# Stock Assessment Study of Chilkoot Lake Sockeye Salmon, 2016-2019 

by

Shane R. Ransbury
Nicole L. Zeiser
Steven C. Heinl
Chase S. Jalbert
and
Sara E. Miller


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| Weights and measures (metric) |  | General |  | Mathematics, statistics |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| centimeter | cm | Alaska Administrative |  | all standard mathematical |  |
| deciliter | dL | Code | AAC | signs, symbols and |  |
| gram | g | all commonly accepted |  | abbreviations |  |
| hectare | ha | abbreviations | e.g., Mr., Mrs., | alternate hypothesis | $\mathrm{H}_{\text {A }}$ |
| kilogram | kg |  | AM, PM, etc. | base of natural logarithm | $e$ |
| kilometer | km | all commonly accepted |  | catch per unit effort | CPUE |
| liter | L | professional titles | e.g., Dr., Ph.D., | coefficient of variation | CV |
| meter | m |  | R.N., etc. | common test statistics | (F, t, $\chi^{2}$, etc.) |
| milliliter | mL | at | @ | confidence interval | CI |
| millimeter | mm | compass directions: east | E | correlation coefficient (multiple) | R |
| Weights and measures (English) |  | north | N | correlation coefficient |  |
| cubic feet per second | $\mathrm{ft}^{3} / \mathrm{s}$ | south | S | (simple) | r |
| foot | ft | west | W | covariance | cov |
| gallon | gal | copyright | © | degree (angular) | - |
| inch | in | corporate suffixes: |  | degrees of freedom | df |
| mile | mi | Company | Co. | expected value | E |
| nautical mile | nmi | Corporation | Corp. | greater than | $>$ |
| ounce | oz | Incorporated | Inc. | greater than or equal to | $\geq$ |
| pound | lb | Limited | Ltd. | harvest per unit effort | HPUE |
| quart | qt | District of Columbia | D.C. | less than | < |
| yard | yd | et alii (and others) | et al. | less than or equal to | $\leq$ |
|  |  | et cetera (and so forth) | etc. | logarithm (natural) | $1 n$ |
| Time and temperature |  | exempli gratia |  | logarithm (base 10) | $\log$ |
| day | d | (for example) | e.g. | logarithm (specify base) | $\log _{2}$, etc. |
| degrees Celsius | ${ }^{\circ} \mathrm{C}$ | Federal Information |  | minute (angular) | , |
| degrees Fahrenheit | ${ }^{\circ} \mathrm{F}$ | Code | FIC | not significant | NS |
| degrees kelvin | K | id est (that is) | i.e. | null hypothesis | $\mathrm{H}_{0}$ |
| hour | h | latitude or longitude | lat or long | percent | \% |
| minute | $\min$ | monetary symbols |  | probability | P |
| second | S | (U.S.) <br> months (tables and | \$, ¢ | probability of a type I error (rejection of the null |  |
| Physics and chemistry |  | figures): first three |  | hypothesis when true) | $\alpha$ |
| all atomic symbols |  | letters | Jan,...,Dec | probability of a type II error |  |
| alternating current | AC | registered trademark | ${ }^{\circledR}$ | (acceptance of the null |  |
| ampere | A | trademark | тм | hypothesis when false) | $\beta$ |
| calorie | cal | United States |  | second (angular) | " |
| direct current | DC | (adjective) | U.S. | standard deviation | SD |
| hertz | Hz | United States of |  | standard error | SE |
| horsepower | hp | America (noun) | USA | variance |  |
| hydrogen ion activity (negative log of) | pH | U.S.C. | United States Code | population sample | Var var |
| parts per million | ppm | U.S. state | use two-letter |  |  |
| parts per thousand | $\mathrm{ppt},$ |  | abbreviations (e.g., AK, WA) |  |  |
| volts | V |  |  |  |  |
| watts | W |  |  |  |  |

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Shane R. Ransbury and Nicole L. Zeiser
Alaska Department of Fish and Game, Division of Commercial Fisheries, Haines
Steven C. Heinl
Alaska Department of Fish and Game, Division of Commercial Fisheries, Ketchikan
Chase S. Jalbert
Alaska Department of Fish and Game, Gene Conservation Laboratory, Anchorage
and

Sara E. Miller
Alaska Department of Fish and Game, Division of Commercial Fisheries, Juneau

Alaska Department of Fish and Game
Division of Sport Fish, Research and Technical Services 333 Raspberry Road, Anchorage, Alaska, 99518-1565

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Shane R. Ransbury and Nicole L. Zeiser, Alaska Department of Fish and Game, Division of Commercial Fisheries, Mile 1 Haines Highway, Haines, Alaska 99827, USA<br>Steven C. Heinl<br>Alaska Department of Fish and Game, Division of Commercial Fisheries, 2030 Sea Level Drive, Suite 205, Ketchikan, Alaska, 99901, USA<br>Chase S. Jalbert<br>Alaska Department of Fish and Game, Division of Commercial Fisheries, 333 Raspberry Road, Anchorage, AK 99518<br>Sara E. Miller<br>Alaska Department of Fish and Game, Division of Commercial Fisheries, 1255 W. 8th Street, Juneau, AK 99801

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#### Abstract

From 2016 to 2019, the Alaska Department of Fish and Game, Division of Commercial Fisheries, continued a stock assessment program that began in 1976 to estimate escapements and harvests of Chilkoot Lake sockeye salmon (Oncorhynchus nerka). Sockeye salmon were counted through a weir near the outlet of Chilkoot Lake, and age, length, and sex data were collected and analyzed each year. Sockeye salmon escapements at the weir were 86,721 fish in 2016, 43,098 fish in $2017,85,463$ fish in 2018 , and 140,378 fish in 2019 , all of which fell within or exceeded the sustainable escapement goal range of $38,000-86,000$ fish. Ocean-age- 3 sockeye salmon (ages 1.3 and 2.3 combined) of both sexes were the shortest in length since scale sampling began in 1982. The stock compositions of sockeye salmon harvested annually in the District 15 commercial drift gillnet fishery were estimated through scale pattern analysis (2016) and genetic stock identification (2017-2019). Estimated commercial harvests of Chilkoot Lake sockeye salmon were 119,843 fish in 2016, 1,933 fish in 2017, 33,969 fish in 2018, and 149,586 fish in 2019. Estimated harvest rates (including subsistence and sport harvests) were $59 \%$ in $2016,9 \%$ in $2017,31 \%$ in 2018, and $52 \%$ in 2019. Estimated fall sockeye salmon fry populations at Chilkoot Lake were $42 \%$ above average in 2016, $45 \%$ below average in 2017, and at or slightly below average in 2018 and 2019, respectively. Average May-September zooplankton density and biomass at Chilkoot Lake were nearly double the long-term average, and zooplankton density in 2019 and biomass in 2016 were the highest recorded since sampling began in 1987.


Keywords: Chilkoot Lake, Chilkoot River, commercial harvest, escapement, enumeration weir, hydroacoustic survey, Oncorhynchus nerka, scale pattern analysis, genetic stock identification, sockeye salmon, sustainable escapement goal, zooplankton

## INTRODUCTION

The Chilkoot and Chilkat River watersheds, located in northern Southeast Alaska near the town of Haines, support 2 of the largest sockeye salmon (Oncorhynchus nerka) runs in Southeast Alaska (Figure 1). Between 1900 and 1920, the annual commercial harvest of sockeye salmon in northern Southeast Alaska averaged 1.5 million fish, the majority of which were believed to originate from the Chilkat and Chilkoot River watersheds (Rich and Ball 1933). Harvests decreased in the early 1920s and remained at relatively low levels thereafter (Eggers et al. 2009). Historically, Chilkoot Lake sockeye salmon were harvested in the large fish trap and purse seine fisheries in Icy and northern Chatham straits as well as in terminal drift gillnet areas of Lynn Canal. Fish traps were eliminated with Alaska statehood in 1959, and Lynn Canal developed into a designated drift gillnet fishing area (District 15), where most of the commercial harvest of Chilkoot Lake sockeye salmon now takes place. District 15 encompasses Section 15-A (north Lynn Canal), Section 15-B (Berners Bay), and Section 15-C (central Lynn Canal; Figure 1). Historically, the sockeye salmon was the primary species targeted from late June through September (McPherson 1990). In recent decades, however, fishing effort has shifted to Section 15-C to harvest substantial hatchery summer chum salmon (O. keta) runs to Douglas Island Pink and Chum, Inc. (DIPAC) release sites at Boat Harbor and Amalga Harbor Terminal Harvest Areas (THAs), which have attracted record-level effort (Bednarski et al. 2016; Gray et al. 2017). The fall fishery is managed to target wild fall-run chum and coho (O. kisutch) salmon. Following a sharp decline in Chilkat River fall-run chum salmon runs in the early 1990s, management of the fall fishery shifted abruptly from an emphasis on harvesting chum salmon to exploiting abundant coho salmon runs (Shaul et al. 2017).

The annual harvest of sockeye salmon in the District 15 commercial drift gillnet fishery averaged 192,000 fish from 1984 to 2015, of which an average 79,000 fish originated from Chilkat Lake, 92,000 originated from Chilkoot Lake, and the remainder were of mixed stock origin (Bednarski et al. 2016). A smaller portion of the Chilkoot Lake run is harvested in the commercial purse seine fisheries that target pink salmon (O. gorbuscha) in Icy and northern Chatham straits (Ingledue 1989; Gilk-Baumer et al. 2015). Annual contributions to those fisheries are not known and likely vary annually depending on fishing effort and the strength of pink salmon runs. Chilkoot Lake
sockeye salmon are also harvested annually in subsistence fisheries in Chilkoot Inlet and Lutak Inlet, where reported harvests for the period 2010-2019 averaged 3,000 fish per year.


Figure 1.-Haines Management Area with sections and statistical areas for the District 15 commercial drift gillnet fishery. Early in the 2018 and 2019 seasons, the fishery was restricted to the black shaded areas in accordance with management actions implemented in the 2018 Chilkat River Chinook salmon action plan (Lum and Fair 2018) and the 2019 Southeast Alaska drift gillnet fishery management plan (Gray et al. 2019) that were designed to reduce commercial harvest of Chilkat River Chinook salmon.

The stock composition of the sockeye salmon harvest in the District 15 commercial drift gillnet fishery has been estimated from scale pattern analysis and, more recently, genetic stock identification (Bednarski et al. 2017). These projects provided information regarding stock contribution (stratified by time), run timing, and age composition. The Alaska Department of Fish and Game (ADF\&G) initiated a scale pattern analysis program in 1980 to estimate the contribution of Chilkat Lake, Chilkoot Lake, and "other" sockeye salmon stocks to the District 15 commercial drift gillnet fishery harvest. Bergander (1974) first developed a dichotomous key based on distinct
differences in their freshwater scale patterns (Stockley 1950). Marshall et al. (1982) improved the sample design and estimated stock contributions using linear discriminant function analysis. McPherson and Marshall (1986) showed that all age classes of the 2 stocks could be identified accurately using a visual classification technique and blind testing procedure. That technique was expanded to include a group of "other" stocks-a combination of Chilkat River mainstem and Berners Bay stocks that contribute to early-season harvests in Lynn Canal (McPherson 1987a). Blind tests to verify accuracy and correct for misclassification were only conducted in the early 1990s. However, historical stock-specific harvest estimates based solely on visual classification were considered to be highly accurate due to consistent differences in freshwater scale patterns, and the difference between initial and corrected estimates varied by only $2 \%$ or less (McPherson and Marshall 1986; McPherson 1987a, 1987b; McPherson and Jones 1987; McPherson 1989; McPherson et al. 1992; McPherson and Olsen 1992).

Although accurate, scale pattern analysis required highly skilled personnel trained in very specific pattern recognition, which could take years to master, and required intensive field sampling and inseason analysis of a very large number of scale samples (Bednarski et al. 2017), whereas genetic stock identification methods are standardized and used widely throughout the state (Shedd et al. 2016). Multiple blind tests conducted by the Northern Boundary Technical Committee of the Pacific Salmon Commission (years 2003, 2009) and by ADF\&G (Lynn Canal, years 2015-2016) indicated that the 2 methods offered similar estimates of salmon stock contribution but that the genetic techniques were able to discriminate stocks at a finer resolution in less time compared to scale pattern analysis (Anne Reynolds-Manney, ADF\&G fisheries biologist, unpublished data ${ }^{1}$ ). As a result, stock composition of sockeye salmon harvests in the District 15 commercial drift gillnet fishery have been estimated solely through genetic stock identification since 2017 (Bednarski et al. 2017).

Chilkoot Lake sockeye salmon escapements have been counted and sampled annually at an adult counting weir on the Chilkoot River since 1976 (Bachman and Sogge 2006; Bachman et al. 2013 and 2014; Bednarski et al. 2016; Brenner et al. 2018). Historically, the run had 2 components, an early and a late run, which were managed as separate units through 2005 (Geiger et al. 2005). Total annual weir counts averaged 81,000 sockeye salmon through 1993 but declined to an average of only 28,000 fish from 1994 to 1999 . Weir counts later rebounded to an average of 62,000 sockeye salmon from 2000 to 2009 and an average of 83,000 fish from 2010 to 2019. In addition to salmon counts, biological data have been collected annually at the weir to estimate age, size, and sex composition of the escapement and, prior to 2017, for use in scale pattern analysis. Basic information about lake productivity and rearing sockeye salmon fry populations has also been collected through limnological and hydroacoustic sampling conducted most years since 1987 (Barto 1996; Riffe 2006; Bachman et al. 2014). Those studies have been used to assess potential sockeye salmon production from the lake (Barto 1996).
The Chilkoot Lake run has been managed for at least 5 different escapement goals since 1976. Informal goals of 80,000-100,000 fish (1976-1980) and 60,000-80,000 fish (1981-1989; Bergander et al. 1988) were replaced in 1990 by a biological escapement goal of 50,500-91,500 sockeye salmon (McPherson 1990). The goal was divided into separate goals for early (16,500$31,500$ fish ) and late runs ( $34,000-60,000$ fish $)$. In 2006 , the escapement goal was rounded to

[^0]$50,000-90,000$ sockeye salmon and classified as a sustainable escapement goal due to uncertainty in escapement levels based on weir counts (Geiger et al. 2005). Early- and late-run goals were eliminated and replaced with weekly cumulative escapement targets based on historical run timing. The current sustainable escapement goal of $38,000-86,000$ sockeye salmon, along with weekly escapement targets, was established in 2009 based on an updated stock-recruit analysis by Eggers et al. (2009). The escapement goal was subsequently reviewed by Brenner et al. (2018), who recommended maintaining the current sustainable escapement goal and weekly escapement targets.

The primary purpose of the sockeye salmon stock assessment program was to estimate escapement and commercial harvest of Chilkoot Lake sockeye salmon. Information provided by this project, in conjunction with stock assessment projects on the adjacent Chilkat River (Sogge and Bachman 2014; Rhea-Fournier et al. 2018), was used inseason to manage the District 15 commercial drift gillnet fishery to ensure escapement goals were met while maximizing and sustaining the harvest of sockeye salmon from the 2 watersheds. Information on age at return is used in reconstruction of brood-year returns and escapement goal evaluations. In addition, hydroacoustic and limnological surveys of Chilkoot Lake were conducted to estimate populations of rearing sockeye salmon fry and to collect information on zooplankton abundance, light penetration, and water temperature profiles.

## STUDY SITE

Chilkoot Lake (ADF\&G Anadromous Waters Catalogue No. 115-33-10200-0010; 59 ${ }^{\circ} 21^{\prime} 16^{\prime \prime} \mathrm{N}$, $135^{\circ} 35^{\prime} 42^{\prime \prime} \mathrm{W}$ ) is located at the head of Lutak Inlet, approximately 16 km northeast of the city of Haines, Alaska (Figures 1 and 2). It is glacially turbid and it has a surface area of $7.2 \mathrm{~km}^{2}(1,734$ acres), a mean depth of 55 m , a maximum depth of 89 m , and a total volume of $382.4 \times 106 \mathrm{~m}^{3}$. The Chilkoot River originates at glacier terminuses east of the Takshunak Mountains and west of the Ferebee Glacier. The glacial river flows approximately 26 km southeast into Chilkoot Lake, then flows approximately 2 km into Lutak Inlet. Early-run sockeye salmon spawn in small lake and river tributaries, and late-run fish spawn in the main channel of the Chilkoot River and along lake beaches where upwelling water occurs (McPherson 1990). Chilkoot Lake is located within the northern temperate rainforest that dominates the Pacific Northwest coast of North America. Although the climate is characterized by cold winters and cool, wet summers, the lake is set in a transitional zone, with warmer and drier summers and cooler winters than the rest of Southeast Alaska (Bieniek et al. 2012). Average precipitation in the study area is approximately $165 \mathrm{~cm} / \mathrm{year}$ (Bugliosi 1988). Sitka spruce (Picea sitchensis), western hemlock (Tsuga heterophylla), and Sitka alder (Alnus viridis) dominate the forested watershed.


Figure 2.-Map showing Lutak Inlet, Chilkoot Lake, and the location of the limnology stations and salmon counting weir.

## OBJECTIVES

## Primary Objectives:

1. Enumerate adult salmon by species through the Chilkoot River weir from the first week of June to the second week of September.
2. Estimate the seasonal age, sex, and length composition of the Chilkoot River sockeye salmon escapement such that the estimated proportions are within $5 \%$ of the true value with at least $95 \%$ probability.
3. Estimate the weekly and annual stock composition of the sockeye salmon harvest by age in the District 15 commercial drift gillnet fishery in 2016 using visual scale pattern analysis.
4. Estimate the weekly stock composition of the sockeye salmon harvest in the District 15 commercial drift gillnet fishery during 2017-2019 using genetic stock identification, such that the estimates are within $7 \%$ of the true value with at least $90 \%$ probability.
5. Estimate the seasonal age-specific stock composition of the sockeye salmon harvest in the District 15 commercial drift gillnet fishery for major age classes (i.e., those contributing $>0.5 \%$; ages $0.3,1.2,1.3,2.2$, and 2.3 ) and "other" age classes combined (e.g., minor age classes, such as ages 1.4, 2.4, 3.3).

## Secondary Objectives:

1. Estimate the abundance and density of sockeye salmon fry and other pelagic fish species in Chilkoot Lake such that the coefficient of variation is no greater than $15 \%$ of the point estimate.
2. Measure water column temperature, record light penetration profiles, and estimate zooplankton species composition, size, density, and biomass in Chilkoot Lake on a monthly basis during the middle of the month, May-September.

## METHODS

## ESCAPEMENT

The Chilkoot Lake adult salmon escapement was counted through a weir located in the Chilkoot River 1 km downstream from Chilkoot Lake. The weir was operated from at least the first week of June through the second week of September each year. The weir is supported by a 110 m long permanent steel structure anchored with 20 cm steel pilings driven approximately 7 m into the bottom of the Chilkoot River channel. Pickets of black iron pipe were installed into the support structure to form a fence across the river channel. The pickets were 2 to 3 m long, with a 2.5 cm outside diameter, and spaced 3.8 cm apart. The weir was regularly inspected, and gaps or small openings were blocked with sandbags or plastic-coated wire mesh to prevent fish from passing undetected. A fish trap, recovery box, counting station, and sampling stations were installed near the center of the weir structure.

In order to minimize handling, most fish were passed by temporarily removing up to 4 pickets at a counting station located between 2 weir-mounted counting chairs near the center of the weir. Fish were counted by species as they passed through the opening. To facilitate identification and enumeration of fish, white plywood panels were stacked in front of and below the opening to force fish higher in the water column as they passed upstream. Fish were trapped or caught with a dip net as they passed through the counting station in the weir and sampled for age, sex, and length. Sampled fish were released into a $2 \mathrm{~m} \times 2 \mathrm{~m} \times 2.5 \mathrm{~m}$ plywood recovery box on the upstream side of the weir to recover from handling. Once fish recuperated, they exited the recovery box by swimming through a large hole in the side of the box.
Stream height and water temperature were recorded at approximately 6:30 am each day. Stream height ( cm ) was measured on a stadia rod, and water temperature $\left({ }^{\circ} \mathrm{C}\right)$ was measured with a permanently installed thermometer near the east end of the weir.

## Weir Passage Estimates

In some years, brief periods of flooding required removal of pickets to prevent structural damage to the weir, and therefore upstream salmon passage had to be estimated for days the weir was inoperable. Estimates were assumed to be zero if passage was probably negligible based on historical or inseason data. Otherwise, estimates for missed passage were calculated following methods used at the Kogrukluk River weir in western Alaska (Hansen and Blain 2013). When the weir was not operated for all of one day, an estimate for that day $\left(\hat{n}_{i}\right)$ was calculated as the average of the number of fish counted on the 2 days before ( $n_{b}$ and $n_{b-1}$ ) and the 2 days after ( $n_{a}$ and $n_{a+1}$ ) the missed day:

$$
\begin{equation*}
\hat{n}_{i}=\left(\frac{n_{b}+n_{b-1}+n_{a}+n_{a+1}}{4}\right) \tag{1}
\end{equation*}
$$

When the weir was not operated for a period of 3 or more days, passage estimates for the missed days were calculated using linear interpolation. This method was appropriate for short periods of inoperability when fish passage was reasonably assumed to have a linear relationship with time. Average fish counts from the 2 days before and 2 days after the inoperable period were used to estimate the counts during the period of missed passage. The estimated fish count ( $\hat{n}$ ) on day $(i)$ of the inoperable period, where $D$ is the total number of inoperable days, was estimated as:

$$
\begin{equation*}
\hat{n}_{i}=\left(\frac{n_{b}+n_{b-1}}{2}\right)+i\left(\frac{\left(n_{a}+n_{a+1}\right)-\left(n_{b}+n_{b-1}\right)}{2(D+1)}\right) \tag{2}
\end{equation*}
$$

## Escapement Age, Sex, AND LengTh Composition

The seasonal age composition of the Chilkoot Lake sockeye salmon escapement was determined from a minimum sample of 665 fish captured at the weir. This sample size was based on work by Thompson (2002) to estimate proportions of 4 or more major age classes. A sample of 510 fish is needed to ensure the estimated proportion of each major age class will be within $5 \%$ of the true value with at least $95 \%$ probability. The sample size was increased to 665 fish to ensure the sampling goal will be met, even if age cannot be determined from $30 \%$ of the sampled fish. In addition, beginning in 2017, 3 scales were sampled from each fish to increase the proportion of readable scales.

In 2016, scale samples were collected at the weir from a daily sample of 40 sockeye salmon (Sogge 2016). This sampling goal was established to ensure sufficient samples of each age class for use in scale pattern analysis of fishery samples (McPherson and Olson 1992) and was far above the number required to estimate the age composition of the escapement. Approximately 20 fish were sampled during the morning shift and 20 more fish in the afternoon or evening shift. The length of each fish was measured from mid eye to tail fork to the nearest 5 mm . Sex was determined by examining external dimorphic sexual maturation characteristics, such as kype development, belly shape, and trunk depth. In 2016, one scale per fish was taken from the preferred area above the lateral line on the left side of the fish on a diagonal downward from the posterior insertion of the dorsal fin to the anterior insertion of the anal fin (INPFC 1963) and placed on a gum card. Date of sample, sex, length, and data regarding the condition of each fish were recorded on standard optical scan forms.

Scale sampling goals were reduced starting in 2017, when escapement samples were no longer required to provide known-origin samples for scale pattern analysis. Up to 10 sockeye salmon ( 70 fish/week) were sampled for matched scales, sex, and length data each morning, and three scales were collected from each sampled fish. This sampling goal was sufficient to meet the objective criteria for estimating age composition because the total seasonal sample typically exceeded the goal of 665 samples. This sample size also met seasonal sex and length composition requirements, because only 385 samples (assuming no data loss) were needed to achieve the precision criteria (within $5 \%$ of the true value $95 \%$ of the time) for estimating sex composition (Thompson 2002).

Scale samples were analyzed at the ADF\&G Region 1 Scale Aging Laboratory in Douglas, Alaska. Scale impressions were made in cellulose acetate and prepared for analysis as described by Clutter and Whitesel (1956). Scales were examined under moderate ( $70 \times$ ) magnification to determine age. Age classes were designated by the European aging system where freshwater and saltwater years were separated by a period (e.g., 1.3 denoted a fish with 1 freshwater and 3 ocean years; Koo 1962). Age, length, and sex data were entered into the Region 1 Commercial Fisheries Database by Douglas staff. The weekly age distribution, the seasonal age distribution weighted by week,
and the mean length by age and sex weighted by week were calculated using standard sampling summary statistics from Cochran (1977; Appendix A).

## Commercial Harvest Estimate

The stock composition of sockeye salmon harvested in the District 15 commercial drift gillnet fishery was determined by visual scale pattern analysis from 1976 to 2016 and by genetic stock identification from 2017 onward (Bednarski et al. 2016). The District 15 commercial drift gillnet fishery season typically begins at 12:00 noon on the third Sunday of June. Openings are then conducted weekly starting at 12:00 noon on Sunday. Each week typically begins with a 48 -hour opening, with the possibility of an extension depending on fishery performance. Commercial harvest data for the District 15 commercial drift gillnet fishery, stratified by statistical week, were obtained from the Region 1 Commercial Fisheries Database. ADF\&G statistical weeks begin on Sunday at 12:01 am and end the following Saturday at midnight and are numbered sequentially starting from the beginning of the calendar year (Appendix B).

## Fishery Sampling

Matched sockeye salmon scale and genetic tissue samples were collected from District 15 commercial drift gillnet fishery landings by ADF\&G port sampling personnel primarily at fish processing facilities in Excursion Inlet and Juneau (Buettner et al. 2017), and also at Petersburg's processing facility in 2018 and 2019. Sampling was stratified by statistical week, and sampling effort spanned the first 10 weeks of the fishery. During the previous 10 years, 2006-2015, an average $92 \%$ of the sockeye salmon harvest occurred during the first 10 weeks of the fishery. In 2016 and 2017, sampling goals were set at 600 fish per week ( 300 at Excursion Inlet and 300 at Juneau; Buettner et al. 2017). In 2018 and 2019, sampling goals were set at 150 fish each from Juneau and Excursion Inlet, and 100 fish from Petersburg. If Excursion Inlet or Juneau were short of samples in a given week, more were collected from Petersburg. A sample of 510 fish was sufficient to describe the weekly estimated sockeye salmon age composition within $5 \%$ of the true proportion with at least $95 \%$ probability. In addition, according to sample theory, under the worstcase scenario (stocks contributing equal proportions) a minimum sample of 200 fish should provide weekly estimates of relative stock composition proportions within $7 \%$ of the true value $90 \%$ of the time (Thompson 1987).

Starting in 2018, sockeye salmon harvested in the District 15 commercial drift gillnet fishery were sampled regardless of the harvest type and all samples were recorded as traditional harvest (harvest code 11). Previously, sockeye salmon harvested in the Boat Harbor terminal harvest area (THA; statistical area 115-11) were not sampled, including sockeye salmon on tenders with fish mixed from traditional and terminal harvest (harvest code 12) fisheries. The Boat Harbor THA is designated to harvest hatchery chum salmon released inside Boat Harbor; however, the THA encompasses a portion of central Lynn Canal (Figure 1) through which mixed stocks of sockeye salmon must migrate, and sockeye salmon are harvested incidentally in the fishery. Over the 10 years 2008-2017, an average $21 \%$ (range: 12-36\%) of sockeye salmon harvested in central Lynn Canal (statistical areas 115-10 and 115-11) were harvested in the Boat Harbor THA. Since 2018, all sockeye salmon samples have been identified as harvest code 11 .

Sampling protocols ensured that samples were as representative of harvests as possible to account for fluctuations in harvest and effort over the course of a weekly fishery. Deliveries with harvests mixed from more than one gear type or fishing district were not sampled, no more than 40 samples were collected from a single delivery, no more than 200 samples were collected from a single tender
delivery, samples were collected without regard to size or sex of fish, and, whenever possible, samples were systematically collected from the entire hold as it was offloaded to ensure they were representative of the entire delivery. Sampled fish were identified to sex, and one scale per fish was taken from the preferred area (INPFC 1963). Samples were processed and aged at the ADF\&G salmon-aging laboratory following procedures described above for escapement samples.

## Scale Pattern Analysis

The general methods of District 15 scale pattern analysis remained unchanged from the mid-1980s through 2016: escapement scale samples from 3 stocks of known origin, Chilkoot Lake, Chilkat Lake, and "other" (Chilkat River mainstem and Berners Bay stocks), were aged and compared to scale samples from the commercial fisheries (McPherson and Olson 1992). Known-origin scale samples were collected weekly from sockeye salmon at the Chilkoot River weir (this study), at Chilkat Lake, and from Chilkat River mainstem spawning populations (Rhea-Fournier et al. 2018). Samples were also collected annually from spawning populations in Berners Bay (Berners and Lace Rivers) and along the mainstem of the Chilkat River. These latter samples may not have been representative of the entire Berners River and Chilkat River mainstem populations because they were collected opportunistically and were sometimes temporally and spatially limited. Samples were processed and aged at the ADF\&G scale aging laboratory following procedures described above for escapement samples.
Known-origin scale samples were analyzed inseason on a weekly basis at the ADF\&G scale aging laboratory, after which commercial fishery samples were analyzed and assigned to 1 of 3 stocks, Chilkoot Lake, Chilkat Lake, and "other", based on scale characteristics. The size of the freshwater annulus and the number of circuli in the freshwater growth zones were the principal scale characteristics used to distinguish between runs; however, the total size of the freshwater growth zone, size of the freshwater-plus growth zone, and completeness of circuli and spacing between circuli in the freshwater growth zone were also considered. Differences in age composition between stocks and migratory timing by age were also accounted for inseason. The weekly proportions of classified scale samples were applied to the District 15 commercial drift gillnet harvest to provide weekly estimates of stock contribution for inseason management and postseason estimates of total harvest by stock, weighted by statistical week. The stock proportions in the last sampled statistical week in 2016 (statistical week 35) were used to estimate contribution for the final weeks of the fishery, statistical weeks 35-40, which accounted for $2 \%$ of the sockeye salmon harvest.

## Genetic Stock Identification

Beginning in 2017, stock composition of the sockeye salmon harvest in the District 15 commercial drift gillnet fishery was estimated through genetic stock identification. Laboratory analysis, including quality control, was performed by the ADF\&G Gene Conservation Laboratory following methods outlined in Dann et al. (2012). Sockeye salmon were identified to 7 reporting groups: Chilkat Lake, Chilkat mainstem, Chilkoot Lake, Juneau mainland, Snettisham, Taku River/Stikine mainstem, and Other; however, reporting groups were reduced to Chilkat Lake, Chilkoot Lake, and Other for postseason reporting (Appendix C; Zeiser et al. 2019). Stock composition was estimated inseason for each statistical week using a Bayesian mixed stock analysis (MSA) approach as implemented in the R package rubias $^{2}$ (Moran and Anderson 2019), which compared

[^1]fishery samples against the genetic baseline described in Rogers Olive et al. (2018). Postseason, samples were reanalyzed with age composition data from the harvest using Mark and Ageenhanced Genetic Mixture Analysis (MAGMA), an extension of the Pella-Masuda genetic stock identification model (Pella and Masuda 2001) that incorporates ages from matched scale samples to provide age-specific stock composition estimates for all major contributing age classes (i.e., those contributing $>0.5 \%$ ). MAGMA was used to analyze stock composition in the first ten statistical weeks of the sockeye salmon fishery. The stock proportions in the last sampled statistical week were used to estimate contribution for the final weeks of the fishery, generally statistical weeks $35-40$, which accounted for an average $9 \%$ (range $3-15 \%$ ) of the sockeye salmon harvest during 2017-2019.

## Laboratory Analysis

Genomic DNA was extracted from tissue samples using a NucleoSpin 96 Tissue Kit by MachereyNagel (Düren, Germany). A multiplexed preamplification polymerase chain reaction (PCR) of 48 screened single nucleotide polymorphism (SNP) markers was used to increase the concentration of template DNA. Samples were genotyped for 48 screened SNP markers using two sets of Fluidigm 192.24 Dynamic ArrayTM Integrated Fluidic Circuits, which systematically combined up to 24 assays and 192 samples into 4,608 parallel reactions (https://www.fluidigm.com). The Dynamic Arrays were read on a Fluidigm EP1 System after amplification and scored using Fluidigm SNP Genotyping Analysis software. If necessary, SNPs were rescreened on a QuantStudio 12K Flex Real-Time PCR System (Life Technologies) as a backup method for assaying genotypes. Genotypes were imported and archived in the Gene Conservation Laboratory Oracle database, LOKI.
A quality control (QC) analysis was conducted postseason to identify laboratory errors and to measure the background discrepancy rate of the genotyping process. The QC analyses were performed by staff not involved in the original genotyping as described in detail by Dann et al. (2012). Briefly, the method consisted of re-extracting $8 \%$ of project fish and genotyping them for the same SNPs assayed in the original genotyping process. Discrepancy rates were calculated as the number of conflicting genotypes, divided by the total number of genotypes compared. These rates describe the difference between original project data and QC data for all SNPs and can identify extraction, assay plate, and genotyping errors. Assuming that discrepancies among analyses are due equally to errors during the original genotyping and during QC, error rates in the original genotyping was estimated as half the rate of discrepancies. If there were many discrepancies, a duplicate check was performed to determine if the QC fish were a better match to any other project fish. A QC fish matching other project fish would indicate that fish were swapped during the extraction process. This information was used to identify which, and how many, fish should be re-extracted.

## Statistical Analysis

Genotypes in the LOKI database were imported into the statistical program R for analysis. Prior to statistical analysis, 3 statistical quality control analyses were performed to ensure high-quality data: 1) individuals missing $>20 \%$ of their genotype data (markers) were identified and removed from analyses because this is indicative of low-quality DNA ( $80 \%$ rule; Dann et al. 2012); 2) duplicate individuals were identified and removed; and 3) non-sockeye salmon were identified and removed.

Stock composition for each stratum was estimated using the R package rubias. Markov Chain Monte Carlo (MCMC) methods, using a single chain with starting values equal among all populations, formed the posterior distribution that described the stock composition of each stratum. Summary statistics were tabulated from these distributions to describe stock compositions. Stock composition estimates of commercial harvest were applied to observed harvest (obtained from fish ticket data) to quantify stock-specific harvests within each week. Postseason, age-specific stock composition for all major contributing age classes ( $>0.5 \%$ ) was estimated seasonally through a MAGMA model. Weekly and seasonal estimates were provided, by age group, using MAGMA. This method requires 2 sets of parameters: 1) a vector of stock compositions summing to 1 weighted by harvest per stratum; and 2) a matrix of age composition, with a row for each stock summing to 1 and a column for each age class. This information was "completed" iteratively by stochastically assigning each fish to a population, then estimating the stock proportions based on summaries of assignment from each iteration. In this process, all available information (i.e., age and genotype) was used to assign individuals to stock of origin.

To initialize the algorithm, all fish with unknown origin or age were stochastically assigned to a population or age group, and then proportions for populations and age groups were estimated in the following steps:

1) All age data were summarized by assigned and observed populations for both wild and hatchery individuals;
2) Population and age compositions were estimated from previous summaries (accounting for sampling error);
3) Each wild fish with genotypes was stochastically assigned to a wild population of origin based on the product of its genotypic frequency, age frequency, and population proportion;
4) Each wild fish without genotypes was stochastically assigned to a population of origin based on the product of its age frequency and population proportion; and
5) Steps 1-4 were repeated while updating the estimates of the stock proportions and age compositions with each iteration.
This algorithm was run for 40,000 repetitions, and the first 20,000 repetitions were discarded to eliminate the effect of the initial state. Five MCMC chains were run and checked for convergence among chains using the Gelman-Rubin convergence diagnostic (Gelman and Rubin 1992; Brooks and Gelman 1998). The point estimates and credible intervals for stock-specific age compositions were summary statistics of the output.

## Fry Population Estimate

Hydroacoustic sampling methods were used to estimate abundance of sockeye salmon fry and other small pelagic fish in Chilkoot Lake. To control year-to-year variation in our estimates, surveys were conducted annually along the same 12 transects ( 2 from each of 6 sampling sections of the lake) that were randomly chosen in 2002 as permanent transects (Riffe 2006). Hydroacoustic surveys were conducted annually between early October and mid-November, and a midwater trawl survey was performed in 2019 to examine species composition.
Hydroacoustic sampling was conducted after sunset, and all transects were sampled in the same night. A Biosonics DT-X scientific echosounder ( $430 \mathrm{kHz}, 7.3^{\circ}$ split-beam transducer) with Biosonics Visual Acquisition version 5.0 software was used to collect data. The ping rate was set
to 5 pings/sec and the pulse width was set to 0.3 ms . Surveys were conducted at a constant boat speed of about $2.0 \mathrm{~m} / \mathrm{sec}$. A target strength of -40 dB to -70 dB was used to represent fish within the size range of juvenile sockeye salmon and other small pelagic fish.

Fish-target density $\widehat{M}_{i j}$ (targets $/ \mathrm{m}^{2}$ ) in section $i$ across transect $j$ was estimated using Biosonics Visual Analyzer version 4.1 software, using echo integration methods (MacLennan and Simmonds 1992). Methods for calculating fish population estimates were similar to DeCino (2001) and DeCino and Willette (2014) and adapted from Burczynski and Johnson (1986). The population estimate of each transect $j$ in a section $i$ was estimated as:

$$
\begin{equation*}
\widehat{N}_{i j}=a_{i} \widehat{M}_{i j}, \tag{3}
\end{equation*}
$$

where $a_{i}$ represents the surface area $\left(\mathrm{m}^{2}\right)$ of the lake in section $i$. Using transects as the sampling unit (Burczynski and Johnson 1986), fish abundance ( $\widehat{N}_{i}$ ) across each section was estimated from the mean abundance of the replicate transects $j$ in section $i$,

$$
\begin{equation*}
\widehat{N}_{i}=J^{-1} \sum_{j=1}^{J} \widehat{N}_{i j} \tag{4}
\end{equation*}
$$

with variance

$$
\begin{equation*}
v\left(\widehat{N}_{i}\right)=\sum\left(\widehat{N}_{i j}-\widehat{N}_{i}\right)^{2}(J-1)^{-1} J^{-1} . \tag{5}
\end{equation*}
$$

The sum of the 6 section estimates $\left(\widehat{N}_{i}\right)$ provided an estimate of total targets for the entire lake $(\widehat{N})$. Note that target density was expressed as average targets per unit of lake surface area $a_{i}$, not per unit of volume. Because the estimate of total targets in each section was essentially independent (neglecting any movement of fry from one section to the other during surveys), the sample variance of the estimate of the total targets in the entire lake, $v(\widehat{N})$, was estimated by summing the sample variances $v\left(\widehat{N}_{i}\right)$ across all 6 sections. Sampling error for the estimate of total targets for the entire lake was measured and reported with the coefficient of variation (CV; Sokal and Rohlf 1981). The CV of population estimates was $15 \%$ or less in 8 of 12 years from 2004 to 2015 (Sogge 2016).

Historically, estimates of total targets were partitioned into species categories based on the proportion of each species captured in annual midwater trawls. A $2 \mathrm{~m} \times 2 \mathrm{~m}$ elongated trawl net was used to capture pelagic fish and estimate species composition (Riffe 2006). Four to six nighttime trawls were conducted at various depths, ranging from near surface to 15 m . Trawl depths and duration were determined from observations of fish densities and distributions throughout the lake during the hydroacoustic survey. Fish were counted by species and released. Beginning in 2015, the frequency of the trawl surveys was reduced because the vast majority of fish captured in past trawl surveys were sockeye salmon fry (median $=99 \% ; n=26$ years; Bednarski et al. 2016). The most recent trawl, conducted in 2019, was performed to ensure that species composition in the lake had not changed markedly since 2014.

## Limnological Assessment

Basic limnological data, including zooplankton, light, and temperature sampling, were collected monthly between May and October. Since 2008, all limnological sampling has been conducted at stations $1 \mathrm{~A}\left(59^{\circ} 21.88^{\prime} \mathrm{N}, 135^{\circ} 36.64^{\prime} \mathrm{W}\right)$ and $2 \mathrm{~A}\left(59^{\circ} 20.81^{\prime} \mathrm{N}, 135^{\circ} 35.79^{\prime} \mathrm{W}\right.$; Figure 2), which are marked by anchored buoys in the lake (Bednarski et al. 2016; Zeiser et al. 2019). The stations were marked by anchored buoys at the beginning of the season using a GPS navigational device. Results were averaged between stations by month and season, and the season was standardized to May-September average to be comparable over all years.

## Light and Temperature Profiles

Light penetration measurements were used to estimate the euphotic zone depth (EZD) of the lake, defined as the depth at which light (photosynthetically available radiation at $400-700 \mathrm{~nm}$ ) is attenuated to $1 \%$ of the intensity just below the lake surface (Schindler 1971). Photometric illuminance was recorded as lumens per square meter $\left(\mathrm{lm} / \mathrm{m}^{2}\right)$ at 0.5 m intervals, from just below the lake surface to the depth at which ambient light level equals $1 \%$ of the subsurface recording. The natural log of the ratio of light intensity $I$ just below the surface ( $I_{0}$ ) to light intensity at depth $Z$, or $\ln \left(I_{0} / I z\right)$, was calculated for each depth. The vertical light extinction coefficient $\left(K_{d}\right)$, the rate $\left(\mathrm{m}^{-1}\right)$ at which light dims with increasing depth, was estimated as the slope of the regression of $\ln \left(I_{0} / I_{z}\right)$ versus depth, and EZD was calculated as $4.6502 / K_{d}$ (Kirk 1994; Edmundson et al. 2000). Only the measurements recorded from 5 cm below the surface to just below $1 \%$ of the subsurface light level were used in the calculations, because use of data at depths below $1 \%$ of the initial subsurface measurement would skew the estimate of EZD.
Light profiles were collected at each station using an ILT 1400 International Light Technologies Photometer. A Protomatic light meter that measures illumination in foot candles or a secchi disk (Koenings et al. 1987) were occasionally used as a backup.
Temperature $\left({ }^{\circ} \mathrm{C}\right)$ was measured with a Yellow Springs Instruments Model 58 meter. Temperature was recorded at 1 m intervals from the lake surface to a depth of 20 m , and at 5 m intervals from 20 m to a depth of 50 m .

## Secondary Production

Zooplankton samples were collected at each sampling station using a 0.5 m diameter, $153 \mu \mathrm{~m}$ mesh conical net. Vertical zooplankton tows were pulled from a depth of 50 m to the surface at a constant speed of $0.5 \mathrm{~m} / \mathrm{sec}$. Once the top of the net cleared the surface, the rest of the net was pulled slowly out of the water and rinsed from the outside with lake water to wash organisms into the screened sampling container at the cod end of the net. All specimens in the sampling container were carefully rinsed into a 250 ml sampling bottle and preserved in buffered $10 \%$ formalin or $10 \% \mathrm{EtOH}$. Samples were analyzed at the ADF\&G Kodiak Limnology Lab using methods detailed in the ADF\&G Limnology Field and Laboratory Manual (Koenings et al. 1987). Results were averaged between stations by month and season and standardized to May-September in order to compare across all years of sampling.

## RESULTS

## ESCAPEMENT

## 2016

In 2016, 86,721 sockeye, 8,354 pink, 116 chum, 53 coho, and 2 Chinook salmon were enumerated through the Chilkoot River weir between 3 June and 9 September (statistical weeks 23-37; Table 1; Figure 3; Appendices D and E). A high-water event during 28-29 July required removing pickets from the weir, which allowed fish to pass uncounted for approximately 27 hours. An interpolation of 1,749 sockeye salmon ( $2 \%$ of the total weir count) was calculated to estimate sockeye salmon passage during those 2 days. Weekly sockeye salmon escapements were below the lower-bound escapement goal targets for the first 5 weeks of the season, rose above the lower bound beginning in statistical week 28, and then increased dramatically in statistical week 32 . The total sockeye salmon escapement of 86,721 slightly exceeded the upper bound of the escapement goal range
(Table 1; Figure 3). The pink salmon escapement of 8,354 fish was well below the long-term (1976-2015) average of 26,594 fish (Appendix D).

## 2017

In 2017, 43,098 sockeye, 58,664 pink, 529 chum, 12 coho, and 11 Chinook salmon were enumerated through the Chilkoot River weir between 2 June and 6 September (statistical weeks 22-36; Table 2; Figure 3; Appendices D and F). A high-water event during 9-11 June required removing pickets from the weir, which allowed fish to pass uncounted for approximately 34 hours. An interpolation of 60 sockeye salmon ( $<1 \%$ of the total weir count) was calculated to estimate the sockeye salmon passage during those 3 days. Weekly sockeye salmon escapements were below the lower-bound escapement goal targets for the first 7 weeks of the season but rose above the lower bound escapement goal target in statistical week 30. The total sockeye salmon escapement of 43,098 fish exceeded the lower bound of the escapement goal range (Table 2; Figure 3). The pink salmon escapement of 58,664 was well above the long-term (1976-2015) average (Appendix D).


Figure 3.-Comparison of weekly cumulative escapement of sockeye salmon through the Chilkoot River weir compared to the 1976-2015 average and upper and lower bounds of the weekly escapement goal targets (based on Eggers et al. 2009).

Table 1.-Weekly escapement of sockeye salmon at the Chilkoot River weir compared to weekly management targets and sustainable escapement goal range, 2016.

|  | Escapement |  |  | Escapement goal $^{\mathrm{a}}$ |  |
| :---: | ---: | ---: | ---: | :---: | :---: |
| Statistical <br> week | Weekly | Cumulative |  | Cumulative <br> lower bound | Cumulative <br> upper bound |
| 23 | 0 | 0 |  | 378 | 856 |
| 24 | 73 | 73 |  | 1,924 | 4,354 |
| 25 | 1,067 | 1,140 |  | 4,593 | 10,396 |
| 26 | 955 | 2,095 |  | 6,852 | 15,508 |
| 27 | 5,410 | 7,505 |  | 8,333 | 18,858 |
| 28 | 5,863 | 13,368 |  | 10,102 | 22,863 |
| 29 | 7,167 | 20,535 |  | 13,286 | 30,069 |
| 30 | 8,429 | 28,964 |  | 17,689 | 40,032 |
| 31 | 4,552 | 33,516 |  | 23,236 | 52,587 |
| 32 | 40,054 | 73,570 |  | 28,267 | 63,973 |
| 33 | 7,723 | 81,293 |  | 31,565 | 71,437 |
| 34 | 1,816 | 83,109 |  | 34,371 | 77,787 |
| 35 | 1,633 | 84,742 |  | 36,275 | 82,096 |
| 36 | 1,455 | 86,197 | 37,524 | 84,923 |  |
| 37 | 524 | 86,721 | 38,000 | 86,000 |  |

a Weekly escapement goal targets are from Eggers et al. (2009).

Table 2.-Weekly escapement of sockeye salmon at the Chilkoot River weir compared to weekly management targets and sustainable escapement goal range, 2017.

|  | Escapement |  |  | Escapement goal ${ }^{\text {a }}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Statistical <br> week | Weekly | Cumulative |  | Cumulative <br> lower bound | Cumulative <br> upper bound |
| $22-23$ | 137 | 137 |  | 378 | 856 |
| 24 | 82 | 219 |  | 1,924 | 4,354 |
| 25 | 391 | 610 |  | 4,593 | 10,396 |
| 26 | 1,091 | 1,701 |  | 6,852 | 15,508 |
| 27 | 2,403 | 4,104 |  | 8,333 | 18,858 |
| 28 | 2,717 | 6,821 |  | 10,102 | 22,863 |
| 29 | 6,293 | 13,114 |  | 13,286 | 30,069 |
| 30 | 6,539 | 19,653 |  | 17,689 | 40,032 |
| 31 | 10,169 | 29,822 |  | 23,236 | 52,587 |
| 32 | 4,039 | 33,861 |  | 28,267 | 63,973 |
| 33 | 3,853 | 37,714 |  | 31,565 | 71,437 |
| 34 | 1,995 | 39,709 |  | 34,371 | 77,787 |
| 35 | 2,429 | 42,138 |  | 36,275 | 82,096 |
| 36 | 960 | 43,098 |  | 37,524 | 84,923 |
| 37 | $-^{b}$ | $-{ }^{\text {b }}$ |  |  | 38,000 |

[^2]
## 2018

In 2018, 85,463 sockeye, 5,475 pink, 225 chum, 95 coho, and 31 Chinook salmon were enumerated through the Chilkoot River weir between 3 June and 8 September (statistical weeks 23-36; Table 3; Figure 3; Appendices D and G). A high-water event during 9-10 August required removing pickets from the weir, which allowed fish to pass uncounted for approximately 27 hours. An interpolation of 897 sockeye salmon ( $1 \%$ of the total weir count) was calculated to estimate the sockeye salmon passage during those 2 days. Weekly sockeye salmon escapements were below the lower-bound escapement goal targets for the first 6 weeks of the season. The total sockeye salmon escapement of 85,463 fish was near the upper bound of the escapement goal range (Table 3; Figure 3). The pink salmon escapement of 5,475 fish was well below the long-term (1976-2015) average (Appendix D).

## 2019

In $2019,140,378$ sockeye, 17,156 pink, 396 chum, 80 coho, and 64 Chinook salmon were enumerated through the Chilkoot River weir between 7 June and 8 September (statistical weeks 23-37; Table 4; Figure 3; Appendices D and H). No interpolation for missed days was required during the season. Weekly sockeye salmon escapements were below the lower-bound escapement goal targets for the first 4 weeks of the season, but by statistical week 30 , a dramatic increase in sockeye salmon numbers pushed escapement over the upper bound of the escapement goal. The 2019 sockeye salmon escapement was the largest on record (Table 4; Figure 3), surpassing the record of 118,166 sockeye salmon set in 2012. The pink salmon escapement of 17,156 fish was below the historical average (Appendix D).

Table 3.-Weekly escapement of sockeye salmon at the Chilkoot River weir compared to weekly management targets and sustainable escapement goal range, 2018.

| Statistical week | Escapement |  | Escapement goal ${ }^{\text {a }}$ |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Weekly | Cumulative | Cumulative lower bound | Cumulative upper bound |
| 23 | 4 | 4 | 378 | 856 |
| 24 | 26 | 30 | 1,924 | 4,354 |
| 25 | 330 | 360 | 4,593 | 10,396 |
| 26 | 1,518 | 1,878 | 6,852 | 15,508 |
| 27 | 2,359 | 4,237 | 8,333 | 18,858 |
| 28 | 5,421 | 9,658 | 10,102 | 22,863 |
| 29 | 11,108 | 20,766 | 13,286 | 30,069 |
| 30 | 37,968 | 58,734 | 17,689 | 40,032 |
| 31 | 13,262 | 71,996 | 23,236 | 52,587 |
| 32 | 4,304 | 76,300 | 28,267 | 63,973 |
| 33 | 3,692 | 79,992 | 31,565 | 71,437 |
| 34 | 3,732 | 83,724 | 34,371 | 77,787 |
| 35 | 1,275 | 84,999 | 36,275 | 82,096 |
| 36 | 464 | 85,463 | 37,524 | 84,923 |
| 37 | - ${ }^{\text {b }}$ | ${ }^{\text {b }}$ | 38,000 | 86,000 |

[^3]Table 4.-Weekly escapement of sockeye salmon at the Chilkoot River weir compared to weekly management targets and sustainable escapement goal range, 2019.

|  | Escapement |  |  | Escapement goal $^{\text {a }}$ |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Statistical <br> week | Weekly | Cumulative |  | Cumulative <br> lower bound | Cumulative <br> upper bound |
| 23 | 2 | 2 |  | 378 | 856 |
| 24 | 17 | 19 |  | 1,924 | 4,354 |
| 25 | 924 | 943 |  | 4,593 | 10,396 |
| 26 | 3,323 | 4,266 |  | 6,852 | 15,508 |
| 27 | 4,509 | 8,775 |  | 8,333 | 18,858 |
| 28 | 9,718 | 18,493 |  | 10,102 | 22,863 |
| 29 | 12,072 | 30,565 |  | 13,286 | 30,069 |
| 30 | 56,078 | 86,643 |  | 17,689 | 40,032 |
| 31 | 27,446 | 114,089 |  | 23,236 | 52,587 |
| 32 | 7,150 | 121,239 |  | 28,267 | 63,973 |
| 33 | 9,182 | 130,421 |  | 31,565 | 71,437 |
| 34 | 4,600 | 135,021 |  | 34,371 | 77,787 |
| 35 | 3,458 | 138,479 |  | 36,275 | 82,096 |
| 36 | 1,694 | 140,173 | 37,524 | 84,923 |  |
| 37 | 205 | 140,378 |  | 38,000 | 86,000 |

a Weekly escapement goal targets are from Eggers et al. (2009).

## COMMERCIAL HARVEST ESTIMATE

## 2016

In 2016, 188 , 844 sockeye salmon were harvested in the District 15 commercial drift gillnet fishery. Scale samples from a total sample of 3,995 fish (about $2 \%$ of the commercial harvest) were used in scale pattern analysis to determine weekly stock proportions of the commercial sockeye salmon harvest. Chilkoot Lake sockeye salmon accounted for an estimated $64 \%$ of the total harvest, all weeks combined, or approximately 119,843 fish (Table 5; Appendix I). The Chilkoot Lake sockeye salmon harvest was dominated by age-1.3 fish ( $86 \%$ ), followed by age- 2.3 fish ( $12 \%$ ) and age- 1.2 fish ( $1 \%$ ). The total run was estimated to be 211,830 fish, including the estimated subsistence ( $5,051 \mathrm{fish}$ ) and sport ( 215 fish) harvests, and the total harvest rate was estimated to be $59 \%$ (Appendix M).

## 2017

In 2017, 39,716 sockeye salmon were harvested in the District 15 commercial drift gillnet fishery. A total of 2,915 sockeye salmon were sampled, of which 1,840 fish (about $5 \%$ of the commercial harvest) were genotyped for use in genetic stock identification analysis. Chilkoot Lake sockeye salmon accounted for an estimated $5 \%$ of the total harvest, all weeks combined, or approximately 1,933 fish ( $90 \%$ CI = 1,524-2,370 fish; Table 6; Appendices I and J). The Chilkoot Lake sockeye salmon harvest was dominated by age-1.3 fish (61\%), followed by age-1.2 fish (18\%) and age-2.3 fish ( $14 \%$ ). The total run was estimated to be 47,366 fish, including the estimated subsistence ( $2,102 \mathrm{fish}$ ) and sport ( 233 fish ) harvests, and the total harvest rate was estimated to be $9 \%$ (Appendix M).

Table 5.-Estimated commercial harvest of Chilkoot Lake, Chilkat Lake, and other sockeye salmon stocks in the District 15 commercial drift gillnet fishery based on visual scale pattern analysis, 2016.

| Statistical week | Commercial harvest | Sample size | Estimated stock composition |  |  | Estimated harvest |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Chilkoot Lake | Chilkat Lake | Other ${ }^{\text {a }}$ | Chilkoot Lake | Chilkat Lake | Other ${ }^{\text {a }}$ |
| 25 | 0 | 0 | 0\% | 0\% | 0\% | 0 | 0 | 0 |
| 26 | 1,896 | 212 | 33\% | 41\% | 26\% | 626 | 778 | 492 |
| 27 | 2,514 | 378 | 35\% | 22\% | 43\% | 878 | 552 | 1,084 |
| 28 | 8,361 | 436 | 40\% | 10\% | 50\% | 3,318 | 863 | 4,181 |
| 29 | 15,335 | 436 | 38\% | 23\% | 39\% | 5,768 | 3,552 | 6,014 |
| 30 | 32,443 | 481 | 58\% | 20\% | 22\% | 18,886 | 6,408 | 7,150 |
| 31 | 9,941 | 450 | 48\% | 14\% | 38\% | 4,772 | 1,370 | 3,800 |
| 32 | 21,154 | 309 | 61\% | 19\% | 19\% | 13,007 | 4,108 | 4,039 |
| 33 | 58,180 | 440 | 81\% | 12\% | 7\% | 47,205 | 7,140 | 3,835 |
| 34 | 22,461 | 416 | 74\% | 20\% | 6\% | 16,630 | 4,589 | 1,242 |
| $35-40^{\text {b }}$ | 16,559 | 437 | 53\% | 40\% | 7\% | 8,753 | 6,631 | 1,175 |
| Total | 188,844 | 3,995 | 63\% | 19\% | 17\% | 119,843 | 35,991 | 33,010 |

a Other includes Chilkat River mainstem spawning stocks.
b Harvest proportions and numbers for statistical weeks 36-40 were estimated using the proportions from the last statistical week with scale pattern analysis estimates, in this case statistical week 35.

Table 6.-Estimated commercial harvest of Chilkoot Lake, Chilkat Lake, and other sockeye salmon stocks in the District 15 commercial drift gillnet fishery based on MAGMA genetic stock identification analysis, 2017.

| Statistical week | Commercial harvest | Estimated stock composition |  |  | Estimated Chilkoot harvest and CI |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Chilkoot Lake | Chilkat Lake | Other ${ }^{\text {a }}$ | Harvest | Lower 90\% | Upper 90\% |
| 25 | 1,358 | 5\% | 13\% | 82\% | 62 | 31 | 100 |
| 26 | 2,623 | 6\% | 11\% | 82\% | 167 | 91 | 259 |
| 27 | 1,743 | 6\% | 21\% | 73\% | 113 | 60 | 176 |
| 28 | 759 | 11\% | 16\% | 72\% | 86 | 49 | 129 |
| 29 | 6,077 | 2\% | 6\% | 92\% | 145 | 42 | 296 |
| 30 | 5,834 | 3\% | 6\% | 91\% | 202 | 82 | 356 |
| 31 | 5,590 | 6\% | 9\% | 86\% | 312 | 148 | 512 |
| 32 | 5,772 | 5\% | 36\% | 59\% | 288 | 136 | 480 |
| 33 | 2,171 | 8\% | 17\% | 74\% | 184 | 110 | 269 |
| $34-40^{\text {b }}$ | 7,789 | 5\% | 14\% | 82\% | 374 | 166 | 625 |
| Total | 39,716 | 5\% | 14\% | 81\% | 1,933 | 1,524 | 2,370 |

[^4]
## 2018

In 2018, 81,688 sockeye salmon were harvested in the District 15 commercial drift gillnet fishery. A total of 3,407 sockeye salmon were sampled, of which 1,794 fish (about $2 \%$ of the commercial harvest) were genotyped for use in genetic stock identification analysis. Chilkoot Lake sockeye salmon accounted for an estimated $42 \%$ of the total harvest, all weeks combined, or approximately 33,969 fish ( $90 \%$ CI $=32,077-35,850$ fish; Table 7; Appendices I and K). The Chilkoot Lake sockeye salmon harvest was dominated by age-1.3 fish (73\%), followed by age-1.2 fish ( $21 \%$ ) and age-2.3 fish (5\%). The total run was estimated to be 123,997 fish, including the estimated subsistence ( 4,406 fish) and sport ( 159 fish) harvests, and the total harvest rate was estimated to be $31 \%$ (Appendix M).

## 2019

In 2019, 241,533 sockeye salmon were harvested in the District 15 commercial drift gillnet fishery. A total of 3,803 sockeye salmon were sampled, of which 1,880 fish (about $1 \%$ of the commercial harvest) were genotyped for use in genetic stock identification analysis. Chilkoot Lake sockeye salmon accounted for an estimated $62 \%$ of the total harvest, all weeks combined, or approximately 149,586 fish ( $90 \%$ CI $=144,305-154,702$ fish; Table 8; Appendices I and L). The Chilkoot Lake sockeye salmon harvest was dominated by age-1.3 fish ( $90 \%$ ) and age-1.2 fish ( $8 \%$ ). The total run was estimated to be 293,723 fish, including the estimated subsistence ( 3,673 fish) and sport ( 86 fish) harvests, and the total harvest rate was estimated to be $52 \%$ (Appendix M).

Table 7.-Estimated commercial harvest of Chilkoot Lake, Chilkat Lake, and other sockeye salmon stocks in the District 15 commercial drift gillnet fishery based on MAGMA genetic stock identification analysis, 2018.

| Statistical week | Commercial harvest | Estimated stock composition |  |  | Estimated Chilkoot harvest and CI |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Chilkoot Lake | Chilkat Lake | Other ${ }^{\text {a }}$ | Harvest | Lower 90\% | Upper 90\% |
| 25 | 263 | 18\% | 27\% | 55\% | 47 | 33 | 62 |
| 26 | 904 | 8\% | 26\% | 66\% | 74 | 44 | 108 |
| 27 | 3,630 | 16\% | 24\% | 60\% | 580 | 428 | 745 |
| 28 | 6,450 | 27\% | 16\% | 57\% | 1,725 | 1,380 | 2,081 |
| 29 | 4,303 | 28\% | 17\% | 55\% | 1,216 | 976 | 1,468 |
| 30 | 10,149 | 39\% | 21\% | 40\% | 3,977 | 3,373 | 4,583 |
| 31 | 19,931 | 50\% | 21\% | 29\% | 10,026 | 8,803 | 11,209 |
| 32 | 8,880 | 42\% | 28\% | 31\% | 3,702 | 3,189 | 4,228 |
| 33 | 8,357 | 24\% | 26\% | 50\% | 1,995 | 1,582 | 2,431 |
| $34-40^{\text {b }}$ | 18,821 | 56\% | 28\% | 15\% | 10,628 | 9,553 | 11,676 |
| Total | 81,688 | 42\% | 24\% | 35\% | 33,969 | 32,077 | 35,850 |

[^5]Table 8.-Estimated commercial harvest of Chilkoot Lake, Chilkat Lake, and other sockeye salmon stocks in the District 15 commercial drift gillnet fishery based on MAGMA genetic stock identification analysis, 2019.

| Statistical week | Commercial harvest | Estimated stock composition |  |  | Estimated Chilkoot harvest and CI |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Chilkoot Lake | Chilkat Lake | Other ${ }^{\text {a }}$ | Harvest | Lower 90\% | Upper 90\% |
| 25-26 | 2,215 | 17\% | 27\% | 57\% | 366 | 275 | 466 |
| 27 | 6,573 | 20\% | 16\% | 64\% | 1,325 | 1,026 | 1,639 |
| 28 | 10,573 | 28\% | 21\% | 52\% | 2,927 | 2,383 | 3,495 |
| 29 | 18,540 | 49\% | 22\% | 29\% | 9,137 | 8,078 | 10,202 |
| 30 | 42,029 | 60\% | 15\% | 25\% | 25,109 | 22,636 | 27,486 |
| 31 | 69,841 | 77\% | 13\% | 10\% | 53,766 | 50,233 | 57,109 |
| 32 | 36,104 | 79\% | 11\% | 10\% | 28,406 | 26,719 | 29,952 |
| 33 | 33,072 | 54\% | 23\% | 23\% | 17,738 | 15,836 | 19,596 |
| 34 | 15,126 | 53\% | 20\% | 27\% | 8,048 | 7,188 | 8,928 |
| 35-40 ${ }^{\text {b }}$ | 7,460 | 37\% | 33\% | 30\% | 2,763 | 2,350 | 3,192 |
| Total | 241,533 | 62\% | 17\% | 21\% | 149,586 | 144,305 | 154,702 |

a Other includes Chilkat River mainstem spawning stocks.
b Harvest proportions and numbers for statistical weeks 36-40 were estimated using the proportions from the last statistical week with genetic samples, in this case statistical week 35 .

## Escapement Age, Sex, and LengTh Composition

## 2016

In 2016, the sockeye salmon escapement was composed primarily of age-1.3 (84.2\%) and age-2.3 (12.7\%) fish (Table 9; Appendix N). The remainder of the escapement (3.1\%) was composed of age-1.1, age-1.2, age-1.4, and age- 2.2 fish. The mean length of age- 1.3 fish was 555 mm for males and 543 mm for females, and the mean length of age- 2.3 fish was 549 mm for males and 543 mm for females (Table 10; Appendices O and P ).

Table 9.-Age composition of the Chilkoot Lake sockeye salmon escapement, weighted by statistical week, 2016.

| Brood year and age class | 2013 | 2012 | 2011 | 2010 | 2011 | 2010 | 2009 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1.1 | 1.2 | 1.3 | 1.4 | 2.2 | 2.3 | 3.3 |  |
| Sample size | 1 | 33 | 1,376 | 2 | 9 | 207 | 1 | 1,629 |
| Escapement | 5 | 2,186 | 73,061 | 73 | 362 | 11,024 | 9 | 86,721 |
| Escapement SE | 5 | 521 | 1,214 | 52 | 133 | 1,126 | 8 |  |
| Percent | 0.0\% | 2.5\% | 84.2\% | 0.1\% | 0.4\% | 12.7\% | 0.0\% |  |
| Percent SE | 0.0\% | 0.6\% | 1.4\% | 0.1\% | 0.2\% | 1.3\% | 0.0\% |  |

Table 10.-Average length (mid eye to tail fork) of Chilkoot Lake sockeye salmon by age class and sex, 2016. (A dash indicates the age class was not present.)

| Brood year and age class | 2013 | 2012 | 2011 | 2010 | 2011 | 2010 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1.1 | 1.2 | 1.3 | 1.4 | 2.2 | 2.3 |  |
| Male |  |  |  |  |  |  |  |
| Sample size | 1 | 31 | 644 | 1 | 7 | 97 | 781 |
| Mean length (mm) | 330 | 476 | 555 | 550 | 474 | 549 | 489 |
| SE | 0.0 | 6.5 | 1.0 | 0.0 | 8.7 | 3.0 | 1.1 |
| Female |  |  |  |  |  |  |  |
| Sample size | - | 2 | 638 | 1 | 2 | 82 | 725 |
| Mean length (mm) | - | 478 | 543 | 520 | 535 | 543 | 524 |
| SE | - | 20.0 | 0.7 | 0.0 | 15.0 | 2.0 | 0.6 |
| All Fish |  |  |  |  |  |  |  |
| Sample size | 1 | 33 | 1,282 | 2 | 9 | 179 | 1,506 |
| Mean length (mm) | 330 | 476 | 549 | 535 | 487 | 547 | 487 |
| SE | 0.0 | 6.2 | 0.6 | 15.0 | 11.1 | 1.9 | 0.7 |

## 2017

In 2017, the sockeye salmon escapement was composed primarily of age-1.3 (68.0\%) and age-1.2 ( $20.2 \%$ ) fish (Table 11; Appendix N). The remainder of the escapement (11.8\%) was composed of age- 0.1 , age- 0.3 , age- 1.4 , age- 2.2 , age 2.3 , and age- 2.4 fish. The mean length of age- 1.3 fish was 559 mm for males and 547 mm for females, and the mean length of age- 1.2 fish was 484 mm for males and 496 mm for females (Table 12; Appendices O and P ).

Table 11.-Age composition of the Chilkoot Lake sockeye escapement weighted by statistical week, 2017.

| Brood year and age class | 2015 | 2013 | 2013 | 2012 | 2011 | 2012 | 2011 | 2010 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.1 | 0.3 | 1.2 | 1.3 | 1.4 | 2.2 | 2.3 | 2.4 |  |
| Sample size | 1 | 2 | 124 | 504 | 18 | 10 | 43 | 3 | 705 |
| Escapement | 117 | 55 | 8,702 | 29,286 | 737 | 799 | 3,265 | 137 | 43,098 |
| Escapement SE | 116 | 38 | 867 | 1,050 | 202 | 328 | 644 | 92 |  |
| Percent | 0.3\% | 0.1\% | 20.2\% | 68.0\% | 1.7\% | 1.9\% | 7.6\% | 0.3\% |  |
| Percent SE | 0.3\% | 0.1\% | 2.0\% | 2.4\% | 0.5\% | 0.8\% | 1.5\% | 0.2\% |  |

Table 12.-Average length (mid eye to tail fork) of Chilkoot Lake sockeye salmon by age class and sex, 2017. (A dash indicates the age class was not present.)

| Brood year and age class | 2015 | 2013 | 2013 | 2012 | 2011 | 2012 | 2011 | 2010 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.1 | 0.3 | 1.2 | 1.3 | 1.4 | 2.2 | 2.3 | 2.4 |  |
| Male |  |  |  |  |  |  |  |  |  |
| Sample size | 1 | 1 | 105 | 266 | 12 | 6 | 23 | 3 | 417 |
| Mean length (mm) | 490 | 565 | 484 | 559 | 581 | 484 | 550 | 560 | 534 |
| SE | 0.0 | 0.0 | 3.6 | 1.6 | 5.4 | 16.2 | 6.3 | 21.9 | 2.2 |
| Female |  |  |  |  |  |  |  |  |  |
| Sample size | - | 1 | 19 | 238 | 6 | 4 | 20 | - | 288 |
| Mean length (mm) | - | 510 | 496 | 547 | 565 | 502 | 538 | - | 526 |
| SE | - | 0.0 | 6.5 | 1.5 | 7.9 | 2.9 | 3.6 | - | 1.6 |
| All Fish |  |  |  |  |  |  |  |  |  |
| Sample size | 1 | 2 | 124 | 504 | 18 | 10 | 43 | 3 | 705 |
| Mean length (mm) | 490 | 538 | 486 | 554 | 576 | 490 | 544 | 560 | 530 |
| SE | 0.0 | 30.0 | 3.2 | 1.1 | 4.7 | 10.2 | 3.8 | 21.9 | 1.4 |

## 2018

In 2018, the sockeye salmon escapement was composed primarily of age-1.3 (47.5\%) and age-1.2 (47.2\%) fish (Table 13; Appendix N). The remainder of the escapement (5.3\%) was composed of age- 0.3 , age-1.1, age-1.4, age-2.3, and age- 2.4 fish. The mean length of age- 1.3 fish was 562 mm for males and 548 mm for females, and the mean length of age- 1.2 fish was 477 mm for males and 490 mm for females (Table 14; Appendices O and P).

Table 13.-Age composition of the Chilkoot Lake sockeye escapement weighted by statistical week, 2018.

| Brood year and age class | 2014 | 2015 | 2014 | 2013 | 2012 | 2012 | 2011 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.3 | 1.1 | 1.2 | 1.3 | 1.4 | 2.3 | 2.4 |  |
| Sample size | 1 | 2 | 205 | 442 | 7 | 28 | 1 | 686 |
| Escapement | 24 | 128 | 40,331 | 40,570 | 819 | 3,581 | 9 | 85,463 |
| Escapement SE | 24 | 90 | 2,885 | 2,857 | 673 | 1,198 | 9 |  |
| Percent | 0.0\% | 0.1\% | 47.2\% | 47.5\% | 1.0\% | 4.2\% | 0.0\% |  |
| Percent SE | 0.0\% | 0.1\% | 3.4\% | 3.3\% | 0.8\% | 1.4\% | 0.0\% |  |

Table 14.-Average length (mid eye to tail fork) of Chilkoot Lake sockeye salmon by age class and sex, 2018. (A dash indicates the age class was not present.)

| Brood year and age class | 2014 | 2015 | 2014 | 2013 | 2012 | 2012 | 2011 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.3 | 1.1 | 1.2 | 1.3 | 1.4 | 2.3 | 2.4 |  |
| Male |  |  |  |  |  |  |  |  |
| Sample size | - | 2 | 176 | 201 | 7 | 14 | 1 | 401 |
| Mean length (mm) | - | 333 | 477 | 562 | 562 | 576 | 565 | 525 |
| SE | - | 5.0 | 2.7 | 1.5 | 8.7 | 5.6 | 0.0 | 2.6 |
| Female |  |  |  |  |  |  |  |  |
| Sample size | 1 | - | 29 | 241 | - | 14 | - | 285 |
| Mean length (mm) | 545 | - | 490 | 548 | - | 551 | - | 541 |
| SE | 0.0 | - | 4.6 | 1.2 | - | 5.2 | - | 1.6 |
| All Fish |  |  |  |  |  |  |  |  |
| Sample size | 1 | 2 | 205 | 442 | 7 | 28 | 1 | 686 |
| Mean length (mm) | 545 | 333 | 479 | 554 | 562 | 563 | 565 | 514 |
| SE | 0.0 | 5.0 | 2.5 | 1.0 | 8.7 | 4.4 | 0.0 | 1.7 |

## 2019

In 2019, the sockeye salmon escapement was composed primarily of age-1.3 (80.8\%) and age-1.2 (17.1\%) fish (Table 15; Appendix N). The remainder of the escapement ( $2.1 \%$ ) was composed of age-1.4, age-2.2, and age-2.3 fish. The mean length of age- 1.3 fish was 562 mm for males and 544 mm for females, and the mean length of age-1.2 fish was 479 mm for males and 478 mm for females (Table 16; Appendices O and P).

Table 15.-Age composition of the Chilkoot Lake sockeye escapement weighted by statistical week, 2019.

| Brood year and age class | 2015 | 2014 | 2013 | 2014 | 2013 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1.2 | 1.3 | 1.4 | 2.2 | 2.3 |  |
| Sample size | 92 | 700 | 2 | 4 | 13 | 811 |
| Escapement | 23,987 | 113,393 | 407 | 557 | 2,034 | 140,378 |
| Escapement SE | 3,141 | 3,252 | 392 | 295 | 966 |  |
| Percent | 17.1\% | 80.8\% | 0.3\% | 0.4\% | 1.4\% |  |
| Percent SE | 2.2\% | 2.3\% | 0.3\% | 0.2\% | 0.7\% |  |

Table 16.-Average length (mid eye to tail fork) of Chilkoot Lake sockeye salmon by age class and sex, 2019.

| Brood year | 2015 | 2014 | 2013 | 2014 | 2013 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1.2 | 1.3 | 1.4 | 2.2 | 2.3 | Total |
| Male |  |  |  |  |  |  |
| Sample size | 54 | 296 | 1 | 3 | 5 | 359 |
| Mean length (mm) | 479 | 562 | 605 | 488 | 537 | 549 |
| SE | 4.8 | 1.5 | 0.0 | 15.3 | 13.0 | 2.1 |
| Female |  |  |  |  |  |  |
| Sample size | 38 | 404 | 1 | 1 | 8 | 452 |
| Mean length (mm) | 478 | 544 | 554 | 453 | 532 | 538 |
| SE | 4.6 | 1.2 | 0.0 | 0.0 | 10.9 | 1.5 |
| All Fish |  |  |  |  |  |  |
| Sample size | 92 | 700 | 2 | 4 | 13 | 811 |
| Mean length (mm) | 479 | 552 | 580 | 479 | 534 | 525 |
| SE | 3.4 | 1.0 | 30.0 | 14.7 | 8.0 | 1.3 |

## Fry Population Estimate

Hydroacoustic surveys were conducted at Chilkoot Lake on 26 October 2016, 16 November 2017, 2 October 2018, and 31 October 2019 (Table 17). The precision of pelagic fish estimates met our objective for a CV $\leq 15 \%$ in all four years. In 2016, the pelagic fish population was estimated to be 1,294,334 fish (CV $=4 \%$ ), about 42\% above the long-term average (1987-2015; 908,000 fish). The 2017 estimate of 491,901 fish ( $\mathrm{CV}=5 \%$ ) was the smallest since 2007 and about $45 \%$ below average. Pelagic fish populations improved to about average in 2018 (919,761 fish; CV = 11\%) but were about $21 \%$ below average in 2019 ( 719,165 fish; CV $=8 \%$ ). Trawl surveys conducted on 6 November 2019 to check species composition resulted in a catch of 107 sockeye salmon fry; no other species of fish were captured.

Table 17.-Number of fish collected in trawl samples by species, percentage of sockeye salmon in trawl samples, and estimated total number of fish (hydroacoustic targets) and sockeye salmon fry in autumn surveys of Chilkoot Lake, 1987-2019.

| Year | Trawl samples |  |  |  | Percent sockeye | Hydroacoustic estimates |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total fish | Sockeye | Stickleback | Other |  | Targets | CV | Sockeye |
| 1987 | 194 | 141 | 41 | 12 | 73\% | 1,344,951 | ND | 977,516 |
| 1988 | 85 | 83 | 0 | 2 | 98\% | 3,066,118 | ND | 2,993,974 |
| 1989 | 209 | 208 | 1 | 0 | 100\% | 874,794 | ND | 870,608 |
| 1990 | 240 | 238 | 0 | 2 | 99\% | 607,892 | ND | 602,826 |
| 1991 | 47 | 38 | 9 | 0 | 81\% | 475,404 | ND | 384,369 |
| 1992 | ND | ND | ND | ND | ND | ND | ND | ND |
| 1993 | ND | ND | ND | ND | ND | ND | ND | ND |
| 1994 | ND | ND | ND | ND | ND | ND | ND | ND |
| 1995 | 775 | 708 | 52 | 15 | 91\% | 260,797 | ND | 238,250 |
| 1996 | 174 | 173 | 0 | 1 | 99\% | 418,152 | ND | 415,749 |
| 1997 | 117 | 116 | 0 | 1 | 99\% | 637,628 | ND | 632,178 |
| 1998 | 526 | 523 | 0 | 3 | 99\% | 1,309,711 | ND | 1,302,241 |
| 1999 | 263 | 248 | 11 | 4 | 94\% | 400,307 | ND | 377,476 |
| 2000 | 15 | 14 | 0 | 1 | 93\% | 1,380,950 | ND | 1,288,887 |
| 2001 | 61 | 29 | 23 | 9 | 48\% | 1,351,068 | ND | 642,311 |
| 2002 | 289 | 288 | 1 | 0 | 100\% | 1,389,712 | 4\% | 1,384,903 |
| 2003 | 139 | 138 | 1 | 0 | 99\% | 1,384,754 | ND | 1,384,754 |
| 2004 | 199 | 187 | 4 | 8 | 94\% | 1,059,963 | 10\% | 996,200 |
| 2005 | 25 | 25 | 0 | 0 | 100\% | 247,283 | 22\% | 247,283 |
| 2006 | 80 | 80 | 0 | 0 | 100\% | 356,957 | 17\% | 356,957 |
| 2007 | 48 | 48 | 0 | 0 | 100\% | 99,781 | 6\% | 99,781 |
| 2008 | 534 | 531 | 1 | 2 | 99\% | 1,020,388 | 14\% | 1,014,655 |
| 2009 | 60 | 60 | 0 | 0 | 100\% | 832,991 | 14\% | 832,991 |
| 2010 | 379 | 379 | 0 | 0 | 100\% | 741,537 | 5\% | 741,537 |
| 2011 | 82 | 82 | 0 | 0 | 100\% | 651,847 | 24\% | 651,847 |
| 2012 | 142 | 142 | 0 | 0 | 100\% | 752,212 | 13\% | 752,212 |
| 2013 | 131 | 131 | 0 | 0 | 100\% | 642,256 | 6\% | 642,256 |
| 2014 | 551 | 546 | 0 | 5 | 99\% | 1,160,985 | 8\% | 1,150,450 |
| 2015 | ND | ND | ND | ND | ND | 1,148,335 | 7\% | 1,148,335 |
| 2016 | ND | ND | ND | ND | ND | 1,294,334 | 4\% | 1,294,334 |
| 2017 | ND | ND | ND | ND | ND | 491,901 | 5\% | 491,901 |
| 2018 | ND | ND | ND | ND | ND | 919,761 | 11\% | 919,761 |
| 2019 | 107 | 107 | 0 | 0 | 100\% | 719,165 | 8\% | 719,165 |

Note: Bold estimates are historical records that have been updated since the last project report by Bednarski et al. (2016). No hydroacoustic surveys were conducted from 1992 to 1994.

## Limnological Assessment

## Light and Temperature Profiles

In most years, the euphotic zone depth in Chilkoot Lake was deepest at the beginning of the sampling season (May), gradually became shallower as the season progressed, and increased again in October. In 2016, the average euphotic zone depth ranged from 8.3 m in May to 2.7 m in August and averaged 4.9 m for the season (Table 18). In 2017, the average euphotic zone depth ranged from 13.9 m in May to 3.4 m in September and averaged 6.1 m for the season. In 2018, the average euphotic zone depth ranged from 11.6 m in June to 2.3 m in August and averaged 6.3 m for the
season. In 2019, the euphotic zone depth was remarkably shallow throughout the season, particularly in the spring. The average euphotic zone depth in 2019 ranged from 4.2 m in October to 1.4 m in September and averaged 2.2 m for the season.
In all years (2016-2019), weak thermoclines (the depths at which temperature change was $>1^{\circ} \mathrm{C}$ per m) were detected in only 1 or 2 months between May and September and only to 3 or 4 m below the surface (Figure 4). The maximum lake surface temperature recorded per season was $13.3^{\circ} \mathrm{C}$ on 13 July $2016,11.0^{\circ} \mathrm{C}$ on 14 July $2017,13.6^{\circ} \mathrm{C}$ on 18 August 2018 , and $16.5^{\circ} \mathrm{C}$ on 16 August 2019.

## Zooplankton Composition

Zooplankton samples from Chilkoot Lake were composed predominantly of copepods (Cyclops sp.) in all years (Tables 19 and 20). Despite wide fluctuations during this period, average zooplankton density and biomass (standardized to May-September period) were nearly double the long-term average. Seasonal mean zooplankton density in 2019 ( $290,000 \mathrm{no} . / \mathrm{m}^{2}$ ) and mean biomass in $2016\left(570 \mathrm{mg} / \mathrm{m}^{2}\right)$ were the largest recorded since sampling began in 1987 (Figure 5; Appendix Q). No zooplankton samples were collected in August 2018, making it difficult to compare this year directly to the other years; however, examination of the months that were sampled shows that zooplankton populations were at relatively high levels (Table 19).

Table 18.-Euphotic zone depths (m) in Chilkoot Lake, 2016-2019.

| Year | Date | Station 1A | Station 2A | Average |
| :---: | :---: | :---: | :---: | :---: |
| 2016 | 18-May | 8.5 | 8.1 | 8.3 |
|  | 21-Jun | 5.1 | 4.8 | 4.9 |
|  | 13-Jul | 4.5 | 2.7 | 3.6 |
|  | 17-Aug | 3.2 | 2.2 | 2.7 |
|  | 17-Sep | 3.9 | 3.1 | 3.5 |
|  | 28-Oct | 5.8 | 7.0 | 6.4 |
|  | Average (May-Oct.) | 5.2 | 4.7 | 4.9 |
| 2017 | 19-May | 13.8 | 14.0 | 13.9 |
|  | 18-Jun | 5.6 | 5.8 | 5.7 |
|  | 14-Jul | 5.2 | 5.5 | 5.3 |
|  | 17-Aug | 3.8 | 4.2 | 4.0 |
|  | 15-Sep | 3.0 | 3.8 | 3.4 |
|  | 16-Oct | 4.1 | 4.1 | 4.1 |
|  | Average (May-Oct.) | 5.9 | 6.2 | 6.1 |
| 2018 | 16-May | 10.6 | 11.8 | 11.2 |
|  | 13-Jun | 13.5 | 9.6 | 11.6 |
|  | 18-Jul | 4.0 | 3.4 | 3.7 |
|  | 18-Aug | 2.2 | 2.4 | 2.3 |
|  | 16-Sep | 1.5 | 4.0 | 2.7 |
|  | October | ND | ND | ND |
|  | Average (May-Oct.) | 6.3 | 6.3 | 6.3 |
| 2019 | 15-May | 1.2 | 3.3 | 2.2 |
|  | June | ND | ND | ND |
|  | 17-Jul | 1.7 | 1.9 | 1.8 |
|  | 16-Aug | 1.5 | 1.6 | 1.6 |
|  | 16-Sep | 1.3 | 1.5 | 1.4 |
|  | 17-Oct | 4.0 | 4.5 | 4.2 |
|  | Average (May-Oct.) | 1.9 | 2.6 | 2.2 |



Figure 4.-Water temperature profiles by date (averaged between stations 1A and 2A) at Chilkoot Lake, 2016-2019.

Table 19.-Mean density of zooplankton per $\mathrm{m}^{2}$ of lake surface area, by sampling date and taxon, in Chilkoot Lake 2016-2019. Density estimates were the average of 2 sampling stations. Ovigerous (ovig.) individuals were separated from non-egg bearing individuals.

| Year | Taxon/Date | Macrozooplankton density (number $/ \mathrm{m}^{2}$ ) by sampling date |  |  |  |  |  |  | Seasonal mean |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 20-Apr | 18-May | 21-Jun | 13-Jul | 17-Aug | 17-Sep | 28-Oct | Density | \% Density |
| 2016 | Cyclops sp. | 530,226 | 300,561 | 139,668 | 68,390 | 71,659 | 91,696 | 123,620 | 189,403 | 86\% |
|  | Ovig. Cyclops | 0 | 1,698 | 2,123 | 5,604 | 17,830 | 43,301 | 11,887 | 11,777 | 5\% |
|  | Nauplii | 39,905 | 849 | 1,274 | 2,165 | 6,580 | 16,132 | 70,131 | 70,131 | 9\% |
|  | Total | 570,131 | 303,108 | 143,064 | 76,159 | 96,069 | 151,129 | 205,638 | 220,757 |  |
| 2017 |  | Apr | 19-May | 18-Jun | 14-Jul | 17-Aug | 15-Sep | Oct | Density | \% Density |
|  | Cyclops sp. | ND | 240,406 | 113,771 | 137,629 | 73,612 | 181,525 | ND | 149,389 | 82\% |
|  | Ovig. Cyclops | ND | 0 | 9,594 | 29,759 | 12,226 | 29,334 | ND | 16,183 | 9\% |
|  | Nauplii | ND | 11,420 | 1,613 | 5,986 | 2,038 | 62,447 | ND | 16,701 | 9\% |
|  | Total | ND | 251,825 | 124,979 | 173,374 | 87,876 | 273,306 | ND | 182,272 |  |
| 2018 |  | Apr | 18-May | 13-Jun | 18-Jul | Aug | 16-Sep | Oct | Density | \% Density |
|  | Cyclops sp. | ND | 164,120 | 249,193 | 143,488 | ND | 88,589 | ND | 161,347 | 83\% |
|  | Ovig. Cyclops | ND | 0 | 0 | 11,717 | ND | 6,673 | ND | 4,598 | 2\% |
|  | Nauplii | ND | 26,830 | 5,837 | 2,547 | ND | 55,400 | ND | 22,653 | 12\% |
|  | Bosmina | ND | 0 | 0 | 5,094 | ND | 2,725 | ND | 1,955 | 1\% |
|  | Daphnia sp. | ND | 0 | 0 | 7,641 | ND | 1,987 | ND | 2,407 | 1\% |
|  | Daphnia rosea | ND | 0 | 0 | 0 | ND | 1,274 | ND | 318 | <1\% |
|  | Total | ND | 130,752 | 320,937 | 220,751 | ND | 156,648 | ND | 193,279 |  |
| 2019 |  | Apr | 14-May | 13-Jun | 17-Jul | 16-Aug | 16-Sep | 17-Oct | Density | \% Density |
|  | Cyclops sp. | ND | 343,352 | 185,940 | 286,127 | 184,412 | 264,137 | 145,950 | 234,986 | 86\% |
|  | Ovig. Cyclops | ND | 0 | 0 | 0 | 0 | 1,019 | 255 | 212 | <1\% |
|  | Nauplii | ND | 44,999 | 509 | 679 | 75,310 | 53,235 | 30,990 | 34,287 | 13\% |
|  | Bosmina | ND | 0 | 0 | 0 | 2,802 | 3,566 | 2,632 | 1,500 | 1\% |
|  | Ovig. Bosmina | ND | 0 | 0 | 0 | 0 | 0 | 255 | 42 | $<1 \%$ |
|  | Daphnia sp. | ND | 0 | 0 | 0 | 764 | 4,075 | 1,443 | 1,047 | <1\% |
|  | Daphnia rosea | ND | 0 | 0 | 0 | 0 | 509 | 0 | 85 | <1\% |
|  | Total | ND | 388,351 | 186,449 | 286,806 | 263,287 | 326,541 | 181,525 | 272,160 |  |

Table 20.-Mean length and biomass of zooplankton by sampling date and taxon in Chilkoot Lake, 20162019. Biomass estimates were the average of the 2 sampling stations. Ovigerous (ovig.) individuals were separated from non-egg bearing individuals. (A dash indicates the taxon was not present.)

| Year | Taxon/Date | Macrozooplankton length (mm) by sampling date |  |  |  |  |  | Seasonal means (weighted) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 18-May | 21-Jun | 13-Jul | 17-Aug | 17-Sep | 28-Oct | Length (mm) | $\begin{aligned} & \hline \begin{array}{l} \text { Biomass } \\ \left(\mathrm{mg} / \mathrm{m}^{2}\right) \end{array} \end{aligned}$ | $\begin{gathered} \% \\ \text { biomass } \end{gathered}$ |
| 2016 | Cyclops sp. | 0.91 | 1.11 | 1.17 | 1.06 | 0.90 | 0.83 | 0.82 | 443 | 85\% |
|  | Ovig. Cyclops | 1.24 | 1.39 | 1.28 | 1.33 | 1.25 | 1.34 | 1.31 | 76 | 15\% |
|  | Total |  |  |  |  |  |  |  | 519 |  |
| 2017 |  | 19-May | 18-Jun | 14-Jul | 17-Aug | 15-Sep | Oct | Length (mm) | Biomass $\left(\mathrm{mg} / \mathrm{m}^{2}\right)$ | $\begin{gathered} \hline \% \\ \text { biomass } \\ \hline \end{gathered}$ |
|  | Cyclops sp. | 0.56 | 0.92 | 1.06 | 1.23 | 0.73 | ND | 0.81 | 376 | 80\% |
|  | Ovig. Cyclops | - | 1.13 | 1.14 | 1.29 | 1.34 | ND | 1.23 | 92 | 20\% |
|  | Total |  |  |  |  |  |  |  | 467 |  |
| 2018 |  | 18-May | 13-Jun | 18-Jul | Aug | 16-Sep | Oct | $\begin{gathered} \text { Length } \\ (\mathrm{mm}) \end{gathered}$ | $\begin{aligned} & \text { Biomass } \\ & \left(\mathrm{mg} / \mathrm{m}^{2}\right) \end{aligned}$ | $\begin{gathered} \hline \% \\ \text { biomass } \end{gathered}$ |
|  | Cyclops sp. | 0.53 | 0.85 | 1.10 | ND | 0.68 | ND | 0.80 | 363 | 91\% |
|  | Ovig. Cyclops | - | - | 1.33 | ND | 1.30 | ND | 1.31 | 29 | 7\% |
|  | Bosmina | - | - | 0.28 | ND | 0.34 | ND | 0.29 | <1 | <1\% |
|  | Daphnia sp. | - | - | 0.95 | ND | 0.71 | ND | 0.62 | 4 | 1\% |
|  | Daphnia rosea | - | - | - | ND | 1.05 | ND | 1.05 | 1 | $<1 \%$ |
|  |  |  |  |  |  |  |  |  | 398 |  |
| 2019 |  | 14-May | 13-Jun | 17-Jul | 16-Aug | 16-Sep | 17-Oct | Length (mm) | $\begin{aligned} & \text { Biomass } \\ & \left(\mathrm{mg} / \mathrm{m}^{2}\right) \end{aligned}$ | $\begin{gathered} \% \\ \text { biomass } \\ \hline \end{gathered}$ |
|  | Cyclops | 0.54 | 0.86 | 1.07 | 0.86 | 0.70 | 0.59 | 0.77 | 487 | 99\% |
|  | Ovig. Cyclops | - | - | - | - | 1.29 | - | 1.12 | 1 | $<1 \%$ |
|  | Bosmina | - | - | - | 0.41 | 0.31 | 0.36 | 0.25 | 1 | $<1 \%$ |
|  | Ovig. Bosmina | - | - | - | - | - | 0.50 | 0.50 | $<1$ | $<1 \%$ |
|  | Daphnia rosea | - | - | - | 1.07 | 1.12 | 1.06 | 1.01 | 4 | 1\% |
|  | Ovig. D. rosea | - | - | - | - | 1.30 | - | 1.30 | 1 | <1\% |
|  | Total |  |  |  |  |  |  |  | 493 |  |



Figure 5.-Annual (standardized to May-September period) zooplankton density and biomass in Chilkoot Lake, 1987-2019. Estimates not included for 1992-1994 (no samples were collected), 1995 (samples were only collected in the months of June-August), or 2018 (no samples collected in August).

## DISCUSSION

Chilkoot Lake sockeye salmon escapements annually met or exceeded the current escapement goal range of $38,000-86,000$ fish during 2016-2019. The 2019 escapement of 140,378 fish was the largest recorded since the project first started in 1976. However, total runs (escapement plus District 15 fishery harvest) fluctuated dramatically over the 4 -year period, and the 2017 run of 47,366 fish was the fifth lowest on record (Figure 6; Appendix M). Total runs in 2016 (211,830 fish) and 2018 ( 123,997 fish) were closer to the average for this system and fell into the 67th and 44th percentiles, respectively. The 2019 run ( 293,723 fish) was in the 86th percentile and was the largest run since 1991. Harvest rates on Chilkoot Lake sockeye salmon (including commercial, subsistence, and sport harvest) averaged $38 \%$ over the four years 2016-2019, which was below the long-term average of $48 \%$. The poor run in 2017 required conservative fishery management in order to meet the escapement goal, which limited the harvest rate to $9 \%$, far below the harvest rates in $2016(59 \%), 2018(31 \%)$, and 2019 (52\%).


Figure 6.-Estimated total runs (escapement plus District 15 fishery harvest) of Chilkoot Lake sockeye salmon, 1976-2019. Harvest includes commercial, sport, and subsistence.

## SOURCES OF UNCERTAINTY

Total Chilkoot sockeye salmon run estimates presented in this report are defined as the annual escapement plus terminal subsistence, sport, and commercial (District 15) harvests. The total run estimates represent minimum point estimates and currently do not incorporate sources of uncertainty, including 1) variability in the annual escapement estimate (e.g., interpolation for missed days, fish escaping into the lake after the weir is removed); 2) inconsistent or lack of reporting of subsistence and sport harvest; 3) uncertainty in harvest estimates generated through visual scale pattern analysis (e.g., in 2016) or GSI (2017-2019); 4) unaccounted for incidental commercial fishing mortality (Patterson et al. 2017); 5) variability in the commercial harvest estimates through the weight-to-numbers conversion on fish tickets; and, in particular, 6)
unaccounted for commercial harvest of Chilkoot sockeye salmon outside of District 15. Much of this uncertainty is probably minimal, with the exception of unaccounted for harvest outside of District 15, which would require genetic stock identification to be conducted for those fisheries (Gilk-Baumer et al. 2015; Miller and Heinl 2018).

## District 15 Management

The District 15 commercial drift gillnet fishery has been managed in accordance with the Lynn Canal and Chilkat River King Salmon Fishery Management Plan (5 AAC 33.384) since 2003. The overall management goal is to achieve desired spawning escapement levels while harvesting the available surplus for a long-term maximum sustainable yield of all Lynn Canal salmon stocks. Management decisions are guided by inseason run projections based on daily weir counts and stock composition information from the fishery. Openings early in the season are typically designed to harvest large hatchery runs of summer chum salmon in Section 15-C (central Lynn Canal; Figure 1) while minimizing the harvest of northbound sockeye salmon and other wild stocks until run strength can be determined.
In 2018, the Alaska Board of Fisheries designated the Chilkat River Chinook salmon run as a stock of management concern after multiple years of failing to achieve the Chinook salmon escapement goal. The board adopted the Chilkat River and King Salmon River King Salmon Stock Status and Action Plan, 2018 (Lum and Fair 2018), which outlined management measures intended to reduce the harvest rate on Chilkat River Chinook salmon stocks and rebuild the run to consistently achieve escapements within the escapement goal range. In 2018, management of the District 15 commercial drift gillnet fishery followed the action plan; however, in 2019, additional time and area restrictions beyond those prescribed in the action plan were implemented (Thynes et al. 2020).

The management actions taken to reduce harvest of Chilkat River Chinook salmon subsequently limited harvest opportunities for hatchery chum and wild sockeye salmon in 2018 and 2019. During years of high Chilkoot Lake sockeye salmon abundance, additional time and area in Section $15-\mathrm{A}$ are normally granted north of the latitude of Mud Bay Point (Figure 1), and during very strong years, such as 2019, Lutak Inlet (Figure 1) is also usually opened to harvest fish surplus to escapement needs. In 2019, the Chilkoot Lake sockeye salmon escapement exceeded the upper bound of the weekly management targets during statistical week 29 and reached the upper bound of the escapement goal range of $38,000-86,000$ fish by the end of statistical week 30. Due to Chilkat River Chinook salmon conservation measures outlined in the action plan (Lum and Fair 2018; Thynes et al. 2020), restrictions could not be liberalized in Section 15-A until after the fifth week of the fishery (statistical week 29). In statistical week 30, the open area was extended north to Talsani Island, and by the seventh week of the fishery (statistical week 31) the open area was extended to the terminus of the Chilkoot River. Due to the delay, the gillnet fleet could not effectively intercept the run and the Chilkoot Lake sockeye salmon escapement goal was greatly exceeded.

## Reduced Size and Growth of Sockeye Salmon

During 2015-2019, Chilkoot Lake sockeye salmon in each of the major age classes (ages 1.2, 1.3, 2.2, and 2.3) were smaller than the historical average, with the exception of age- 1.2 males in 2017 and age-2.2 females in 2016 (Appendices O and P ). Ocean-age-2 male sockeye salmon (ages 1.2 and 2.2 combined; Figure 7) appeared to have recovered to near average size from the notably small size observed in 2015 (Bednarski et al. 2016). Ocean-age-2 females were larger than average
in 2016 but subsequently decreased in size. Ocean-age-3 sockeye salmon (ages 1.3 and 2.3 combined) of both sexes were the smallest in the entire time series, 1982-2019 (Figure 7).

The mechanism responsible for the reduced size and growth remains poorly understood, but the widespread nature of the decline suggests that the mechanism is large and affects broad ocean communities. After 2010, sockeye salmon runs across all 4 regions of Alaska declined in average body size, and a $2.1 \%$ decrease was documented in Southeast Alaska sockeye salmon (Oke et al. 2020). The small size of Chilkoot and other sockeye salmon stocks starting in 2015 (Bednarski et al. 2016; Brunette and Piston 2019) was thought to be a product of the "blob", a warm water anomaly that developed off the Gulf of Alaska in 2013 (Bond et al. 2015; Di Lorenzo and Mantua 2016). The discovery of a second, warmer "blob" in the Gulf of Alaska in 2019 (Amaya et al. 2020) suggests that continued decreases in the size and number of Chilkoot Lake sockeye salmon may occur in future years.


Figure 7.-Average annual sockeye salmon mid eye to fork length by sex and ocean age for the major age classes (ages 1.2, 1.3, 2.2, and 2.3) in the Chilkoot Lake escapement compared to the 1982-2019 averages (horizontal lines). Ocean-age-2 refers to age classes 1.2 and 2.2, and ocean-age- 3 refers to age classes 1.3 and 2.3.

Previous estimates of rearing sockeye salmon populations in Chilkoot Lake suggested good runs of adult sockeye salmon during 2017-2019 (Bednarski et al. 2016). Although this prediction held true for the large run in 2019, the 2017 sockeye salmon run was one of the smallest on record. The fall fry population in 2016 ( $1,294,000$ fish) was the largest in 13 years and $42 \%$ greater than the long-term average of 908,000 fish. However, by 2017 the rearing population had declined to 492,000 fish, and the lake held relatively moderate numbers of rearing fry in 2018 ( 920,000 fish $)$ and 2019 ( 719,000 fish), despite recording the highest numbers and biomass of zooplankton since 1987 (Figure 5). Although there has been no relationship (adjusted $R^{2}=<0.01 ; p$-value $=0.31$ ) between the size of the spawning escapement in the parent year and the fall fry population 1 year later, there is a very weak positive correlation (adjusted $R^{2}=0.25 ; p$-value $<0.01$ ) between the size of the fall fry population and subsequent adult returns (Figure 8). Some known sources of error may be weakening the correlation. For example, we assumed that all sockeye salmon fry are age-1, which is not true; however, a very large portion (average $=82 \%$ ) of the adult return (by brood year) to Chilkoot Lake spent only 1 year in freshwater. Further research may uncover a stronger correlation between rearing fry abundance and adult returns.


Figure 8.-Comparison of Chilkoot Lake sockeye salmon parent year escapement (1986-1990; 19942018) to the rearing fry population (1987-1991; 1995-2019) 1 year later (left), and comparison of the rearing fry population (1987-1991; 1995-2015) to the subsequent adult return, brood years 1986-1990; 1994-2014 (right). No hydroacoustic surveys were conducted during 1992-1994. The adjusted $R^{2}$ and pvalues ( $p$ ) from the regression are shown on each figure.

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## APPENDICES

The weekly sockeye salmon age-sex distribution, the seasonal age-sex distribution weighted by week, and the mean length by age and sex weighted by week, were calculated using equations from Cochran (1977).
Let

| $h$ | $=\quad$ index of the stratum (week), |
| :--- | :--- |
| $j$ | $=\quad$ index of the age class, |
| $p_{h j}$ | $=\quad$ proportion of the sample taken during stratum $h$ that is age $j$, |
| $n_{h}$ | $=\quad$ number of fish sampled in week $h$, and |
| $n_{h j}$ | $=\quad$ number observed in class $j$, week $h$. |

Then the age distribution was estimated for each week of the escapement in the usual manner:

$$
\begin{equation*}
\hat{p}_{h j}=n_{h j} / n_{h} . \tag{1}
\end{equation*}
$$

If $N_{h}$ equals the number of fish in the escapement in week $h$, standard errors of the weekly age class proportions are calculated in the usual manner (Cochran 1977, page 52, equation 3.12):

$$
\begin{equation*}
S E\left(\hat{p}_{h j}\right)=\sqrt{\left[\frac{\left(\hat{p}_{h j}\right)\left(1-\hat{p}_{h j}\right)}{n_{h}-1}\right]\left[1-n_{h} / N_{h}\right]} . \tag{2}
\end{equation*}
$$

The age distributions for the total escapement were estimated as a weighted sum (by stratum size) of the weekly proportions. That is,

$$
\begin{equation*}
\hat{p}_{j}=\sum_{h} p_{h j}\left(N_{h} / N\right) \tag{3}
\end{equation*}
$$

such that $N$ equals the total escapement. The standard error of a seasonal proportion is the square root of the weighted sum of the weekly variances (Cochran 1977, pages 107-108):

$$
\begin{equation*}
S E\left(\hat{p}_{j}\right)=\sqrt{\sum_{j}^{h}\left[S E\left(\hat{p}_{h j}\right)\right]^{2}\left(N_{h} / N\right)^{2}} \tag{4}
\end{equation*}
$$

The mean length, by sex and age class (weighted by week of escapement), and the variance of the weighted mean length, were calculated using the following equations from Cochran (1977, pages 142-144) for estimating means over subpopulations. That is, let $i$ equal the index of the individual fish in the age-sex class $j$, and $y_{h i j}$ equal the length of the $i$ th fish in class $j$, week $h$, so that,

$$
\begin{gather*}
\hat{\bar{Y}}_{j}=\frac{\sum_{h}\left(N_{h} / n_{h}\right) \sum_{i} y_{h i j}}{\sum_{h}\left(N_{h} / n_{h}\right) n_{h j}} \text {, and }  \tag{5}\\
\hat{V}\left(\hat{Y}_{j}\right)=\frac{1}{\hat{N}_{j}^{2}} \sum_{h} \frac{N_{h}^{2}\left(1-n_{h} / N_{h}\right)}{n_{h}\left(n_{h}-1\right)}\left[\sum_{i}\left(y_{h i j}-\bar{y}_{h j}\right)^{2}+n_{h j}\left(1-\frac{n_{h j}}{n_{h}}\right)\left(\bar{y}_{h j}-\hat{Y}_{j}\right)^{2}\right] .
\end{gather*}
$$

Appendix B.-ADF\&G statistical weeks, 2016-2019.

| Statistical week | 2016 |  | 2017 |  | 2018 |  | 2019 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Beginning | Ending | Beginning | Ending | Beginning | Ending | Beginning | Ending |
| 23 | 29-May | 4-Jun | 4-Jun | 10-Jun | 3-Jun | $9-J u n$ | 2-Jun | 8-Jun |
| 24 | 5-Jun | 11-Jun | 11-Jun | 17-Jun | 10-Jun | 16-Jun | $9-\mathrm{Jun}$ | 15-Jun |
| 25 | 12-Jun | 18-Jun | 18-Jun | 24-Jun | 17-Jun | 23-Jun | 16-Jun | 22-Jun |
| 26 | 19-Jun | 25-Jun | 25-Jun | 1-Jul | 24-Jun | 30-Jun | 23-Jun | 29-Jun |
| 27 | 26-Jun | 2-Jul | 2-Jul | 8-Jul | 1-Jul | 7-Jul | 30-Jun | 6-Jul |
| 28 | 3-Jul | 9 -Jul | 9-Jul | 15-Jul | 8-Jul | 14-Jul | 7-Jul | 13-Jul |
| 29 | 10-Jul | 16-Jul | 16-Jul | 22-Jul | 15-Jul | 21-Jul | 14-Jul | 20-Jul |
| 30 | 17-Jul | 23-Jul | 23-Jul | 29-Jul | 22-Jul | 28-Jul | 21-Jul | 27-Jul |
| 31 | 24-Jul | 30-Jul | 30-Jul | 5-Aug | 29-Jul | 4-Aug | 28-Jul | 3-Aug |
| 32 | 31-Jul | 6-Aug | 6-Aug | 12-Aug | 5-Aug | 11-Aug | 4-Aug | 10-Aug |
| 33 | 7-Aug | 13-Aug | 13-Aug | 19-Aug | 12-Aug | 18-Aug | 11-Aug | 17-Aug |
| 34 | 14-Aug | 20-Aug | 20-Aug | 26-Aug | 19-Aug | 25-Aug | 18-Aug | 24-Aug |
| 35 | 21-Aug | 27-Aug | 27-Aug | 2-Sep | 26-Aug | 1-Sep | 25-Aug | 31-Aug |
| 36 | 28-Aug | 3-Sep | 3-Sep | 9-Sep | 2-Sep | 8-Sep | 1-Sep | 7-Sep |
| 37 | 4-Sep | 10-Sep | 10-Sep | 16-Sep | 9-Sep | 15-Sep | 8-Sep | 14-Sep |
| 38 | 11-Sep | 17-Sep | 17-Sep | 23-Sep | 16-Sep | 22-Sep | 15-Sep | 21-Sep |
| 39 | 18-Sep | 24-Sep | 24-Sep | 30-Sep | 23-Sep | 29-Sep | 22-Sep | 28-Sep |
| 40 | 25-Sep | 1-Oct | 1-Oct | 7-Oct | 30-Sep | 6-Oct | 29-Sep | 5-Oct |
| 41 | 2-Oct | 8-Oct | 8-Oct | 14-Oct | 7-Oct | 13-Oct | 6-Oct | 12-Oct |
| 42 | 9-Oct | 15-Oct | 15-Oct | 21-Oct | 14-Oct | 20-Oct | 13-Oct | 19-Oct |

Appendix C.-ADF\&G collection code, location, reporting group, and the number ( $n$ ) of sockeye salmon used in the genetic baseline for mixed stock analysis in District 15 commercial drift gillnet fishery (Zeiser et al. 2019).

| ADF\&G collection code | Location | Reporting group | $n$ |
| :---: | :---: | :---: | :---: |
| SCKAT07E | Chilkat Lake07 Early | Chilkat Lake | 95 |
| SCKAT07L | Chilkat Lake07 Late | Chilkat Lake | 95 |
| SCKAT13 | Chilkat Lake 13 | Chilkat Lake | 189 |
| SBEARFL07 | Bear Flats - Chilkat | Chilkat Mainstem | 95 |
| SMULE03.SMULE07 | Mule Meadows - Chilkat | Chilkat Mainstem | 190 |
| SMOSQ07 | Mosquito Lake - Chilkat | Chilkat Mainstem | 95 |
| SCHIK03 | Chilkoot River | Chilkoot | 159 |
| SCHILBC07 | Chilkoot Lake - Bear Creek | Chilkoot | 233 |
| SCHILB07 | Chilkoot Lake - beaches | Chilkoot | 251 |
| SLACE13 | Lace River | Juneau Mainland | 63 |
| SBERN03.SBERN13 | Berners Bay | Juneau Mainland | 165 |
| SANTGILK13 | Antler-Gilkey River | Juneau Mainland | 53 |
| SWIND03.SWIND07 | Windfall Lake | Juneau Mainland | 142 |
| SSTEE03 | Steep Creek | Juneau Mainland | 91 |
| SAUKE13baseline.SLAKECR14 | Lake Creek (Auke Creek Weir) | Juneau Mainland | 318 |
| SKUTH06 | Kuthai Lake | Taku River/Stikine Mainstem | 171 |
| SKSLK10.SKSLK11 | King Salmon Lake | Taku River/Stikine Mainstem | 214 |
| SLTRA90.SLTRA06 | Little Trapper Lake | Taku River/Stikine Mainstem | 237 |
| SLTAT11 | Little Tatsamenie 11 | Taku River/Stikine Mainstem | 59 |
| STATS05.STATS06 | Tatsamenie Lake | Taku River/Stikine Mainstem | 288 |
| SHACK08 <br> SNAHL03.SNAHL07. | Hackett River | Taku River/Stikine Mainstem | 52 |
| SNAHL12 | Nahlin River | Taku River/Stikine Mainstem | 179 |
| STAKU07 | Taku River Taku Mainstem - | Taku River/Stikine Mainstem | 95 |
| STAKWA09 | Takwahoni/Sinwa | Taku River/Stikine Mainstem | 67 |
| SSUSTA08.SSHUST09 <br> STUCH08.SCHUNK09.STUSK08.SBEARSL09. | Shustahini Slough | Taku River/Stikine Mainstem | 185 |
| STUSKS08.STUSKS09 <br> SYELLB08.SYELLB10. | Tuskwa/Chunk Slough | Taku River/Stikine Mainstem | 356 |
| SYELLB11 <br> STULS07.STULS08. | Yellow Bluff Slough | Taku River/Stikine Mainstem | 81 |
| STULS09 | Tulsequah River | Taku River/Stikine Mainstem | 156 |
| SFISHCR09.SFISHCR10 | Fish Creek | Taku River/Stikine Mainstem | 160 |
| SYEHR07.SYEHR09 | Yehring Creek | Taku River/Stikine Mainstem | 171 |
| SCHUT08 | Chutine River | Taku River/Stikine Mainstem | 94 |
| SCHUTL09.SCHUT11 <br> SFOWL07.SFOWL08.SFOWL09.SANDY07. | Chutine Lake | Taku River/Stikine Mainstem | 224 |
| SANDY09 | Andy Smith slough | Taku River/Stikine Mainstem | 54 |
| SPORCU07.SPORCU11 | Porcupine | Taku River/Stikine Mainstem | 74 |
| SDEVIL07.SDEVIL08 | Devil's Elbow0708 | Taku River/Stikine Mainstem | 148 |

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| ADF\&G collection code | Location | Reporting group | $n$ |
| :---: | :---: | :---: | :---: |
| SDEVIL09 | Devil's Elbow09 | Taku River/Stikine Mainstem | 53 |
| SSCUD07.SSCUD08.SSCUD09 | Scud River | Taku River/Stikine Mainstem | 192 |
| SISKU85.SISKU86.SISKU02.SISKU06. |  |  |  |
| SISKU08.SISKU09 | Iskut River | Taku River/Stikine Mainstem | 153 |
| SISKU07 | Iskut River (Craigson Slough) | Taku River/Stikine Mainstem | 42 |
| SCRAIG06.SCRAIG07.SCRAIG08 | Craig River-CAN | Taku River/Stikine Mainstem | 38 |
| SBRON08.SBRON09 | Bronson Slough | Taku River/Stikine Mainstem | 78 |
| SSHAKS06.SSHAKES07.SSHAKS09 | Shakes Slough | Taku River/Stikine Mainstem | 67 |
| SCHRI11.SCHRI12 | Christina Lake | Taku River/Stikine Mainstem | 70 |
| SCRES03 | Crescent Lake | Snettisham | 194 |
| SSPEE03 | Speel Lake | Snettisham | 95 |
| SSNET06.SSPEE07 | Snettisham Hatchery0607 | Snettisham | 190 |
| SSPEE13 | Snettisham Hatchery 13 | Snettisham | 146 |
| SVIVID93 | Vivid Lake | Other | 48 |
| SSECLK14.SSECLKIN14 | Seclusion Lake | Other | 117 |
| SNBERG91 | North Berg Bay Inlet91 | Other | 53 |
| SNBERG92 | North Berg Bay Inlet92 | Other | 100 |
| SBART13 | Bartlett River | Other | 69 |
| SNEVA08 | Neva Lake08 | Other | 94 |
| SNEVA09.SNEVA13 | Neva Lake0913 | Other | 255 |
| SHOKTAI04 | Hoktaheen - main inlet | Other | 47 |
| SHOKTAO04 | Hoktaheen - outlet | Other | 49 |
| SHOKTAM14 | Hoktaheen - marine waters | Other | 47 |
| SKLAG09 | Klag Bay Stream | Other | 200 |
| SFORD04 | Ford Arm Lake | Other | 207 |
| SFORD13 | Ford Arm Creek | Other | 199 |
| SREDOUBT13 | Redoubt Lake | Other | 200 |
| SSALML07.SSALML08 | Salmon Lake | Other | 185 |
| SNECKER91.SNECKER93 | Benzeman Lake | Other | 95 |
| SFALL03.SFALL10 | Falls Lake | Other | 190 |
| SREDB93 | Redfish Lake | Other | 94 |
| SKUTL03 | Kutlaku03 | Other | 95 |
| SKUTL12 | Kutlaku12 | Other | 78 |
| SKUTL13 | Kutlaku13 | Other | 50 |
| SPAVLOF12.SPAVLOFR13 | Pavlof River | Other | 174 |
| SKOOK07.SKOOK10L.SKOOK12L | Kook Lake Late | Other | 194 |
| SKOOK12E.SKOOK13 | Kook Lake early | Other | 148 |

-continued-

Appendix C.--Page 3 of 7.

| ADF\&G collection code | Location | Reporting group | $n$ |
| :---: | :---: | :---: | :---: |
| SSITK03.SSITK11. |  |  |  |
| SSITK12 | Sitkoh Lake | Other | 351 |
| SLEVA12 | Lake Eva | Other | 115 |
| SHASSEL12.SHASSELR13 | Hasselborg Lake | Other | 209 |
| SKANA07.SKANA10.SKANAL13 | Kanalku Lake | Other | 319 |
| SBAIN10 | Bainbridge Lake | Other | 95 |
| SCOGH91.SCOG92HL.SCOG92ES. |  |  |  |
| SCOGH10 | Coghill Lake | Other | 378 |
| SESHAR08.SESHA91 | Eshamy Creek | Other | 185 |
| SMAIN91 | Main Bay | Other | 96 |
| SMINE91.SMINE09 | Miners Lake | Other | 191 |
| SEYAM07 | Eyak Lake - Middle Arm | Other | 95 |
| SEYASB07 | Eyak Lake - South beaches | Other | 87 |
| SEYAK10 | Eyak Lake - Hatchery Creek | Other | 95 |
| SMEND08.SMEND09 | Mendeltna Creek | Other | 188 |
| SSWEDE08 | Swede Lake | Other | 95 |
| SFISHC08 | East Fork Gulkana River | Other | 95 |
| SGULK08EF | Gulkana River - East Fork | Other | 75 |
| SPAXSO09 | Paxson Lake | Other | 75 |
| SMENT08 | Mentasta Lake | Other | 95 |
| STANA05 | Tanada Creek | Other | 94 |
| STANAO09 | Tanada Lake - lower outlet | Other | 95 |
| STANAS09 | Tanada Lake - shore | Other | 93 |
| SKLUT08 | Klutina River | Other | 95 |
| SKLUTI08.SKLUTI09 | Klutina Lake | Other | 95 |
| SBEARH08 | Bear Hole - Klutina | Other | 94 |
| SBANA08 | Banana Lake - Klutina | Other | 80 |
| SSANN05.SSTACR08 | St. Anne Creek | Other | 186 |
| SMAHL08 | Mahlo River | Other | 94 |
| STONSL09 | Tonsina Lake | Other | 94 |
| SLONGLK05 | Long Lake | Other | 95 |
| STEBA08 | Tebay River | Other | 93 |
| SSTEAM08 | Steamboat Lake - Bremner | Other | 95 |
| SSALMC08 | Salmon Creek - Bremner | Other | 93 |
| SCLEAR07 | Clear Creek | Other | 87 |
| SMCKI07 | McKinley Lake07 | Other | 95 |
| SMCKI08 | McKinley Lake08 | Other | 95 |
| SMCKI91 | McKinley Lake91 | Other | 95 |
| SMCKSC07 | McKinley Lake - Salmon Creek | Other | 93 |
| SMART07.SMART08 | Martin Lake | Other | 187 |

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| ADF\&G collection code | Location | Reporting group | $n$ |
| :---: | :---: | :---: | :---: |
| SMARTR08 | Martin River Slough | Other | 95 |
| STOKUN08.STOKUN09 | Tokun Lake | Other | 189 |
| SBERI91 | Bering Lake | Other | 95 |
| SKUSH07.SKUSH08 | Kushtaka Lake | Other | 189 |
| SSITU07 | Mountain Stream | Other | 159 |
| SSITU13 | Situk Lake | Other | 190 |
| SOSITU07 | Old Situk River | Other | 163 |
| SLOST03B | Lost/Tahwah Rivers | Other | 93 |
| SAHRN07 | Ahrnklin River | Other | 90 |
| SDANG09 | Dangerous River | Other | 95 |
| SAKWE09 | Akwe River | Other | 95 |
| SEAST03B | East Alsek River | Other | 94 |
| SDATLAS12 | Datlasaka Creek | Other | 95 |
| SGOATC07.SGOATC12 | Goat Creek | Other | 56 |
| SBORD07.SBORD08 | Border Slough0708 | Other | 71 |
| SBORD09.SBORD11 | Border Slough0911 | Other | 70 |
| STWEED07 | Tweedsmuir07 | Other | 48 |
| STWEED09 | Tweedsmuir09 | Other | 46 |
| SVERNR09.SVERNR10 | Vern Ritchie | Other | 114 |
| SNESK07 | Neskataheen Lake | Other | 195 |
| SKLUK06 | Klukshu River06 | Other | 95 |
| SKLUK07 | Klukshu River07 | Other | 94 |
| SKUDW09.SKUDW10.SKUDW11 | Kudwat Creek | Other | 100 |
| SBRIDGE11.SBRIDGE12 | Tatshenshini - Bridge/Silver | Other | 105 |
| SSTINKY11 | Tatshenshini - Stinky Creek | Other | 40 |
| SUTATS03 | Upper Tatshenshini | Other | 95 |
| SLTATS01.SLTATS03 | Little Tatshenshini Lake | Other | 65 |
| SKWAT11 | Kwatini River | Other | 65 |
| SBLAN07 | Blanchard River07 | Other | 89 |
| SBLAN09 | Blanchard River09 | Other | 62 |
| SLTAH90 | Tahltan Lake90 | Other | 95 |
| STAHL06 | Tahltan Lake06 | Other | 196 |
| SPETL04 | Petersburg Lake | Other | 95 |
| SKAHS03 | Kah Sheets Lake | Other | 96 |
| SMILLC07E | Mill Creek Weir Early | Other | 94 |
| SMILLC07L | Mill Creek Weir Late | Other | 95 |
| SKUNK03 | Kunk Lake | Other | 96 |
| STHOM04.STHOM14 | Thoms Lake | Other | 93 |
| SREDBL04 | Red Bay Lake | Other | 95 |

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| ADF\&G collection code | Location | Reporting group | $n$ |
| :---: | :---: | :---: | :---: |
| SSALM04.SSALM07 | Salmon Bay Lake | Other | 170 |
| SSHIP03 | Shipley Lake | Other | 94 |
| SSARK00.SSARF05 | Sarkar Lakes | Other | 91 |
| SHATC03.SHATC07 | Hatchery Creek | Other | 142 |
| SLUCK04 | Luck Lake | Other | 94 |
| SBIGLK10.SBIGLA14 | Big Lake | Other | 161 |
| SMCDO01.SMCDO03.SMCDO07.SMCDO13 | McDonald Lake | Other | 369 |
| SKART92.SMCGI03.SMCGI04.SMCGI16 | Karta River | Other | 472 |
| SGENE07 | Unuk River07 | Other | 95 |
| SGENE08 | Unuk River08 | Other | 69 |
| SHELM05 | Helm Lake | Other | 94 |
| SHECK04.SHECK07 | Heckman Lake | Other | 189 |
| SMAHO03.SMAHO07 | Mahoney Creek | Other | 154 |
| SKEGA04 | Kegan Lake | Other | 95 |
| SFILLM05 | Fillmore Lake | Other | 52 |
| STHRE04.STHRE10 | Klawock - Three Mile | Other | 181 |
| SINCK03.SINCK08.SHALF08 | Klawock - Inlet Creek | Other | 212 |
| SHETT03.SHETT08.SHETT09L | Hetta Lake | Other | 281 |
| SHETT09M | Hetta Creek - middle run | Other | 95 |
| SHETT10E | Hetta Creek - early run | Other | 95 |
| SEEK04.SEEK07 | Eek Creek | Other | 50 |
| SKLAK04 | Klakas Lake | Other | 95 |
| SBAR04 | Essowah Lake | Other | 95 |
| SHSMI92.SHUGH13 | Hugh Smith | Other | 155 |
| SHUGH04 | HS - Buschmann | Other | 151 |
| SCOBB07 | HS - Cobb Creek | Other | 99 |
| SKWIN01.SKWIN12U | Kwinageese | Other | 76 |
| SBOWS01 | Bowser Lake | Other | 94 |
| SBONN01.SBONN12 | Bonney Creek | Other | 164 |
| SDAMD01 | Damdochax Creek | Other | 93 |
| SMERI01.SMEZIB06 | Meziadin Lake | Other | 186 |
| SHANNA06 | Hanna Creek | Other | 93 |
| STINT06 | Tintina Creek | Other | 94 |
| SGING97 | Gingit Creek | Other | 94 |
| SALAS87.SALAS06 | Alastair Lake | Other | 118 |
| SLAKEL06 | Lakelelse Lake | Other | 93 |
| SSUST01 | Sustut River | Other | 79 |
| SSALIX87.SSALIX88 | Salix Bear | Other | 94 |

Appendix C.-Page 6 of 7.

| ADF\&G collection code | Location | Reporting group | $n$ |
| :---: | :---: | :---: | :---: |
| SMOTA87 | Motase Lake | Other | 47 |
| SSLAM06 | Slamgeesh River | Other | 95 |
| SUBAB06 | Babine River | Other | 95 |
| SFMILE06 | Four Mile Creek | Other | 85 |
| SPINK94.SPINK06 | Pinkut Creek | Other | 187 |
| SGRIZ87 | Grizzly Creek | Other | 76 |
| SPIER06 | Pierre Creek | Other | 95 |
| SFULT06 | Fulton River | Other | 95 |
| SMORR07 | Morrison | Other | 92 |
| SLTAH94 | Lower Tahlo River | Other | 78 |
| STAHLO07 | Tahlo Creek | Other | 95 |
| SMCDON02.SMCDON06 | McDonell Lake (Zymoetz River) | Other | 131 |
| SKALUM06 | Kitsumkalum Lake06 | Other | 56 |
| SKALUM12 | Kitsumkalum Lake12 | Other | 94 |
| SKITW12 | Kitwanga River | Other | 92 |
| SSTECR01 | Stephens Creek | Other | 95 |
| SNANG06 | Nangeese River | Other | 40 |
| SKISP02 | Kispiox River | Other | 53 |
| SSWANLK06 | Swan Lake | Other | 93 |
| SNANI88.SNANI07 | Nanika River | Other | 114 |
| SKYNO97 | Trembleur - Kynock | Other | 94 |
| STACH01 | Tachie River | Other | 94 |
| SSTEL07 | Stellako River | Other | 94 |
| SFRAS96 | Fraser Lake | Other | 85 |
| SMITCH01 | Mitchell River | Other | 94 |
| SLHOR01.SUHOR01.SHORSE07 | Horsefly River | Other | 274 |
| SNAHAT02 | Nahatlatch River | Other | 92 |
| SCULT02 | Cultus Lake | Other | 91 |
| SCHILW04 | Chilliwack Lake | Other | 90 |
| SCHILK01 | Chilko Lake | Other | 87 |
| SRAFT01 | Raft River | Other | 84 |
| SLADA02.SADAM07 | Adams River | Other | 187 |
| SMSHU02 | Middle Shuswap River | Other | 91 |
| SSCOT00 | Scotch River | Other | 91 |
| SGATES09 | Gates Creek | Other | 90 |
| SBIRK07 | Birkenhead River | Other | 90 |
| SWEAV01 | Weaver Creek | Other | 89 |
| SHARR07 | Harrison River | Other | 95 |
| SNTHOM05 | North Thompson | Other | 95 |

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Appendix C.-Page 7 of 7.

| ADF\&G collection code | Location | Reporting group | $n$ |
| :--- | :--- | :--- | :--- |
| SNADE95 | Naden River | Other | 95 |
| SYAKO93 | QCI - Yakoun Lake | Other | 70 |
| SKITIM10 | Kitimat River | Other | 93 |
| SBLOOM05 | Bloomfield Lake | Other | 94 |
| STANK03 | Tankeeah River03 | Other | 47 |
| STANK05 | Tankeeah River05 | Other | 47 |
| SAMBA04 | Central Coast - Amback Creek | Other | 91 |
| SKITL06 | Kitlope Lake | Other | 95 |
| SGCENLK02 | Great Central Lake | Other | 95 |
| SQUAT03 | Vancouver Island - Quatse River | Other | 95 |
| SOKAN02 | Okanagan River | Other | 95 |
| SLAKE97 | Lake Pleasant | Other | Other |
| SISSA96 | Issaquah Creek | Other | 89 |
| SWENA98 | Lake Wenatchee |  | 82 |

Appendix D.-Chilkoot River weir dates of operation, annual estimates of sockeye salmon escapement, and counts of other species, 1976-2019. (Numbers in bold are historical records that have been updated since the last report by Bednarski et al. [2016]).

| Year | Date in | Date out | Sockeye | Pink | Chum | Coho | Chinook |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 29-May | 4-Nov | 71,296 | 1,250 | 241 | 991 | ND |
| 1977 | 28-May | 18-Sep | 97,368 | 5,270 | 195 | 5 | ND |
| 1978 | 6-Jun | 8 -Nov | 35,454 | 112 | 382 | 1,092 | ND |
| 1979 | $9-\mathrm{Jun}$ | 4-Nov | 95,948 | NA | 253 | 899 | ND |
| 1980 | 15-Jun | 4-Oct | 96,513 | 4,683 | 719 | 628 | ND |
| 1981 | 10-Jun | 12-Oct | 84,047 | 34,821 | 405 | 1,585 | ND |
| 1982 | 3-Jun | 14-Sep | 103,038 | 6,665 | 507 | 5 | 6 |
| 1983 | 4-Jun | 12-Nov | 80,141 | 11,237 | 501 | 1,844 | 0 |
| 1984 | 3-Jun | 14-Sep | 100,781 | 5,034 | 372 | 321 | 0 |
| 1985 | 5-Jun | 28-Oct | 69,141 | 33,608 | 1,031 | 2,202 | 5 |
| 1986 | 4-Jun | $28-$ Oct | 88,024 | 1,249 | 508 | 1,966 | 6 |
| 1987 | 4-Jun | $2-\mathrm{Nov}$ | 94,208 | 6,689 | 431 | 576 | 3 |
| 1988 | 9 -Jun | 12-Nov | 81,274 | 5,274 | 450 | 1,476 | 1 |
| 1989 | 3-Jun | 30-Oct | 54,900 | 2,118 | 223 | 3,998 | 0 |
| 1990 | 3-Jun | 30-Oct | 76,119 | 10,398 | 216 | 988 | 0 |
| 1991 | 7-Jun | 8-Oct | 90,754 | 2,588 | 357 | 4,000 | 0 |
| 1992 | 2-Jun | 26-Sep | 67,071 | 7,836 | 193 | 1,518 | 1 |
| 1993 | 3-Jun | 30-Sep | 52,080 | 357 | 240 | 322 | 203 |
| 1994 | 4-Jun | 24-Sep | 37,007 | 22,472 | 214 | 463 | 118 |
| 1995 | 5-Jun | 10-Sep | 7,177 | 1,243 | 99 | 95 | 7 |
| 1996 | 6-Jun | 11-Sep | 50,741 | 2,867 | 305 | 86 | 19 |
| 1997 | 4-Jun | 9-Sep | 44,254 | 26,197 | 268 | 17 | 6 |
| 1998 | 4-Jun | 13-Sep | 12,335 | 44,001 | 368 | 131 | 11 |
| 1999 | 2-Jun | 13-Sep | 19,284 | 56,692 | 713 | 11 | 29 |
| 2000 | 3-Jun | 12-Sep | 43,555 | 23,636 | 1050 | 47 | 10 |
| 2001 | 7-Jun | 12-Sep | 76,283 | 32,294 | 810 | 103 | 24 |
| 2002 | 8-Jun | 11-Sep | 58,361 | 79,639 | 352 | 304 | 36 |
| 2003 | 5-Jun | 9-Sep | 74,459 | 55,424 | 498 | 15 | 12 |
| 2004 | 3-Jun | 12-Sep | 75,596 | 107,994 | 617 | 89 | 17 |
| 2005 | 5-Jun | 12-Sep | 51,178 | 90,486 | 262 | 23 | 9 |
| 2006 | 4-Jun | 13-Sep | 96,203 | 33,888 | 257 | 158 | 1 |
| 2007 | 4-Jun | 12-Sep | 72,678 | 61,469 | 252 | 13 | 39 |
| 2008 | 4-Jun | 12-Sep | 33,117 | 15,105 | 327 | 50 | 31 |
| 2009 | 5-Jun | 10-Sep | 33,705 | 34,483 | 171 | 11 | 12 |
| 2010 | 6-Jun | 14-Sep | 71,657 | 30,830 | 410 | 90 | 6 |
| 2011 | 3-Jun | 6-Sep | 65,915 | 76,244 | 118 | 18 | 43 |
| 2012 | 1-Jun | 12-Sep | 118,166 | 40,753 | 494 | 139 | 47 |
| 2013 | 1-Jun | 7-Sep | 46,329 | 8,195 | 566 | 43 | 139 |
| 2014 | 27-May | 9-Sep | 105,713 | 12,457 | 126 | 162 | 83 |
| 2015 | 2-Jun | 8-Sep | 71,515 | 41,592 | 185 | 11 | 22 |
| 2016 | 3-Jun | 9-Sep | 86,721 | 8,354 | 116 | 53 | 2 |
| 2017 | 2-Jun | 6-Sep | 43,098 | 58,664 | 529 | 12 | 11 |
| 2018 | 3-Jun | 8-Sep | 85,463 | 5,475 | 225 | 95 | 31 |
| 2019 | 7-Jun | 8-Sep | 140,378 | 17,156 | 396 | 80 | 64 |
| Average ${ }^{\text {a }}$ | 1-Jun | 27-Sep | 67,448 | 26,594 | 392 | 662 | 28 |

[^6]Appendix E.-Daily and cumulative (cum.) Chilkoot River weir salmon counts by species, and water temperature and gauge heights, 2016.

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | Water level (cm) | Water temp. $\left({ }^{\circ} \mathrm{C}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 3-Jun | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 146 | 8.5 |
| 4-Jun | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 149 | 5.5 |
| 5-Jun | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 152 | 8.0 |
| 6-Jun | 2 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 154 | 8.0 |
| 7-Jun | 26 | 29 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 156 | 8.0 |
| 8-Jun | 8 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 150 | 8.0 |
| 9-Jun | 6 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 146 | 8.5 |
| 10-Jun | 16 | 59 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 153 | 8.5 |
| 11-Jun | 14 | 73 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 155 | 9.0 |
| 12-Jun | 73 | 146 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 153 | 8.5 |
| 13-Jun | 61 | 207 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 153 | 8.0 |
| 14-Jun | 59 | 266 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 152 | 9.0 |
| 15-Jun | 177 | 443 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 149 | 11.0 |
| 16-Jun | 186 | 629 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 156 | 10.5 |
| 17-Jun | 261 | 890 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 153 | 10.5 |
| 18-Jun | 250 | 1,140 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 156 | 10.5 |
| 19-Jun | 176 | 1,316 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 155 | 10.0 |
| 20-Jun | 123 | 1,439 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 152 | 10.0 |
| 21-Jun | 64 | 1,503 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 146 | 10.5 |
| 22-Jun | 24 | 1,527 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 144 | 11.0 |
| 23-Jun | 187 | 1,714 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 149 | 10.5 |
| 24-Jun | 190 | 1,904 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 153 | 10.5 |
| 25-Jun | 191 | 2,095 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 150 | 10.0 |
| 26-Jun | 450 | 2,545 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 153 | 10.0 |
| 27-Jun | 1,022 | 3,567 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 160 | 10.5 |
| 28-Jun | 212 | 3,779 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 154 | 10.0 |
| 29-Jun | 321 | 4,100 | 0 | 1 | 1 | 2 | 0 | 0 | 0 | 0 | 150 | 10.0 |
| 30-Jun | 724 | 4,824 | 0 | 1 | 2 | 4 | 0 | 0 | 0 | 0 | 150 | 10.0 |
| 1-Jul | 1,455 | 6,279 | 0 | 1 | 0 | 4 | 0 | 0 | 0 | 0 | 150 | 10.0 |
| 2-Jul | 1,226 | 7,505 | 0 | 1 | 1 | 5 | 0 | 0 | 0 | 0 | 148 | 12.0 |
| 3-Jul | 849 | 8,354 | 0 | 1 | 3 | 8 | 0 | 0 | 0 | 0 | 148 | 11.0 |
| 4-Jul | 841 | 9,195 | 0 | 1 | 0 | 8 | 0 | 0 | 0 | 0 | 142 | 10.0 |
| 5-Jul | 1,150 | 10,345 | 0 | 1 | 1 | 9 | 0 | 0 | 0 | 0 | 140 | 11.0 |
| 6-Jul | 601 | 10,946 | 0 | 1 | 3 | 12 | 0 | 0 | 0 | 0 | 138 | 11.5 |

-continued-

Appendix E.-Page 2 of 3.

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | Water level (cm) | Water temp. $\left({ }^{\circ} \mathrm{C}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 7-Jul | 342 | 11,288 | 1 | 2 | 0 | 12 | 0 | 0 | 0 | 0 | 137 | 10.5 |
| 8-Jul | 1,178 | 12,466 | 2 | 4 | 2 | 14 | 0 | 0 | 0 | 0 | 138 | 11.0 |
| 9-Jul | 902 | 13,368 | 2 | 6 | 3 | 17 | 0 | 0 | 0 | 0 | 144 | 12.0 |
| 10-Jul | 544 | 13,912 | 3 | 9 | 0 | 17 | 0 | 0 | 0 | 0 | 149 | 12.0 |
| 11-Jul | 671 | 14,583 | 0 | 9 | 2 | 19 | 0 | 0 | 0 | 0 | 146 | 11.5 |
| 12-Jul | 1,518 | 16,101 | 0 | 9 | 0 | 19 | 0 | 0 | 0 | 0 | 147 | 11.5 |
| 13-Jul | 910 | 17,011 | 0 | 9 | 0 | 19 | 0 | 0 | 0 | 0 | 152 | 11.5 |
| 14-Jul | 1,215 | 18,226 | 0 | 9 | 3 | 22 | 0 | 0 | 0 | 0 | 155 | 11.5 |
| 15-Jul | 1,218 | 19,444 | 10 | 19 | 5 | 27 | 0 | 0 | 0 | 0 | 154 | 12.0 |
| 16-Jul | 1,091 | 20,535 | 23 | 42 | 1 | 28 | 0 | 0 | 0 | 0 | 154 | 11.5 |
| 17-Jul | 524 | 21,059 | 3 | 45 | 0 | 28 | 0 | 0 | 0 | 0 | 155 | 13.0 |
| 18-Jul | 769 | 21,828 | 3 | 48 | 3 | 31 | 0 | 0 | 0 | 0 | 155 | 12.5 |
| 19-Jul | 959 | 22,787 | 5 | 53 | 3 | 34 | 0 | 0 | 0 | 0 | 157 | 13.5 |
| 20-Jul | 1,813 | 24,600 | 13 | 66 | 1 | 35 | 0 | 0 | 1 | 1 | 152 | 12.0 |
| 21-Jul | 1,420 | 26,020 | 34 | 100 | 4 | 39 | 0 | 0 | 0 | 1 | 147 | 12.0 |
| 22-Jul | 2,003 | 28,023 | 11 | 111 | 1 | 40 | 0 | 0 | 0 | 1 | 144 | 11.5 |
| 23-Jul | 941 | 28,964 | 4 | 115 | 2 | 42 | 0 | 0 | 0 | 1 | 152 | 11.0 |
| 24-Jul | 448 | 29,412 | 0 | 115 | 1 | 43 | 0 | 0 | 0 | 1 | 169 | 9.5 |
| 25-Jul | 124 | 29,536 | 0 | 115 | 3 | 46 | 0 | 0 | 0 | 1 | 162 | 10.5 |
| 26-Jul | 287 | 29,823 | 0 | 115 | 3 | 49 | 0 | 0 | 0 | 1 | 157 | 11.0 |
| 27-Jul | 802 | 30,625 | 3 | 118 | 3 | 52 | 0 | 0 | 0 | 1 | 154 | 11.0 |
| $28-\mathrm{Jul}^{\text {a }}$ | 765 | 31,390 | 0 | 118 | 0 | 52 | 0 | 0 | 0 | 1 | 168 | 11.0 |
| $29-\mathrm{Jul}^{\text {a }}$ | 984 | 32,374 | 0 | 118 | 1 | 53 | 0 | 0 | 0 | 1 | 172 | 11.0 |
| 30-Jul | 1,142 | 33,516 | 16 | 134 | 3 | 56 | 0 | 0 | 0 | 1 | 153 | 11.0 |
| 31-Jul | 1,265 | 34,781 | 14 | 148 | 9 | 65 | 0 | 0 | 0 | 1 | 144 | 11.0 |
| 1-Aug | 3,057 | 37,838 | 71 | 219 | 2 | 67 | 0 | 0 | 1 | 2 | 138 | 11.0 |
| 2-Aug | 5,058 | 42,896 | 67 | 286 | 2 | 69 | 0 | 0 | 0 | 2 | 139 | 11.0 |
| 3-Aug | 4,442 | 47,338 | 74 | 360 | 0 | 69 | 0 | 0 | 0 | 2 | 136 | 11.5 |
| 4-Aug | 13,701 | 61,039 | 232 | 592 | 0 | 69 | 0 | 0 | 0 | 2 | 138 | 11.5 |
| 5-Aug | 8,488 | 69,527 | 1,131 | 1,723 | 0 | 69 | 0 | 0 | 0 | 2 | 137 | 11.0 |
| 6-Aug | 4,043 | 73,570 | 919 | 2,642 | 0 | 69 | 0 | 0 | 0 | 2 | 137 | 11.0 |
| 7-Aug | 4,354 | 77,924 | 786 | 3,428 | 0 | 69 | 0 | 0 | 0 | 2 | 140 | 11.0 |
| 8-Aug | 1,607 | 79,531 | 958 | 4,386 | 0 | 69 | 1 | 1 | 0 | 2 | 139 | 11.5 |
| 9-Aug | 343 | 79,874 | 515 | 4,901 | 0 | 69 | 0 | 1 | 0 | 2 | 147 | 11.0 |

[^7]Appendix E.-Page 3 of 3 .

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | Water level (cm) | Water temp. $\left({ }^{\circ} \mathrm{C}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 10-Aug | 468 | 80,342 | 269 | 5,170 | 1 | 70 | 0 | 1 | 0 | 2 | 144 | 10.0 |
| 11-Aug | 282 | 80,624 | 272 | 5,442 | 0 | 70 | 0 | 1 | 0 | 2 | 142 | 11.0 |
| 12-Aug | 338 | 80,962 | 411 | 5,853 | 0 | 70 | 0 | 1 | 0 | 2 | 143 | 11.0 |
| 13-Aug | 331 | 81,293 | 133 | 5,986 | 0 | 70 | 0 | 1 | 0 | 2 | 146 | 11.0 |
| 14-Aug | 392 | 81,685 | 24 | 6,010 | 0 | 70 | 0 | 1 | 0 | 2 | 158 | 11.0 |
| 15-Aug | 175 | 81,860 | 4 | 6,014 | 0 | 70 | 0 | 1 | 0 | 2 | 174 | 11.0 |
| 16-Aug | 402 | 82,262 | 33 | 6,047 | 0 | 70 | 0 | 1 | 0 | 2 | 158 | 11.0 |
| 17-Aug | 198 | 82,460 | 36 | 6,083 | 1 | 71 | 0 | 1 | 0 | 2 | 148 | 11.0 |
| 18-Aug | 112 | 82,572 | 106 | 6,189 | 0 | 71 | 0 | 1 | 0 | 2 | 140 | 10.5 |
| 19-Aug | 209 | 82,781 | 178 | 6,367 | 0 | 71 | 0 | 1 | 0 | 2 | 145 | 10.5 |
| 20-Aug | 328 | 83,109 | 348 | 6,715 | 0 | 71 | 0 | 1 | 0 | 2 | 142 | 10.5 |
| 21-Aug | 319 | 83,428 | 401 | 7,116 | 0 | 71 | 0 | 1 | 0 | 2 | 136 | 10.5 |
| 22-Aug | 447 | 83,875 | 258 | 7,374 | 1 | 72 | 0 | 1 | 0 | 2 | 135 | 10.5 |
| 23-Aug | 447 | 84,322 | 163 | 7,537 | 0 | 72 | 0 | 1 | 0 | 2 | 132 | 10.5 |
| 24-Aug | 124 | 84,446 | 99 | 7,636 | 0 | 72 | 0 | 1 | 0 | 2 | 135 | 10.5 |
| 25-Aug | 102 | 84,548 | 63 | 7,699 | 0 | 72 | 0 | 1 | 0 | 2 | 137 | 11.5 |
| 26-Aug | 112 | 84,660 | 106 | 7,805 | 2 | 74 | 1 | 2 | 0 | 2 | 142 | 11.5 |
| 27-Aug | 82 | 84,742 | 35 | 7,840 | 0 | 74 | 1 | 3 | 0 | 2 | 146 | 12.0 |
| 28-Aug | 304 | 85,046 | 141 | 7,981 | 2 | 76 | 0 | 3 | 0 | 2 | 140 | 12.5 |
| 29-Aug | 272 | 85,318 | 94 | 8,075 | 9 | 85 | 6 | 9 | 0 | 2 | 136 | 12.0 |
| 30-Aug | 360 | 85,678 | 93 | 8,168 | 3 | 88 | 5 | 14 | 0 | 2 | 132 | 11.0 |
| 31-Aug | 59 | 85,737 | 19 | 8,187 | 2 | 90 | 3 | 17 | 0 | 2 | 129 | 11.5 |
| 1-Sep | 103 | 85,840 | 16 | 8,203 | 1 | 91 | 1 | 18 | 0 | 2 | 124 | 12.0 |
| 2-Sep | 147 | 85,987 | 40 | 8,243 | 9 | 100 | 1 | 19 | 0 | 2 | 123 | 11.5 |
| 3-Sep | 210 | 86,197 | 25 | 8,268 | 3 | 103 | 3 | 22 | 0 | 2 | 125 | 12.0 |
| 4-Sep | 143 | 86,340 | 26 | 8,294 | 4 | 107 | 13 | 35 | 0 | 2 | 123 | 12.0 |
| 5-Sep | 206 | 86,546 | 21 | 8,315 | 4 | 111 | 8 | 43 | 0 | 2 | 122 | 11.5 |
| 6-Sep | 23 | 86,569 | 5 | 8,320 | 4 | 115 | 3 | 46 | 0 | 2 | 122 | 11.5 |
| 7-Sep | 94 | 86,663 | 22 | 8,342 | 1 | 116 | 6 | 52 | 0 | 2 | 124 | 10.0 |
| 8-Sep | 55 | 86,718 | 12 | 8,354 | 0 | 116 | 1 | 53 | 0 | 2 | 124 | 10.5 |
| 9-Sep | 3 | 86,721 | 0 | 8,354 | 0 | 116 | 0 | 53 | 0 | 2 | ND | ND |

[^8]Appendix F.-Daily and cumulative (cum.) Chilkoot River weir salmon counts species, and water temperature and gauge heights, 2017.

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | $\begin{gathered} \text { Water } \\ \text { level }(\mathrm{cm}) \\ \hline \end{gathered}$ | Water temp. $\left({ }^{\circ} \mathrm{C}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 1-Jun | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 150 | 9.0 |
| 2-Jun | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 152 | 8.0 |
| 3-Jun | 8 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 152 | 9.0 |
| 4-Jun | 14 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 154 | 8.0 |
| 5-Jun | 16 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 148 | 8.0 |
| 6-Jun | 4 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 142 | 9.0 |
| 7-Jun | 6 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 146 | 9.0 |
| 8-Jun | 45 | 94 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 156 | 10.0 |
| 9-Jun ${ }^{\text {a }}$ | 23 | 117 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 180 | 10.0 |
| 10-Jun ${ }^{\text {a }}$ | 20 | 137 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 177 | 9.0 |
| 11-Jun ${ }^{\text {a }}$ | 17 | 154 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 158 | 8.0 |
| 12-Jun | 13 | 167 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 154 | 7.5 |
| 13-Jun | 14 | 181 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 147 | 7.0 |
| 14-Jun | 17 | 198 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 142 | 7.5 |
| 15-Jun | 9 | 207 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 137 | 7.5 |
| 16-Jun | 2 | 209 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 139 | 8.0 |
| 17-Jun | 10 | 219 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 148 | 7.5 |
| 18-Jun | 8 | 227 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 144 | 8.0 |
| 19-Jun | 6 | 233 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 141 | 9.0 |
| 20-Jun | 179 | 412 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 141 | 8.5 |
| 21-Jun | 11 | 423 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 142 | 8.0 |
| 22-Jun | 36 | 459 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 139 | 9.0 |
| 23-Jun | 35 | 494 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 136 | 7.5 |
| 24-Jun | 116 | 610 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 137 | 8.0 |
| 25-Jun | 46 | 656 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 138 | 9.0 |
| 26-Jun | 462 | 1,118 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 140 | 9.0 |
| 27-Jun | 117 | 1,235 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 141 | 9.0 |
| 28-Jun | 48 | 1,283 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 140 | 9.0 |
| 29-Jun | 84 | 1,367 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 139 | 8.5 |
| 30-Jun | 212 | 1,579 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 2 | 139 | 8.0 |
| 1-Jul | 122 | 1,701 | 5 | 6 | 0 | 0 | 0 | 0 | 0 | 2 | 145 | 8.0 |
| 2-Jul | 300 | 2,001 | 7 | 13 | 1 | 1 | 0 | 0 | 0 | 2 | 150 | 9.0 |
| 3-Jul | 271 | 2,272 | 0 | 13 | 0 | 1 | 0 | 0 | 0 | 2 | 154 | 8.5 |
| 4-Jul | 213 | 2,485 | 0 | 13 | 0 | 1 | 0 | 0 | 0 | 2 | 149 | 9.0 |

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Appendix F.-Page 2 of 3.

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | $\begin{gathered} \text { Water } \\ \text { level }(\mathrm{cm}) \end{gathered}$ | Water <br> temp. $\left({ }^{\circ} \mathrm{C}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 5-Jul | 318 | 2,803 | 9 | 22 | 0 | 1 | 0 | 0 | 0 | 2 | 146 | 9.0 |
| 6-Jul | 343 | 3,146 | 51 | 73 | 1 | 2 | 0 | 0 | 1 | 3 | 146 | 9.0 |
| 7-Jul | 352 | 3,498 | 63 | 136 | 0 | 2 | 0 | 0 | 0 | 3 | 151 | 10.0 |
| 8 -Jul | 606 | 4,104 | 15 | 151 | 2 | 4 | 0 | 0 | 1 | 4 | 153 | 9.5 |
| $9-\mathrm{Jul}$ | 285 | 4,389 | 21 | 172 | 1 | 5 | 0 | 0 | 0 | 4 | 152 | 9.0 |
| 10-Jul | 404 | 4,793 | 18 | 190 | 6 | 11 | 0 | 0 | 0 | 4 | 155 | 9.0 |
| 11-Jul | 555 | 5,348 | 17 | 207 | 5 | 16 | 0 | 0 | 0 | 4 | 151 | 9.5 |
| 12-Jul | 572 | 5,920 | 40 | 247 | 2 | 18 | 0 | 0 | 0 | 4 | 157 | 8.5 |
| 13-Jul | 533 | 6,453 | 52 | 299 | 4 | 22 | 0 | 0 | 0 | 4 | 150 | 8.0 |
| 14-Jul | 297 | 6,750 | 66 | 365 | 3 | 25 | 0 | 0 | 0 | 4 | 150 | 8.0 |
| 15-Jul | 71 | 6,821 | 39 | 404 | 1 | 26 | 0 | 0 | 0 | 4 | 148 | 7.5 |
| 16-Jul | 386 | 7,207 | 163 | 567 | 3 | 29 | 0 | 0 | 0 | 4 | 149 | 8.5 |
| 17-Jul | 576 | 7,783 | 136 | 703 | 2 | 31 | 0 | 0 | 1 | 5 | 152 | 8.5 |
| 18-Jul | 700 | 8,483 | 116 | 819 | 7 | 38 | 0 | 0 | 0 | 5 | 150 | 9.0 |
| 19-Jul | 920 | 9,403 | 246 | 1,065 | 6 | 44 | 0 | 0 | 0 | 5 | 148 | 9.0 |
| 20-Jul | 1,107 | 10,510 | 449 | 1,514 | 13 | 57 | 0 | 0 | 1 | 6 | 146 | 8.0 |
| 21-Jul | 1,348 | 11,858 | 559 | 2,073 | 16 | 73 | 0 | 0 | 0 | 6 | 142 | 8.5 |
| 22-Jul | 1,256 | 13,114 | 858 | 2,931 | 13 | 86 | 0 | 0 | 3 | 9 | 140 | 8.0 |
| 23-Jul | 1,016 | 14,130 | 2,095 | 5,026 | 6 | 92 | 0 | 0 | 0 | 9 | 138 | 9.0 |
| 24-Jul | 1,139 | 15,269 | 568 | 5,594 | 6 | 98 | 0 | 0 | 0 | 9 | 139 | 9.5 |
| 25-Jul | 371 | 15,640 | 144 | 5,738 | 4 | 102 | 0 | 0 | 0 | 9 | 140 | 9.5 |
| 26-Jul | 1,077 | 16,717 | 446 | 6,184 | 8 | 110 | 0 | 0 | 1 | 10 | 140 | 8.5 |
| 27-Jul | 1,270 | 17,987 | 613 | 6,797 | 7 | 117 | 0 | 0 | 0 | 10 | 139 | 9.0 |
| 28-Jul | 1,176 | 19,163 | 459 | 7,256 | 3 | 120 | 0 | 0 | 0 | 10 | 141 | 8.5 |
| 29-Jul | 490 | 19,653 | 141 | 7,397 | 2 | 122 | 0 | 0 | 0 | 10 | 138 | 9.0 |
| 30-Jul | 2,217 | 21,870 | 384 | 7,781 | 10 | 132 | 0 | 0 | 0 | 10 | 135 | 9.5 |
| 31-Jul | 1,370 | 23,240 | 595 | 8,376 | 6 | 138 | 0 | 0 | 0 | 10 | 134 | 9.5 |
| 1-Aug | $1,049$ | $24,289$ | $1,302$ | 9,678 | 1 | 139 | 0 | 0 | 0 | 10 | 134 | 10.5 |
| 2-Aug | 2,385 | 26,674 | 2,288 | 11,966 | 4 | 143 | 0 | 0 | 1 | 11 | 134 | 10.0 |
| 3-Aug | 1,245 | 27,919 | 1,781 | 13,747 | 9 | 152 | 0 | 0 | 0 | 11 | 135 | 9.5 |
| 4-Aug | 859 | 28,778 | 1,826 | 15,573 | 0 | 152 | 0 | 0 | 0 | 11 | 136 | 10.5 |
| 5-Aug | 1,044 | 29,822 | 1,943 | 17,516 | 8 | 160 | 0 | 0 | 0 | 11 | 141 | 11.0 |
| 6-Aug | 947 | 30,769 | 1,776 | 19,292 | 6 | 166 | 0 | 0 | 0 | 11 | 145 | 10.5 |

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Appendix F.-Page 3 of 3.

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | $\begin{gathered} \text { Water } \\ \text { level }(\mathrm{cm}) \\ \hline \end{gathered}$ | Water <br> temp. $\left({ }^{\circ} \mathrm{C}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 7-Aug | 1,144 | 31,913 | 5,919 | 25,211 | 5 | 171 | 0 | 0 | 0 | 11 | 148 | 11.0 |
| 8-Aug | 395 | 32,308 | 2,519 | 27,730 | 3 | 174 | 0 | 0 | 0 | 11 | 146 | 11.0 |
| 9-Aug | 217 | 32,525 | 1,356 | 29,086 | 5 | 179 | 0 | 0 | 0 | 11 | 149 | 11.0 |
| 10-Aug | 380 | 32,905 | 1,198 | 30,284 | 5 | 184 | 0 | 0 | 0 | 11 | 141 | 11.0 |
| 11-Aug | 243 | 33,148 | 1,750 | 32,034 | 9 | 193 | 0 | 0 | 0 | 11 | 142 | 11.5 |
| 12-Aug | 713 | 33,861 | 2,178 | 34,212 | 10 | 203 | 0 | 0 | 0 | 11 | 138 | 11.0 |
| 13-Aug | 305 | 34,166 | 410 | 34,622 | 9 | 212 | 0 | 0 | 0 | 11 | 138 | 11.0 |
| 14-Aug | 634 | 34,800 | 547 | 35,169 | 30 | 242 | 0 | 0 | 0 | 11 | 143 | 9.5 |
| 15-Aug | 48 | 34,848 | 90 | 35,259 | 11 | 253 | 0 | 0 | 0 | 11 | 138 | 10.5 |
| 16-Aug | 541 | 35,389 | 456 | 35,715 | 9 | 262 | 0 | 0 | 0 | 11 | 135 | 10.0 |
| 17-Aug | 628 | 36,017 | 1,118 | 36,833 | 20 | 282 | 0 | 0 | 0 | 11 | 135 | 9.0 |
| 18-Aug | 928 | 36,945 | 3,346 | 40,179 | 14 | 296 | 0 | 0 | 0 | 11 | 138 | 10.0 |
| 19-Aug | 769 | 37,714 | 2,247 | 42,426 | 25 | 321 | 0 | 0 | 0 | 11 | 134 | 9.5 |
| 20-Aug | 502 | 38,216 | 1,901 | 44,327 | 6 | 327 | 0 | 0 | 0 | 11 | 134 | 9.5 |
| 21-Aug | 607 | 38,823 | 2,681 | 47,008 | 10 | 337 | 0 | 0 | 0 | 11 | 139 | 10.0 |
|  | 233 | 39,056 | 1,841 | 48,849 | 4 | 341 | 0 | 0 | 0 | 11 | 149 | 10.0 |
| 23-Aug | 63 | 39,119 | 200 | 49,049 | 2 | 343 | 0 | 0 | 0 | 11 | 160 | 8.5 |
| 24-Aug | 139 | 39,258 | 709 | 49,758 | 4 | 347 | 0 | 0 | 0 | 11 | 152 | 8.5 |
| 25-Aug | 405 | 39,663 | 1,703 | 51,461 | 5 | 352 | 1 | 1 | 0 | 11 | 142 | 9.0 |
| 26-Aug | 46 | 39,709 | 61 | 51,522 | 1 | 353 | 0 | 1 | 0 | 11 | 157 | 9.5 |
| 27-Aug | 167 | 39,876 | 133 | 51,655 | 2 | 355 | 0 | 1 | 0 | 11 | 155 | 8.5 |
| 28-Aug | 677 | 40,553 | 681 | 52,336 | 5 | 360 | 0 | 1 | 0 | 11 | 145 | 9.0 |
| 29-Aug | 706 | 41,259 | 1,837 | 54,173 | 18 | 378 | 0 | 1 | 0 | 11 | 138 | 9.0 |
| 30-Aug | 176 | 41,435 | 1,263 | 55,436 | 28 | 406 | 0 | 1 | 0 | 11 | 139 | 9.5 |
| 31-Aug | 235 | 41,670 | 365 | 55,801 | 6 | 412 | 0 | 1 | 0 | 11 | 148 | 9.5 |
| 1-Sep | 223 | 41,893 | 606 | 56,407 | 4 | 416 | 1 | 2 | 0 | 11 | 144 | 9.0 |
| 2-Sep | 245 | 42,138 | 398 | 56,805 | 9 | 425 | 0 | 2 | 0 | 11 | 135 | 9.0 |
| 3-Sep | 448 | 42,586 | 1,078 | 57,883 | 30 | 455 | 2 | 4 | 0 | 11 | 132 | 9.0 |
| 4-Sep | 276 | 42,862 | 561 | 58,444 | 37 | 492 | 3 | 7 | 0 | 11 | 130 | 9.0 |
| 5-Sep | 101 | 42,963 | 142 | 58,586 | 14 | 506 | 3 | 10 | 0 | 11 | 144 | 9.5 |
| 6-Sep | 135 | 43,098 | 78 | 58,664 | 23 | 529 | 2 | 12 | 0 | 11 | 150 | 10.0 |

[^9]Appendix G.-Daily and cumulative (cum.) Chilkoot River weir salmon counts by species, and water temperature and gauge heights, 2018.

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | Water level (cm) | $\begin{gathered} \text { Water } \\ \text { temp }\left({ }^{\circ} \mathrm{C}\right) \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 3-Jun | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 127 | 7.5 |
| 4-Jun | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 128 | 8.0 |
| 5-Jun | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 132 | 7.0 |
| 6-Jun | 2 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 133 | 8.5 |
| 7-Jun | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 133 | 7.5 |
| 8-Jun | 1 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 132 | 7.5 |
| 9-Jun | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 128 | 8.5 |
| 10-Jun | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 131 | 8.0 |
| 11-Jun | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 130 | 7.5 |
| 12-Jun | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 129 | 7.0 |
| 13-Jun | 1 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 129 | 7.5 |
| 14-Jun | 23 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 133 | 8.5 |
| 15-Jun | 2 | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 142 | 8.0 |
| 16-Jun | 0 | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 137 | 8.0 |
| 17-Jun | 38 | 68 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 139 | 8.0 |
| 18-Jun | 19 | 87 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 144 | 8.0 |
| 19-Jun | 53 | 140 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 154 | 9.0 |
| 20-Jun | 104 | 244 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 161 | 9.0 |
| 21-Jun | 40 | 284 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 166 | 8.3 |
| 22-Jun | 27 | 311 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 164 | 9.5 |
| 23-Jun | 49 | 360 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 157 | 9.5 |
| 24-Jun | 252 | 612 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 146 | 9.0 |
| 25-Jun | 54 | 666 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 142 | 9.5 |
| 26-Jun | 58 | 724 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 138 | 8.5 |
| 27-Jun | 435 | 1,159 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 136 | 9.0 |
| 28-Jun | 319 | 1,478 | 0 | 0 | 1 | 3 | 0 | 0 | 0 | 0 | 136 | 9.0 |
| 29-Jun | 179 | 1,657 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 134 | 9.5 |
| 30-Jun | 221 | 1,878 | 0 | 0 | 2 | 5 | 0 | 0 | 0 | 0 | 134 | 9.5 |
| 1-Jul | 209 | 2,087 | 0 | 0 | 2 | 7 | 0 | 0 | 0 | 0 | 134 | 9.0 |
| 2-Jul | 245 | 2,332 | 0 | 0 | 1 | 8 | 0 | 0 | 0 | 0 | 134 | 9.5 |
| 3-Jul | 593 | 2,925 | 0 | 0 | 2 | 10 | 0 | 0 | 0 | 0 | 139 | 9.5 |
| 4-Jul | 330 | 3,255 | 0 | 0 | 1 | 11 | 0 | 0 | 0 | 0 | 148 | 12.0 |
| 5-Jul | 339 | 3,594 | 0 | 0 | 2 | 13 | 0 | 0 | 1 | 1 | 152 | 10.5 |
| 6-Jul | 270 | 3,864 | 0 | 0 | 3 | 16 | 0 | 0 | 0 | 1 | 154 | 11.0 |

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Appendix G.-Page 2 of 3.

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | Water level (cm) | $\begin{array}{r} \text { Water } \\ \text { temp }\left({ }^{\circ} \mathrm{C}\right) \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 7-Jul | 373 | 4,237 | 0 | 0 | 3 | 19 | 0 | 0 | 0 | 1 | 152 | 11.0 |
| 8-Jul | 178 | 4,415 | 0 | 0 | 1 | 20 | 0 | 0 | 0 | 1 | 146 | 11.0 |
| $9-\mathrm{Jul}$ | 541 | 4,956 | 0 | 0 | 3 | 23 | 0 | 0 | 0 | 1 | 142 | 10.5 |
| 10-Jul | 1,532 | 6,488 | 0 | 0 | 6 | 29 | 0 | 0 | 2 | 3 | 138 | 9.0 |
| 11-Jul | 991 | 7,479 | 0 | 0 | 3 | 32 | 0 | 0 | 0 | 3 | 147 | 9.0 |
| 12-Jul | 579 | 8,058 | 0 | 0 | 1 | 33 | 0 | 0 | 1 | 4 | 143 | 9.0 |
| 13-Jul | 518 | 8,576 | 0 | 0 | 3 | 36 | 0 | 0 | 3 | 7 | 138 | 9.5 |
| 14-Jul | 1,082 | 9,658 | 1 | 1 | 4 | 40 | 0 | 0 | 1 | 8 | 136 | 9.5 |
| 15-Jul | 540 | 10,198 | 0 | 1 | 5 | 45 | 0 | 0 | 1 | 9 | 134 | 9.0 |
| 16-Jul | 229 | 10,427 | 2 | 3 | 2 | 47 | 0 | 0 | 3 | 12 | 142 | 9.0 |
| 17-Jul | 257 | 10,684 | 1 | 4 | 2 | 49 | 0 | 0 | 2 | 14 | 142 | 9.5 |
| 18-Jul | 696 | 11,380 | 7 | 11 | 8 | 57 | 0 | 0 | 3 | 17 | 138 | 10.0 |
| 19-Jul | 282 | 11,662 | 3 | 14 | 1 | 58 | 0 | 0 | 0 | 17 | 136 | 9.5 |
| 20-Jul | 389 | 12,051 | 9 | 23 | 5 | 63 | 0 | 0 | 1 | 18 | 136 | 9.5 |
| 21-Jul | 8,715 | 20,766 | 68 | 91 | 14 | 77 | 0 | 0 | 2 | 20 | 140 | 10.5 |
| 22-Jul | 13,041 | 33,807 | 106 | 197 | 10 | 87 | 0 | 0 | 2 | 22 | 142 | 11.0 |
| 23-Jul | 1,583 | 35,390 | 125 | 322 | 5 | 92 | 0 | 0 | 0 | 22 | 146 | 11.5 |
| 24-Jul | 2,424 | 37,814 | 95 | 417 | 4 | 96 | 0 | 0 | 0 | 22 | 148 | 11.5 |
| 25-Jul | 4,740 | 42,554 | 61 | 478 | 6 | 102 | 0 | 0 | 3 | 25 | 144 | 11.5 |
| 26-Jul | 4,938 | 47,492 | 168 | 646 | 5 | 107 | 0 | 0 | 1 | 26 | 144 | 12.0 |
| 27-Jul | 4,770 | 52,262 | 279 | 925 | 7 | 114 | 0 | 0 | 1 | 27 | 142 | 13.0 |
| 28-Jul | 6,472 | 58,734 | 233 | 1,158 | 3 | 117 | 0 | 0 | 0 | 27 | 144 | 12.0 |
| 29-Jul | 5,002 | 63,736 | 269 | 1,427 | 13 | 130 | 0 | 0 | 0 | 27 | 145 | 12.0 |
| 30-Jul | 3,244 | 66,980 | 228 | 1,655 | 5 | 135 | 0 | 0 | 0 | 27 | 147 | 13.0 |
| 31-Jul | 1,533 | 68,513 | 113 | 1,768 | 1 | 136 | 0 | 0 | 0 | 27 | 148 | 13.5 |
| 1-Aug | 679 | 69,192 | 139 | 1,907 | 1 | 137 | 0 | 0 | 1 | 28 | 150 | 12.5 |
| 2-Aug | 425 | 69,617 | 60 | 1,967 | 1 | 138 | 0 | 0 | 1 | 29 | 148 | 12.5 |
| 3-Aug | 739 | 70,356 | 58 | 2,025 | 1 | 139 | 0 | 0 | 0 | 29 | 142 | 12.5 |
| 4-Aug | 1,640 | 71,996 | 211 | 2,236 | 1 | 140 | 0 | 0 | 0 | 29 | 140 | 12.5 |
| 5-Aug | 944 | 72,940 | 464 | 2,700 | 5 | 145 | 0 | 0 | 0 | 29 | 142 | 13.0 |
| 6-Aug | 1,073 | 74,013 | 486 | 3,186 | 1 | 146 | 0 | 0 | 0 | 29 | 140 | 12.5 |
| 7-Aug | 642 | 74,655 | 287 | 3,473 | 14 | 160 | 0 | 0 | 0 | 29 | 138 | 10.5 |
| 8-Aug | 665 | 75,320 | 93 | 3,566 | 0 | 160 | 0 | 0 | 0 | 29 | 150 | 10.0 |

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Appendix G.-Page 3 of 3 .

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | Water level (cm) | $\begin{gathered} \text { Water } \\ \text { temp }\left({ }^{\circ} \mathrm{C}\right) \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 9-Aug ${ }^{\text {a }}$ | 517 | 75,837 | 167 | 3,733 | 5 | 165 | 0 | 0 | 0 | 29 | 174 | 10.0 |
| $10-\mathrm{Aug}^{\text {a }}$ | 380 | 76,217 | 144 | 3,877 | 3 | 168 | 0 | 0 | 0 | 29 | 162 | 10.5 |
| 11-Aug | 83 | 76,300 | 37 | 3,914 | 1 | 169 | 0 | 0 | 0 | 29 | 150 | 10.5 |
| 12-Aug | 402 | 76,702 | 206 | 4,120 | 1 | 170 | 0 | 0 | 0 | 29 | 141 | 10.5 |
| 13-Aug | 402 | 77,104 | 92 | 4,212 | 2 | 172 | 0 | 0 | 0 | 29 | 137 | 10.5 |
| 14-Aug | 584 | 77,688 | 26 | 4,238 | 4 | 176 | 0 | 0 | 0 | 29 | 156 | 10.5 |
| 15-Aug | 500 | 78,188 | 46 | 4,284 | 2 | 178 | 0 | 0 | 0 | 29 | 152 | 10.0 |
| 16-Aug | 266 | 78,454 | 128 | 4,412 | 3 | 181 | 0 | 0 | 0 | 29 | 140 | 10.0 |
| 17-Aug | 922 | 79,376 | 187 | 4,599 | 5 | 186 | 0 | 0 | 0 | 29 | 132 | 10.0 |
| 18-Aug | 616 | 79,992 | 90 | 4,689 | 4 | 190 | 0 | 0 | 0 | 29 | 128 | 10.0 |
| 19-Aug | 1,193 | 81,185 | 131 | 4,820 | 3 | 193 | 0 | 0 | 0 | 29 | 124 | 10.0 |
| 20-Aug | 562 | 81,747 | 94 | 4,914 | 0 | 193 | 0 | 0 | 1 | 30 | 124 | 10.0 |
| 21-Aug | 343 | 82,090 | 53 | 4,967 | 0 | 193 | 0 | 0 | 0 | 30 | 125 | 10.5 |
| 22-Aug | 176 | 82,266 | 54 | 5,021 | 1 | 194 | 0 | 0 | 0 | 30 | 134 | 10.0 |
| 23-Aug | 258 | 82,524 | 42 | 5,063 | 3 | 197 | 0 | 0 | 0 | 30 | 134 | 11.0 |
| 24-Aug | 1,002 | 83,526 | 98 | 5,161 | 1 | 198 | 0 | 0 | 0 | 30 | 130 | 10.0 |
| 25-Aug | 198 | 83,724 | 63 | 5,224 | 2 | 200 | 0 | 0 | 0 | 30 | 138 | 10.5 |
| 26-Aug | 174 | 83,898 | 64 | 5,288 | 1 | 201 | 0 | 0 | 0 | 30 | 133 | 11.0 |
| 27-Aug | 802 | 84,700 | 69 | 5,357 | 6 | 207 | 0 | 0 | 0 | 30 | 133 | 10.5 |
| 28-Aug | 120 | 84,820 | 58 | 5,415 | 1 | 208 | 0 | 0 | 0 | 30 | 138 | 10.5 |
| 29-Aug | 89 | 84,909 | 21 | 5,436 | 2 | 210 | 1 | 1 | 0 | 30 | 151 | 10.5 |
| 30-Aug | 18 | 84,927 | 11 | 5,447 | 2 | 212 | 0 | 1 | 0 | 30 | 142 | 10.0 |
| 31-Aug | 34 | 84,961 | 5 | 5,452 | 2 | 214 | 1 | 2 | 0 | 30 | 135 | 10.0 |
| 1-Sep | 38 | 84,999 | 7 | 5,459 | 2 | 216 | 2 | 4 | 0 | 30 | 130 | 10.0 |
| 2-Sep | 106 | 85,105 | 3 | 5,462 | 0 | 216 | 1 | 5 | 0 | 30 | 128 | 10.0 |
| 3-Sep | 46 | 85,151 | 1 | 5,463 | 0 | 216 | 0 | 5 | 0 | 30 | 124 | 10.0 |
| 4-Sep | 33 | 85,184 | 2 | 5,465 | 2 | 218 | 0 | 5 | 0 | 30 | 123 | 10.5 |
| 5-Sep | 5 | 85,189 | 0 | 5,465 | 0 | 218 | 0 | 5 | 0 | 30 | 121 | 10.0 |
| 6-Sep | 76 | 85,265 | 1 | 5,466 | 3 | 221 | 2 | 7 | 1 | 31 | 120 | 11.0 |
| 7-Sep | 46 | 85,311 | 6 | 5,472 | 3 | 224 | 28 | 35 | 0 | 31 | 118 | 11.0 |
| 8-Sep | 152 | 85,463 | 3 | 5,475 | 1 | 225 | 60 | 95 | 0 | 31 | 115 | 11.5 |

[^10]Appendix H.-Daily and cumulative (cum.) Chilkoot River weir salmon counts by species, and water temperature and gauge heights, 2019.

| Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | Water level (cm) | Water temp. $\left({ }^{\circ} \mathrm{C}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
| 7-Jun | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 140 | 8.0 |
| 8-Jun | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 142 | 8.5 |
| 9-Jun | 2 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 139 | 8.5 |
| 10-Jun | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 136 | 10.0 |
| 11-Jun | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 135 | 8.5 |
| 12-Jun | 12 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 146 | 8.0 |
| 13-Jun | 0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 150 | 8.0 |
| 14-Jun | 3 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 146 | 9.0 |
| 15-Jun | 0 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 144 | 9.5 |
| 16-Jun | 5 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 146 | 9.0 |
| 17-Jun | 66 | 90 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 154 | 8.5 |
| 18-Jun | 20 | 110 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 165 | 8.0 |
| 19-Jun | 153 | 263 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 160 | 8.0 |
| 20-Jun | 147 | 410 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 7 | 152 | 8.5 |
| 21-Jun | 126 | 536 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 153 | 8.5 |
| 22-Jun | 407 | 943 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 8 | 150 | 9.0 |
| 23-Jun | 517 | 1,460 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 9 | 148 | 9.5 |
| 24-Jun | 381 | 1,841 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 11 | 146 | 11.0 |
| 25-Jun | 245 | 2,086 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 12 | 145 | 9.0 |
| 26-Jun | 617 | 2,703 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 15 | 150 | 10.5 |
| 27-Jun | 506 | 3,209 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 17 | 155 | 12.0 |
| 28-Jun | 544 | 3,753 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 18 | 155 | 12.5 |
| 29-Jun | 513 | 4,266 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 18 | 154 | 10.5 |
| 30-Jun | 195 | 4,461 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 18 | 160 | 11.5 |
| 1-Jul | 564 | 5,025 | 27 | 27 | 0 | 0 | 0 | 0 | 0 | 18 | 158 | 12.5 |
| 2-Jul | 656 | 5,681 | 7 | 34 | 0 | 0 | 0 | 0 | 0 | 18 | 158 | 12.0 |
| 3-Jul | 797 | 6,478 | 5 | 39 | 0 | 0 | 0 | 0 | 1 | 19 | 160 | 12.5 |
| 4-Jul | 467 | 6,945 | 8 | 47 | 0 | 0 | 0 | 0 | 1 | 20 | 157 | 12.5 |
| 5-Jul | 727 | 7,672 | 17 | 64 | 0 | 0 | 0 | 0 | 1 | 21 | 156 | 13.0 |
| 6-Jul | 1,103 | 8,775 | 7 | 71 | 0 | 0 | 0 | 0 | 7 | 28 | 157 | 13.0 |
| 7-Jul | 764 | 9,539 | 43 | 114 | 0 | 0 | 0 | 0 | 3 | 31 | 155 | 12.5 |
| 8-Jul | 928 | 10,467 | 52 | 166 | 0 | 0 | 0 | 0 | 0 | 31 | 154 | 11.5 |
| 9-Jul | 860 | 11,327 | 41 | 207 | 0 | 0 | 0 | 0 | 0 | 31 | 148 | 13.0 |
| 10-Jul | 817 | 12,144 | 13 | 220 | 0 | 0 | 0 | 0 | 0 | 31 | 142 | 12.0 |

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|  | Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | $\begin{gathered} \text { Water } \\ \text { level }(\mathrm{cm}) \\ \hline \end{gathered}$ | Water <br> temp. $\left({ }^{\circ} \mathrm{C}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
|  | 11-Jul | 1,417 | 13,561 | 15 | 235 | 0 | 0 | 0 | 0 | 0 | 31 | 142 | 12.0 |
|  | 12-Jul | 2,571 | 16,132 | 12 | 247 | 0 | 0 | 0 | 0 | 2 | 33 | 138 | 11.0 |
|  | 13-Jul | 2,361 | 18,493 | 8 | 255 | 0 | 0 | 0 | 0 | 1 | 34 | 134 | 11.0 |
|  | 14-Jul | 1,905 | 20,398 | 72 | 327 | 1 | 1 | 0 | 0 | 5 | 39 | 130 | 11.0 |
|  | 15-Jul | 1,457 | 21,855 | 74 | 401 | 3 | 4 | 0 | 0 | 1 | 40 | 130 | 11.5 |
|  | 16-Jul | 2,144 | 23,999 | 63 | 464 | 4 | 8 | 0 | 0 | 4 | 44 | 130 | 11.0 |
|  | 17-Jul | 1,041 | 25,040 | 39 | 503 | 2 | 10 | 0 | 0 | 2 | 46 | 132 | 11.5 |
|  | 18-Jul | 2,859 | 27,899 | 82 | 585 | 2 | 12 | 0 | 0 | 3 | 49 | 133 | 12.0 |
|  | 19-Jul | 1,128 | 29,027 | 62 | 647 | 4 | 16 | 0 | 0 | 0 | 49 | 136 | 10.5 |
|  | 20-Jul | 1,538 | 30,565 | 45 | 692 | 1 | 17 | 0 | 0 | 1 | 50 | 135 | 11.0 |
|  | 21-Jul | 1,323 | 31,888 | 40 | 732 | 5 | 22 | 0 | 0 | 4 | 54 | 134 | 10.5 |
|  | 22-Jul | 1,408 | 33,296 | 80 | 812 | 18 | 40 | 0 | 0 | 0 | 54 | 134 | 12.0 |
|  | 23-Jul | 5,568 | 38,864 | 191 | 1,003 | 18 | 58 | 0 | 0 | 0 | 54 | 136 | 10.5 |
|  | 24-Jul | 8,762 | 47,626 | 215 | 1,218 | 16 | 74 | 0 | 0 | 5 | 59 | 135 | 11.0 |
|  | 25-Jul | 13,069 | 60,695 | 419 | 1,637 | 6 | 80 | 0 | 0 | 1 | 60 | 134 | 10.5 |
| 8 | 26-Jul | 11,225 | 71,920 | 238 | 1,875 | 5 | 85 | 0 | 0 | 1 | 61 | 133 | 11.0 |
|  | 27-Jul | 14,723 | 86,643 | 553 | 2,428 | 13 | 98 | 0 | 0 | 0 | 61 | 134 | 11.5 |
|  | 28-Jul | 12,047 | 98,690 | 145 | 2,573 | 25 | 123 | 0 | 0 | 0 | 61 | 142 | 9.0 |
|  | 29-Jul | 2,582 | 101,272 | 88 | 2,661 | 4 | 127 | 0 | 0 | 0 | 61 | 145 | 9.0 |
|  | 30-Jul | 1,401 | 102,673 | 476 | 3,137 | 2 | 129 | 0 | 0 | , | 62 | 142 | 10.0 |
|  | 31-Jul | 4,267 | 106,940 | 1,357 | 4,494 | 2 | 131 | 0 | 0 | 0 | 62 | 138 | 10.5 |
|  | 1-Aug | 3,518 | 110,458 | 1,414 | 5,908 | 0 | 131 | 0 | 0 | 0 | 62 | 138 | 10.0 |
|  | 2-Aug | 1,508 | 111,966 | 794 | 6,702 | 1 | 132 | 0 | 0 | 0 | 62 | 136 | 10.5 |
|  | 3-Aug | 2,123 | 114,089 | 582 | 7,284 | 1 | 133 | 0 | 0 | 1 | 63 | 135 | 10.5 |
|  | 4-Aug | 1,854 | 115,943 | 484 | 7,768 | 8 | 141 | 0 | 0 | 0 | 63 | 140 | 11.5 |
|  | 5-Aug | 1,194 | 117,137 | 572 | 8,340 | 2 | 143 | 0 | 0 | 0 | 63 | 142 | 11.5 |
|  | 6-Aug | 1,452 | 118,589 | 359 | 8,699 | 2 | 145 | 0 | 0 | 0 | 63 | 144 | 11.5 |
|  | 7-Aug | 907 | 119,496 | 189 | 8,888 | 0 | 145 | 0 | 0 | 0 | 63 | 150 | 12.0 |
|  | 8-Aug | 529 | 120,025 | 147 | 9,035 | 1 | 146 | 0 | 0 | 0 | 63 | 149 | 13.0 |
|  | 9-Aug | 806 | 120,831 | 410 | 9,445 | 9 | 155 | 0 | 0 | 0 | 63 | 140 | 13.5 |
|  | 10-Aug | 408 | 121,239 | 95 | 9,540 | 0 | 155 | 0 | 0 | 0 | 63 | 135 | 12.5 |
|  | 11-Aug | 1,696 | 122,935 | 371 | 9,911 | 3 | 158 | 0 | 0 | 0 | 63 | 132 | 13.0 |
|  | 12-Aug | 2,135 | 125,070 | 401 | 10,312 | 10 | 168 | 0 | 0 | 0 | 63 | 130 | 12.5 |
|  | 13-Aug | 1,454 | 126,524 | 362 | 10,674 | 5 | 173 | 0 | 0 | 1 | 64 | 134 | 12.0 |

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|  | Date | Sockeye salmon |  | Pink salmon |  | Chum salmon |  | Coho salmon |  | Chinook salmon |  | Water level (cm) | Water temp. $\qquad$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |  |  |
|  | 14-Aug | 1,041 | 127,565 | 317 | 10,991 | 1 | 174 | 0 | 0 | 0 | 64 | 132 | 11.5 |
|  | 15-Aug | 1,110 | 128,675 | 229 | 11,220 | 0 | 174 | 0 | 0 | 0 | 64 | 131 | 10.5 |
|  | 16-Aug | 1,088 | 129,763 | 242 | 11,462 | 1 | 175 | 0 | 0 | 0 | 64 | 133 | ND |
|  | 17-Aug | 658 | 130,421 | 374 | 11,836 | 1 | 176 | 0 | 0 | 0 | 64 | 135 | ND |
|  | 18-Aug | 491 | 130,912 | 306 | 12,142 | 9 | 185 | 0 | 0 | 0 | 64 | 135 | 11.0 |
|  | 19-Aug | 1,328 | 132,240 | 499 | 12,641 | 9 | 194 | 0 | 0 | 0 | 64 | 124 | 11.5 |
|  | 20-Aug | 561 | 132,801 | 174 | 12,815 | 48 | 242 | 0 | 0 | 0 | 64 | 120 | 11.0 |
|  | 21-Aug | 692 | 133,493 | 77 | 12,892 | 7 | 249 | 0 | 0 | 0 | 64 | 114 | 11.0 |
|  | 22-Aug | 376 | 133,869 | 97 | 12,989 | 1 | 250 | 0 | 0 | 0 | 64 | 113 | 10.5 |
|  | 23-Aug | 424 | 134,293 | 85 | 13,074 | 2 | 252 | 0 | 0 | 0 | 64 | 114 | 10.5 |
|  | 24-Aug | 728 | 135,021 | 144 | 13,218 | 1 | 253 | 0 | 0 | 0 | 64 | 118 | 10.0 |
|  | 25-Aug | 530 | 135,551 | 173 | 13,391 | 4 | 257 | 4 | 4 | 0 | 64 | 126 | 9.5 |
|  | 26-Aug | 506 | 136,057 | 171 | 13,562 | 11 | 268 | 9 | 13 | 0 | 64 | 126 | 9.5 |
|  | 27-Aug | 716 | 136,773 | 175 | 13,737 | 21 | 289 | 5 | 18 | 0 | 64 | 132 | 10.0 |
|  | 28-Aug | 325 | 137,098 | 239 | 13,976 | 2 | 291 | 1 | 19 | 0 | 64 | 134 | 10.0 |
| の | 29-Aug | 257 | 137,355 | 342 | 14,318 | 0 | 291 | 0 | 19 | 0 | 64 | 128 | 9.5 |
|  | 30-Aug | 417 | 137,772 | 584 | 14,902 | 2 | 293 | 1 | 20 | 0 | 64 | 128 | 10.5 |
|  | 31-Aug | 707 | 138,479 | 696 | 15,598 | 4 | 297 | 3 | 23 | 0 | 64 | 126 | 10.5 |
|  | 1-Sep | 375 | 138,854 | 395 | 15,993 | 0 | 297 | 7 | 30 | 0 | 64 | 124 | 10.5 |
|  | 2-Sep | 339 | 139,193 | 134 | 16,127 | 1 | 298 | 3 | 33 | 0 | 64 | 122 | 9.0 |
|  | 3-Sep | 246 | 139,439 | 217 | 16,344 | 1 | 299 | 3 | 36 | 0 | 64 | 122 | 10.0 |
|  | 4-Sep | 311 | 139,750 | 258 | 16,602 | 2 | 301 | 5 | 41 | 0 | 64 | 122 | 10.0 |
|  | 5-Sep | 173 | 139,923 | 141 | 16,743 | 1 | 302 | 2 | 43 | 0 | 64 | 120 | 10.0 |
|  | 6-Sep | 112 | 140,035 | 161 | 16,904 | 27 | 329 | 12 | 55 | 0 | 64 | 118 | 10.0 |
|  | 7-Sep | 138 | 140,173 | 122 | 17,026 | 46 | 375 | 7 | 62 | 0 | 64 | 118 | 11.0 |
|  | 8-Sep | 205 | 140,378 | 130 | 17,156 | 21 | 396 | 18 | 80 | 0 | 64 | 120 | 11.0 |

Appendix I.-Estimated commercial harvest of Chilkoot Lake, Chilkat Lake, and other sockeye salmon stocks in the District 15 commercial drift gillnet fishery based on scale pattern analysis (1976-2016) and genetic stock identification (2017-2019).

| Year | Harvest |  |  | Percentile rank |  |  | Percent of harvest |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Chilkoot | Chilkat |  | Chilkoot | Chilkat |  | Chilkoot | Chilkat |  |
|  | Lake | Lake | Other ${ }^{\text {a }}$ | Lake | Lake | Other | Lake | Lake | Other |
| 1976 | 61,861 | 58,765 | 4,796 | 0.51 | 0.42 | 0.12 | 49\% | 47\% | 4\% |
| 1977 | 113,555 | 41,477 | 5,389 | 0.65 | 0.23 | 0.14 | 71\% | 26\% | 3\% |
| 1978 | 14,264 | 89,558 | 4,658 | 0.12 | 0.67 | 0.09 | 13\% | 83\% | 4\% |
| 1979 | 69,864 | 115,995 | 7,117 | 0.58 | 0.81 | 0.16 | 36\% | 60\% | 4\% |
| 1980 | 21,244 | 31,267 | 1,588 | 0.21 | 0.16 | 0.02 | 39\% | 58\% | 3\% |
| 1981 | 43,756 | 48,420 | 1,070 | 0.44 | 0.33 | 0.00 | 47\% | 52\% | 1\% |
| 1982 | 144,748 | 127,174 | 1,911 | 0.81 | 0.88 | 0.05 | 53\% | 46\% | 1\% |
| 1983 | 242,034 | 124,180 | 3,965 | 0.93 | 0.84 | 0.07 | 65\% | 34\% | 1\% |
| 1984 | 225,634 | 99,592 | 9,502 | 0.88 | 0.70 | 0.19 | 67\% | 30\% | 3\% |
| 1985 | 153,533 | 131,091 | 18,704 | 0.84 | 0.91 | 0.49 | 51\% | 43\% | 6\% |
| 1986 | 110,114 | 168,006 | 12,174 | 0.60 | 1.00 | 0.30 | 38\% | 58\% | 4\% |
| 1987 | 327,323 | 69,900 | 18,658 | 1.00 | 0.51 | 0.47 | 79\% | 17\% | 4\% |
| 1988 | 248,640 | 76,883 | 26,353 | 0.95 | 0.58 | 0.74 | 71\% | 22\% | 7\% |
| 1989 | 292,830 | 156,160 | 25,908 | 0.98 | 0.98 | 0.72 | 62\% | 33\% | 5\% |
| 1990 | 181,260 | 149,377 | 31,499 | 0.86 | 0.93 | 0.81 | 50\% | 41\% | 9\% |
| 1991 | 228,607 | 60,721 | 24,353 | 0.91 | 0.47 | 0.67 | 73\% | 19\% | 8\% |
| 1992 | 142,471 | 113,146 | 33,729 | 0.79 | 0.79 | 0.91 | 49\% | 39\% | 12\% |
| 1993 | 52,080 | 103,531 | 19,605 | 0.47 | 0.74 | 0.56 | 30\% | 59\% | 11\% |
| 1994 | 25,367 | 126,852 | 19,578 | 0.28 | 0.86 | 0.53 | 15\% | 74\% | 11\% |
| 1995 | 9,637 | 68,737 | 10,302 | 0.09 | 0.49 | 0.23 | 11\% | 78\% | 12\% |
| 1996 | 19,882 | 99,677 | 30,019 | 0.19 | 0.72 | 0.79 | 13\% | 67\% | 20\% |
| 1997 | 31,822 | 73,761 | 13,245 | 0.35 | 0.53 | 0.35 | 27\% | 62\% | 11\% |
| 1998 | 2,838 | 112,630 | 19,469 | 0.02 | 0.77 | 0.51 | 2\% | 83\% | 14\% |
| 1999 | 4,604 | 149,410 | 9,547 | 0.05 | 0.95 | 0.21 | 3\% | 91\% | 6\% |
| 2000 | 14,622 | 78,265 | 16,673 | 0.14 | 0.60 | 0.40 | 13\% | 71\% | 15\% |
| 2001 | 66,355 | 60,183 | 21,273 | 0.53 | 0.44 | 0.60 | 45\% | 41\% | 14\% |
| 2002 | 24,200 | 47,332 | 10,482 | 0.26 | 0.28 | 0.28 | 30\% | 58\% | 13\% |
| 2003 | 32,446 | 49,955 | 12,729 | 0.40 | 0.35 | 0.33 | 34\% | 53\% | 13\% |
| 2004 | 66,498 | 51,110 | 33,637 | 0.56 | 0.37 | 0.88 | 44\% | 34\% | 22\% |
| 2005 | 29,276 | 22,852 | 13,341 | 0.33 | 0.14 | 0.37 | 45\% | 35\% | 20\% |
| 2006 | 119,201 | 15,979 | 10,400 | 0.67 | 0.07 | 0.26 | 82\% | 11\% | 7\% |
| 2007 | 125,199 | 14,208 | 17,529 | 0.74 | 0.02 | 0.44 | 80\% | 9\% | 11\% |
| 2008 | 7,491 | 22,156 | 17,008 | 0.07 | 0.12 | 0.42 | 16\% | 47\% | 36\% |
| 2009 | 16,622 | 85,551 | 24,422 | 0.16 | 0.65 | 0.70 | 13\% | 68\% | 19\% |
| 2010 | 32,064 | 48,079 | 20,830 | 0.37 | 0.30 | 0.58 | 32\% | 48\% | 21\% |
| 2011 | 26,766 | 15,599 | 21,428 | 0.30 | 0.05 | 0.63 | 42\% | 24\% | 34\% |
| 2012 | 124,366 | 54,884 | 45,393 | 0.72 | 0.40 | 0.98 | 55\% | 24\% | 20\% |
| 2013 | 23,111 | 75,588 | 23,404 | 0.23 | 0.56 | 0.65 | 19\% | 62\% | 19\% |
| 2014 | 110,487 | 81,502 | 42,693 | 0.63 | 0.63 | 0.95 | 47\% | 35\% | 18\% |
| 2015 | 58,568 | 33,085 | 39,924 | 0.49 | 0.19 | 0.93 | 45\% | 25\% | 30\% |
| 2016 | 119,843 | 35,991 | 33,010 | 0.70 | 0.21 | 0.86 | 63\% | 19\% | 17\% |
| 2017 | 1,933 | 5,698 | 32,085 | 0.00 | 0.00 | 0.84 | 14\% | 5\% | 81\% |
| 2018 | 33,969 | 19,235 | 28,483 | 0.42 | 0.09 | 0.77 | 42\% | 24\% | 35\% |
| 2019 | 149,586 | 40,935 | 51,012 | 0.81 | 0.23 | 1.00 | 62\% | 17\% | 21\% |
| Average ${ }^{\text {b }}$ | 91,131 | 78,816 | 18,108 |  |  |  | 41\% | 47\% | 12\% |
| Median ${ }^{\text {b }}$ | 60,214 | 74,675 | 18,094 |  |  |  | 44\% | 47\% | 11\% |

Note: Bold estimates are historical records that have been updated since the last project report by Bednarski et al. (2016).
a Other includes Chilkat River mainstem spawning stocks (1976-2019).
b Average and median values use 1976-2015 data.

Appendix J.-District 15 commercial drift gillnet fishery data used in genetic stock identification analysis and results by statistical week and reporting group, 2017.

| Stat. week | Sample <br> size | Genotyped | Aged only | Not genotyped or aged | Reporting group | Mean | SD | CI 5\% | CI 95\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 25 | 300 | 185 | 95 | 20 | Chilkat Lake | 0.131 | 0.026 | 0.091 | 0.175 |
|  |  |  |  |  | Chilkat Mainstem | 0.005 | 0.010 | 0.000 | 0.027 |
|  |  |  |  |  | Chilkoot | 0.045 | 0.016 | 0.023 | 0.074 |
|  |  |  |  |  | Other | 0.819 | 0.030 | 0.767 | 0.866 |
| 26 | 305 | 187 | 95 | 23 | Chilkat Lake | 0.114 | 0.025 | 0.076 | 0.157 |
|  |  |  |  |  | Chilkat Mainstem | 0.033 | 0.015 | 0.011 | 0.061 |
|  |  |  |  |  | Chilkoot | 0.064 | 0.020 | 0.035 | 0.099 |
|  |  |  |  |  | Other | 0.790 | 0.033 | 0.735 | 0.842 |
| 27 | 312 | 185 | 104 | 23 | Chilkat Lake | 0.209 | 0.032 | 0.159 | 0.262 |
|  |  |  |  |  | Chilkat Mainstem | 0.062 | 0.019 | 0.033 | 0.096 |
|  |  |  |  |  | Chilkoot | 0.065 | 0.021 | 0.034 | 0.101 |
|  |  |  |  |  | Other | 0.664 | 0.038 | 0.601 | 0.725 |
| 28 | 190 | 188 | 2 | 0 | Chilkat Lake | 0.164 | 0.039 | 0.103 | 0.233 |
|  |  |  |  |  | Chilkat Mainstem | 0.005 | 0.010 | 0.000 | 0.027 |
|  |  |  |  |  | Chilkoot | 0.113 | 0.033 | 0.065 | 0.170 |
|  |  |  |  |  | Other | 0.718 | 0.049 | 0.634 | 0.795 |
| 29 | 360 | 187 | 152 | 21 | Chilkat Lake | 0.057 | 0.018 | 0.031 | 0.089 |
|  |  |  |  |  | Chilkat Mainstem | 0.006 | 0.007 | 0.000 | 0.020 |
|  |  |  |  |  | Chilkoot | 0.024 | 0.013 | 0.007 | 0.049 |
|  |  |  |  |  | Other | 0.912 | 0.022 | 0.872 | 0.946 |
| 30 | 305 | 183 | 113 | 9 | Chilkat Lake | 0.060 | 0.019 | 0.032 | 0.093 |
|  |  |  |  |  | Chilkat Mainstem | 0.014 | 0.012 | 0.000 | 0.037 |
|  |  |  |  |  | Chilkoot | 0.035 | 0.015 | 0.014 | 0.061 |
|  |  |  |  |  | Other | 0.891 | 0.026 | 0.846 | 0.930 |
| 31 | 240 | 185 | 49 | 6 | Chilkat Lake | 0.088 | 0.023 | 0.053 | 0.129 |
|  |  |  |  |  | Chilkat Mainstem | 0.001 | 0.003 | 0.000 | 0.007 |
|  |  |  |  |  | Chilkoot | 0.056 | 0.020 | 0.026 | 0.092 |
|  |  |  |  |  | Other | 0.855 | 0.029 | 0.804 | 0.900 |
| 32 | 303 | 170 | 111 | 22 | Chilkat Lake | 0.365 | 0.039 | 0.303 | 0.429 |
|  |  |  |  |  | Chilkat Mainstem | 0.028 | 0.013 | 0.010 | 0.051 |
|  |  |  |  |  | Chilkoot | 0.050 | 0.018 | 0.024 | 0.083 |
|  |  |  |  |  | Other | 0.558 | 0.040 | 0.491 | 0.624 |
| 33 | 300 | 184 | 101 | 15 | Chilkat Lake | 0.175 | 0.031 | 0.125 | 0.228 |
|  |  |  |  |  | Chilkat Mainstem | 0.005 | 0.007 | 0.000 | 0.020 |
|  |  |  |  |  | Chilkoot | 0.085 | 0.022 | 0.051 | 0.124 |
|  |  |  |  |  | Other | 0.736 | 0.037 | 0.673 | 0.796 |
| 34-40 | 300 | 186 | 98 | 16 | Chilkat Lake | 0.136 | 0.026 | 0.094 | 0.181 |
|  |  |  |  |  | Chilkat Mainstem | 0.001 | 0.003 | 0.000 | 0.008 |
|  |  |  |  |  | Chilkoot | 0.048 | 0.018 | 0.021 | 0.080 |
|  |  |  |  |  | Other | 0.815 | 0.030 | 0.764 | 0.862 |
| All | 2915 | 1,840 | 920 | 155 | Chilkat Lake | 0.143 | 0.010 | 0.127 | 0.160 |
|  |  |  |  |  | Chilkat Mainstem | 0.013 | 0.003 | 0.008 | 0.019 |
|  |  |  |  |  | Chilkoot | 0.049 | 0.007 | 0.038 | 0.060 |
|  |  |  |  |  | Other | 0.795 | 0.011 | 0.776 | 0.813 |

Appendix K.-District 15 commercial drift gillnet fishery data used in genetic stock identification analysis and results by statistical week and reporting group, 2018.

| Stat. week | Sample size | Genotyped | $\begin{gathered} \text { Aged } \\ \text { only } \end{gathered}$ | $\begin{gathered} \text { Not genotyped } \\ \text { or aged } \end{gathered}$ | Reporting group | Mean | SD | CI 5\% | CI 95\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 25 | 126 | 121 | 5 | 0 | Chilkat Lake | 0.273 | 0.043 | 0.204 | 0.346 |
|  |  |  |  |  | Chilkat Mainstem | 0.061 | 0.026 | 0.022 | 0.109 |
|  |  |  |  |  | Chilkoot | 0.178 | 0.035 | 0.124 | 0.237 |
|  |  |  |  |  | Other | 0.488 | 0.048 | 0.410 | 0.567 |
| 26 | 326 | 186 | 100 | 40 | Chilkat Lake | 0.264 | 0.035 | 0.208 | 0.322 |
|  |  |  |  |  | Chilkat Mainstem | 0.104 | 0.024 | 0.068 | 0.145 |
|  |  |  |  |  | Chilkoot | 0.081 | 0.022 | 0.049 | 0.120 |
|  |  |  |  |  | Other | 0.551 | 0.040 | 0.485 | 0.615 |
| 27 | 413 | 183 | 195 | 35 | Chilkat Lake | 0.240 | 0.031 | 0.190 | 0.293 |
|  |  |  |  |  | Chilkat Mainstem | 0.060 | 0.023 | 0.025 | 0.100 |
|  |  |  |  |  | Chilkoot | 0.160 | 0.027 | 0.118 | 0.205 |
|  |  |  |  |  | Other | 0.540 | 0.040 | 0.475 | 0.606 |
| 28 | 503 | 186 | 261 | 56 | Chilkat Lake | 0.158 | 0.028 | 0.116 | 0.207 |
|  |  |  |  |  | Chilkat Mainstem | 0.023 | 0.015 | 0.001 | 0.050 |
|  |  |  |  |  | Chilkoot | 0.268 | 0.033 | 0.214 | 0.323 |
|  |  |  |  |  | Other | 0.551 | 0.038 | 0.488 | 0.613 |
| 29 | 390 | 186 | 166 | 38 | Chilkat Lake | 0.168 | 0.029 | 0.122 | 0.217 |
|  |  |  |  |  | Chilkat Mainstem | 0.001 | 0.003 | 0.000 | 0.006 |
|  |  |  |  |  | Chilkoot | 0.283 | 0.035 | 0.227 | 0.341 |
|  |  |  |  |  | Other | 0.548 | 0.040 | 0.483 | 0.613 |
| 30 | 310 | 188 | 89 | 33 | Chilkat Lake |  | 0.030 | 0.160 | 0.259 |
|  |  |  |  |  | Chilkat Mainstem | 0.008 | 0.010 | 0.000 | 0.028 |
|  |  |  |  |  | Chilkoot | 0.392 | 0.036 | 0.332 | 0.452 |
|  |  |  |  |  | Other | 0.392 | 0.037 | 0.331 | 0.454 |
| 31 | 399 | 187 | 166 | 46 |  |  |  |  |  |
|  |  |  |  |  | Chilkat Mainstem | 0.003 | 0.006 | 0.000 | 0.015 |
|  |  |  |  |  | Chilkoot | 0.503 | 0.037 | 0.442 | 0.562 |
|  |  |  |  |  | Other | 0.282 | 0.035 | 0.225 | 0.341 |
| 32 | 300 | 182 | 88 | 30 |  |  | 0.030 | 0.230 |  |
|  |  |  |  |  | Chilkat Mainstem | 0.001 | 0.002 | 0.000 | 0.004 |
|  |  |  |  |  | Chilkoot | 0.417 | 0.035 | 0.359 | 0.476 |
|  |  |  |  |  | Other | 0.305 | 0.034 | 0.251 | 0.361 |
| 33 | 340 | 189 | 114 | 37 | Chilkat Lake | 0.262 | 0.029 | 0.215 | 0.311 |
|  |  |  |  |  | Chilkat Mainstem | 0.001 | 0.002 | 0.000 | 0.005 |
|  |  |  |  |  | Chilkoot | 0.239 | 0.031 | 0.189 | 0.291 |
|  |  |  |  |  | Other | 0.499 | 0.036 | 0.440 | 0.558 |
| 34-40 | 300 | 186 | 80 | 34 | Chilkat Lake | 0.283 | 0.030 | 0.234 | 0.333 |
|  |  |  |  |  | Chilkat Mainstem | 0.004 | 0.006 | 0.000 | 0.016 |
|  |  |  |  |  | Chilkoot | 0.565 | 0.034 | 0.508 | 0.620 |
|  |  |  |  |  | Other | 0.149 | 0.027 | 0.107 | 0.194 |
| All | 3,407 | 1,794 | 1,264 | 349 | Chilkat Lake | 0.235 | 0.012 | 0.216 | 0.255 |
|  |  |  |  |  | Chilkat Mainstem | 0.009 | 0.003 | 0.005 | 0.014 |
|  |  |  |  |  | Chilkoot | 0.416 | 0.014 | 0.393 | 0.439 |
|  |  |  |  |  | Other | 0.340 | 0.014 | 0.318 | 0.363 |

Appendix L.-District 15 commercial drift gillnet fishery data used in genetic stock identification analysis and results by statistical week and reporting group, 2019.

| Stat. week | Sample <br> size | Genotyped | Aged only | Not genotyped or aged | Reporting group | Mean | SD | CI 5\% | CI 95\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 25-26 | 539 | 187 | 297 | 55 | Chilkat Lake | 0.266 | 0.052 | 0.187 | 0.357 |
|  |  |  |  |  | Chilkat Mainstem | 0.191 | 0.050 | 0.109 | 0.273 |
|  |  |  |  |  | Chilkoot | 0.165 | 0.026 | 0.124 | 0.211 |
|  |  |  |  |  | Other | 0.378 | 0.039 | 0.315 | 0.444 |
| 27 | 418 | 188 | 186 | 44 | Chilkat Lake | 0.162 | 0.030 | 0.116 | 0.213 |
|  |  |  |  |  | Chilkat Mainstem | 0.149 | 0.032 | 0.099 | 0.203 |
|  |  |  |  |  | Chilkoot | 0.202 | 0.028 | 0.156 | 0.249 |
|  |  |  |  |  | Other | 0.488 | 0.040 | 0.422 | 0.555 |
| 28 | 448 | 190 | 212 | 46 | Chilkat Lake | 0.208 | 0.030 | 0.159 | 0.259 |
|  |  |  |  |  | Chilkat Mainstem | 0.100 | 0.024 | 0.062 | 0.142 |
|  |  |  |  |  | Chilkoot | 0.277 | 0.032 | 0.225 | 0.331 |
|  |  |  |  |  | Other | 0.416 | 0.038 | 0.354 | 0.478 |
| 29 | 289 | 188 | 90 | 11 | Chilkat Lake | 0.217 | 0.030 | 0.168 | 0.268 |
|  |  |  |  |  | Chilkat Mainstem | 0.125 | 0.027 | 0.083 | 0.171 |
|  |  |  |  |  | Chilkoot | 0.493 | 0.035 | 0.436 | 0.550 |
|  |  |  |  |  | Other | 0.165 | 0.030 | 0.118 | 0.216 |
| 30 | 350 | 188 | 151 | 11 | Chilkat Lake | 0.154 | 0.028 | 0.111 | 0.202 |
|  |  |  |  |  | Chilkat Mainstem | 0.060 | 0.020 | 0.030 | 0.095 |
|  |  |  |  |  | Chilkoot | 0.597 | 0.035 | 0.539 | 0.654 |
|  |  |  |  |  | Other | 0.188 | 0.031 | 0.139 | 0.240 |
| 31 | 350 | 187 | 141 | 22 | Chilkat Lake | 0.135 | 0.027 | 0.094 | 0.181 |
|  |  |  |  |  | Chilkat Mainstem | 0.047 | 0.015 | 0.025 | 0.073 |
|  |  |  |  |  | Chilkoot | 0.770 | 0.030 | 0.719 | 0.818 |
|  |  |  |  |  | Other | 0.049 | 0.019 | 0.020 | 0.083 |
| 32 | 470 | 186 | 256 | 28 | Chilkat Lake | 0.111 | 0.023 | 0.075 | 0.151 |
|  |  |  |  |  | Chilkat Mainstem | 0.020 | 0.011 | 0.005 | 0.041 |
|  |  |  |  |  | Chilkoot | 0.787 | 0.028 | 0.740 | 0.830 |
|  |  |  |  |  | Other | 0.082 | 0.020 | 0.052 | 0.118 |
| 33 | 330 | 188 | 127 | 15 | Chilkat Lake | 0.233 | 0.031 | 0.184 | 0.287 |
|  |  |  |  |  | Chilkat Mainstem | 0.008 | 0.010 | 0.000 | 0.029 |
|  |  |  |  |  | Chilkoot | 0.536 | 0.034 | 0.479 | 0.593 |
|  |  |  |  |  | Other | 0.222 | 0.033 | 0.170 | 0.277 |
| 34 | 310 | 188 | 101 | 21 | Chilkat Lake | 0.201 | 0.026 | 0.160 | 0.244 |
|  |  |  |  |  | Chilkat Mainstem | 0.001 | 0.004 | 0.000 | 0.008 |
|  |  |  |  |  | Chilkoot | 0.532 | 0.035 | 0.475 | 0.590 |
|  |  |  |  |  | Other | 0.266 | 0.033 | 0.214 | 0.321 |
| 35-40 | 299 | 190 | 100 | 9 | Chilkat Lake | 0.327 | 0.032 | 0.275 | 0.380 |
|  |  |  |  |  | Chilkat Mainstem | 0.004 | 0.007 | 0.000 | 0.019 |
|  |  |  |  |  | Chilkoot | 0.370 | 0.034 | 0.315 | 0.428 |
|  |  |  |  |  | Other | 0.298 | 0.034 | 0.242 | 0.355 |
| All | 3,803 | 1,880 | 1,661 | 262 | Chilkat Lake | 0.169 | 0.011 | 0.152 | 0.189 |
|  |  |  |  |  | Chilkat Mainstem | 0.048 | 0.007 | 0.038 | 0.059 |
|  |  |  |  |  | Chilkoot | 0.619 | 0.013 | 0.597 | 0.641 |
|  |  |  |  |  | Other | 0.163 | 0.010 | 0.146 | 0.180 |

Note: Gray highlighted rows indicate the GSI estimates did not meet acceptable levels of precision and accuracy to estimate the proportion of mixtures within $10 \%$ of the true mixture $90 \%$ of the time.

Appendix M.-Annual Chilkoot Lake sockeye salmon escapements based on weir counts, and estimated harvests (commercial, sport, and subsistence), total runs (harvest plus escapement), and harvest rates, 1976-2019.

| Year | Escapement goal |  | Escapement estimate | Harvest |  |  |  | Total run | Harvest rate (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Lower | Upper |  | Commercial | Sport | Subsistence | Total |  |  |
| 1976 | 80,000 | 100,000 | 71,291 | 61,861 | ND | ND | 61,861 | 133,152 | 46\% |
| 1977 | 80,000 | 100,000 | 97,368 | 113,555 | 400 | ND | 113,955 | 211,323 | 54\% |
| 1978 | 80,000 | 100,000 | 35,454 | 14,264 | 500 | ND | 14,764 | 50,218 | 29\% |
| 1979 | 80,000 | 100,000 | 96,122 | 69,864 | 300 | ND | 70,164 | 166,286 | 42\% |
| 1980 | 80,000 | 100,000 | 98,673 | 21,244 | 700 | ND | 21,944 | 120,617 | 18\% |
| 1981 | 60,000 | 80,000 | 84,047 | 43,756 | 1,200 | ND | 44,956 | 129,003 | 35\% |
| 1982 | 60,000 | 80,000 | 103,038 | 144,748 | 800 | ND | 145,548 | 248,586 | 59\% |
| 1983 | 60,000 | 80,000 | 80,141 | 242,034 | 600 | ND | 242,634 | 322,775 | 75\% |
| 1984 | 60,000 | 80,000 | 100,781 | 225,634 | 1,000 | ND | 226,634 | 327,415 | 69\% |
| 1985 | 60,000 | 80,000 | 69,141 | 153,533 | 1,100 | 1,055 | 155,688 | 224,829 | 69\% |
| 1986 | 60,000 | 80,000 | 88,024 | 110,114 | 3,000 | 1,640 | 114,754 | 202,778 | 57\% |
| 1987 | 60,000 | 80,000 | 94,208 | 327,323 | 1,700 | 1,237 | 330,260 | 424,468 | 78\% |
| 1988 | 60,000 | 80,000 | 81,274 | 248,640 | 300 | 1,013 | 249,953 | 331,227 | 75\% |
| 1989 | 60,000 | 80,000 | 54,900 | 292,830 | 900 | 2,055 | 295,785 | 350,685 | 84\% |
| 1990 | 50,500 | 91,500 | 76,119 | 181,260 | 2,600 | 2,391 | 186,251 | 262,370 | 71\% |
| 1991 | 50,500 | 91,500 | 92,375 | 228,607 | 600 | 4,399 | 233,606 | 325,981 | 72\% |
| 1992 | 50,500 | 91,500 | 77,601 | 142,471 | 500 | 4,104 | 147,075 | 224,676 | 65\% |
| 1993 | 50,500 | 91,500 | 52,080 | 52,080 | 100 | 2,896 | 55,076 | 107,156 | 51\% |
| 1994 | 50,500 | 91,500 | 37,007 | 25,367 | 400 | 1,589 | 27,356 | 64,363 | 43\% |
| 1995 | 50,500 | 91,500 | 7,177 | 9,637 | 200 | 384 | 10,221 | 17,398 | 59\% |
| 1996 | 50,500 | 91,500 | 50,741 | 19,882 | 475 | 2,311 | 22,668 | 73,409 | 31\% |
| 1997 | 50,500 | 91,500 | 44,254 | 31,822 | 478 | 1,781 | 34,081 | 78,335 | 44\% |
| 1998 | 50,500 | 91,500 | 12,335 | 2,838 | closed | 160 | 2,998 | 15,333 | 20\% |
| 1999 | 50,500 | 91,500 | 19,284 | 4,604 | 27 | 115 | 4,746 | 24,030 | 20\% |
| 2000 | 50,500 | 91,500 | 43,555 | 14,622 | 384 | 251 | 15,257 | 58,812 | 26\% |
| 2001 | 50,500 | 91,500 | 76,283 | 66,355 | 2,344 | 1,499 | 70,198 | 146,481 | 48\% |
| 2002 | 50,500 | 91,500 | 58,361 | 24,200 | 1,503 | 1,258 | 26,961 | 85,322 | 32\% |
| 2003 | 50,500 | 91,500 | 75,065 | 32,446 | 1,509 | 2,091 | 36,046 | 111,111 | 32\% |
| 2004 | 50,500 | 91,500 | 77,660 | 66,498 | 889 | 1,766 | 69,153 | 146,813 | 47\% |
| 2005 | 50,500 | 91,500 | 51,178 | 29,276 | 566 | 1,427 | 31,269 | 82,447 | 38\% |
| 2006 | 50,000 | 90,000 | 96,203 | 119,201 | 520 | 2,279 | 122,000 | 218,203 | 56\% |
| 2007 | 50,000 | 90,000 | 72,678 | 125,199 | 303 | 3,290 | 128,792 | 201,470 | 64\% |
| 2008 | 50,000 | 90,000 | 33,117 | 7,491 | 298 | 1,894 | 9,683 | 42,800 | 23\% |
| 2009 | 38,000 | 86,000 | 33,705 | 16,622 | 165 | 892 | 17,679 | 51,384 | 34\% |
| 2010 | 38,000 | 86,000 | 71,657 | 32,064 | 567 | 2,251 | 34,882 | 106,539 | 33\% |
| 2011 | 38,000 | 86,000 | 65,915 | 26,766 | 973 | 1,976 | 29,715 | 95,630 | 31\% |
| 2012 | 38,000 | 86,000 | 118,166 | 124,366 | 1,025 | 3,080 | 128,471 | 246,637 | 52\% |
| 2013 | 38,000 | 86,000 | 46,329 | 23,111 | 204 | 2,439 | 25,754 | 72,083 | 36\% |
| 2014 | 38,000 | 86,000 | 105,713 | 110,487 | 318 | 3,231 | 114,036 | 219,749 | 52\% |
| 2015 | 38,000 | 86,000 | 71,515 | 58,568 | 912 | 2,222 | 61,702 | 133,217 | 46\% |
| 2016 | 38,000 | 86,000 | 86,721 | 119,843 | 215 | 5,051 | 125,109 | 211,830 | 59\% |

-continued-

Appendix M.-Page 2 of 2.

| Year | Escapement goal |  | Escapementestimate | Harvest |  |  |  | Total run | Harvest rate (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Lower | Upper |  | Commercial | Sport | Subsistence | Total |  |  |
| 2017 | 38,000 | 86,000 | 43,098 | 1,933 | 233 | 2,102 | 4,268 | 47,366 | 9\% |
| 2018 | 38,000 | 86,000 | 85,463 | 33,969 | 159 | 4,406 | 38,534 | 123,997 | 31\% |
| 2019 | 38,000 | 86,000 | 140,378 | 149,586 | 86 | 3,673 | 153,345 | 293,723 | 52\% |
| 1976-2015 Average |  |  | 68,013 | 91,131 | 799 | 1,902 | 93,364 | 161,378 | 48\% |
| 1976-2015 Median |  |  | 72,168 | 60,215 | 567 | 1,894 | 61,782 | 133,185 | 47\% |
| 1976-2015 Lower Quartile |  |  | 49,638 | 23,928 | 335 | 1,248 | 26,659 | 77,104 | 33\% |
| 1976-2015 Upper Quartile |  |  | 89,112 | 129,517 | 993 | 2,351 | 132,981 | 224,714 | 60\% |

Note: Bold estimates are historical records that have been updated since the last project report by Bednarski et al. (2016).

Appendix N.-Historical age composition of the Chilkoot Lake sockeye salmon escapement, weighted by statistical week, 1982-2019.

| Year | Weighted by stat. week | Age class |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.1 | 1.1 | 2.1 | 3.1 | 0.2 | 1.2 | 2.2 | 3.2 | 0.3 | 1.3 | 2.3 | 3.3 | 1.4 | 2.4 |  |
| 1982 | Escapement by age class | 66 | 0 | 65 | 0 | 0 | 19,342 | 560 | 0 | 139 | 80,980 | 914 | 0 | 972 | 0 | 103,038 |
|  | SE of number | 65 | 0 | 65 | 0 | 0 | 938 | 185 | 0 | 98 | 989 | 244 | 0 | 243 | 0 |  |
|  | Proportion by age class | 0.1\% | 0.0\% | 0.1\% | 0.0\% | 0.0\% | 18.8\% | 0.5\% | 0.0\% | 0.1\% | 78.6\% | 0.9\% | 0.0\% | 0.9\% | 0.0\% |  |
|  | SE of \% | 0.1\% | 0.0\% | 0.1\% | 0.0\% | 0.0\% | 0.9\% | 0.2\% | 0.0\% | 0.1\% | 1.0\% | 0.2\% | 0.0\% | 0.2\% | 0.0\% |  |
|  | Sample size | 1 | 0 | 1 | 0 | 0 | 320 | 9 | 0 | 2 | 1,322 | 16 | 0 | 16 | 0 | 1,687 |
| 1983 | Escapement by age class | 0 | 84 | 42 | 0 | 0 | 9,852 | 1,352 | 0 | 95 | 48,435 | 20,043 | 0 | 238 | 0 | 80,141 |
|  | SE of number | 0 | 59 | 42 | 0 | 0 | 637 | 279 | 0 | 69 | 972 | 837 | 0 | 118 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.1\% | 0.1\% | 0.0\% | 0.0\% | 12.3\% | 1.7\% | 0.0\% | 0.1\% | 60.4\% | 25.0\% | 0.0\% | 0.3\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.1\% | 0.1\% | 0.0\% | 0.0\% | 0.8\% | 0.3\% | 0.0\% | 0.1\% | 1.2\% | 1.0\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 2 | 1 | 0 | 0 | 214 | 25 | 0 | 2 | 1,081 | 461 | 0 | 4 | 0 | 1,790 |
| 1984 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 4,712 | 345 | 0 | 0 | 86,112 | 8,635 | 0 | 977 | 0 | 100,781 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 525 | 132 | 0 | 0 | 921 | 751 | 0 | 279 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 4.7\% | 0.3\% | 0.0\% | 0.0\% | 85.4\% | 8.6\% | 0.0\% | 1.0\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.5\% | 0.1\% | 0.0\% | 0.0\% | 0.9\% | 0.7\% | 0.0\% | 0.3\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 86 | 7 | 0 | 0 | 1,649 | 145 | 0 | 15 | 0 | 1,902 |
| 1985 | Escapement by age class | 0 | 46 | 0 | 0 | 0 | 8,132 | 1,661 | 45 | 0 | 45,675 | 11,517 | 0 | 1,857 | 208 | 69,141 |
|  | SE of number | 0 | 46 | 0 | 0 | 0 | 552 | 252 | 45 | 0 | 876 | 700 | 0 | 342 | 93 |  |
|  | Proportion by age class | 0.0\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 11.8\% | 2.4\% | 0.1\% | 0.0\% | 66.1\% | 16.7\% | 0.0\% | 2.7\% | 0.3\% |  |
|  | SE of \% | 0.0\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 0.8\% | 0.4\% | 0.1\% | 0.0\% | 1.3\% | 1.0\% | 0.0\% | 0.5\% | 0.1\% |  |
|  | Sample size | 0 | 1 | 0 | 0 | 0 | 198 | 43 | 1 | 0 | 1,078 | 258 | 0 | 39 | 5 | 1,623 |
| 1986 | Escapement by age class | 0 | 43 | 0 | 0 | 0 | 11,398 | 1,934 | 0 | 0 | 59,561 | 14,425 | 67 | 493 | 102 | 88,024 |
|  | SE of number | 0 | 42 | 0 | 0 | 0 | 627 | 289 | 0 | 0 | 906 | 718 | 67 | 144 | 59 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 12.9\% | 2.2\% | 0.0\% | 0.0\% | 67.7\% | 16.4\% | 0.1\% | 0.6\% | 0.1\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.7\% | 0.3\% | 0.0\% | 0.0\% | 1.0\% | 0.8\% | 0.1\% | 0.2\% | 0.1\% |  |
|  | Sample size | 0 | 1 | 0 | 0 | 0 | 284 | 47 | 0 | 0 | 1,438 | 361 | 1 | 12 | 3 | 2,147 |
| 1987 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 7,706 | 2,074 | 0 | 0 | 62,153 | 21,773 | 79 | 283 | 139 | 94,208 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 537 | 294 | 0 | 0 | 915 | 811 | 79 | 132 | 80 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 8.2\% | 2.2\% | 0.0\% | 0.0\% | 66.0\% | 23.1\% | 0.1\% | 0.3\% | 0.1\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.6\% | 0.3\% | 0.0\% | 0.0\% | 1.0\% | 0.9\% | 0.1\% | 0.1\% | 0.1\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 185 | 49 | 0 | 0 | 1,527 | 437 | 1 | 5 | 3 | 2,207 |

Appendix N.-Page 2 of 7.

| Year | Weighted by stat. week | Age class |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.1 | 1.1 | 2.1 | 3.1 | 0.2 | 1.2 | 2.2 | 3.2 | 0.3 | 1.3 | 2.3 | 3.3 | 1.4 | 2.4 |  |
| $1988$ | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 3,265 | 2,103 | 0 | 0 | 63,381 | 11,060 | 52 | 1,115 | 299 | 81,274 |
|  | SE of number | $0$ | $0$ | $0$ | 0 | 0 | 317 | 263 | 0 | 0 | 705 | 592 | 51 | 196 | 107 |  |
|  | Proportion by age class | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $4.0 \%$ | $2.6 \%$ | $0.0 \%$ | $0.0 \%$ | $78.0 \%$ | 13.6\% | $0.1 \%$ | $1.4 \%$ | $0.4 \%$ |  |
|  | SE of \% | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.4 \%$ | $0.3 \%$ | 0.0\% | $0.0 \%$ | 0.9\% | 0.7\% | 0.1\% | 0.2\% | 0.1\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 117 | 72 | 0 | 0 | $2,074$ | 350 | 1 | 38 | 9 | 2,661 |
| 1989 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 1,743 | 2,169 | 0 | 0 | 30,584 | 19,213 | 304 | 649 | 238 | 54,900 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 178 | 226 | 0 | 0 | 680 | 657 | 102 | 146 | 96 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 3.2\% | 4.0\% | 0.0\% | 0.0\% | 55.7\% | 35.0\% | 0.6\% | 1.2\% | 0.4\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.3\% | 0.4\% | 0.0\% | 0.0\% | 1.2\% | 1.2\% | 0.2\% | 0.3\% | 0.2\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 116 | 130 | 0 | 0 | 1,419 | 866 | 14 | 31 | 10 | 2,586 |
| $1990$ | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 1,227 | 1,006 | 11 | 0 | 35,537 | 36,830 | 64 | 736 | 708 | 76,119 |
|  | SE of number | $0$ | $0$ | $0$ | 0 | 0 | $185$ | $180$ | 10 | 0 | $806$ | 807 | 46 | 161 | $150$ |  |
|  | Proportion by age class | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $1.6 \%$ | $1.3 \%$ | $0.0 \%$ | $0.0 \%$ | $46.7 \%$ | 48.4\% | $0.1 \%$ | $1.0 \%$ | $0.9 \%$ |  |
|  | SE of \% | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.2 \%$ | $0.2 \%$ | $0.0 \%$ | $0.0 \%$ | 1.1\% | 1.1\% | 0.1\% | $0.2 \%$ | $0.2 \%$ |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 55 | 41 | 1 | 0 | 1,277 | 1,382 | 3 | 27 | 29 | 2,815 |
| 1991 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 12,537 | 4,648 | 0 | 0 | 50,513 | 24,249 | 100 | 158 | 169 | 92,375 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 870 | 538 | 0 | 0 | 1,236 | 1,104 | 62 | 53 | 74 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 13.6\% | 5.0\% | 0.0\% | 0.0\% | 54.7\% | 26.3\% | 0.1\% | 0.2\% | 0.2\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.9\% | 0.6\% | 0.0\% | 0.0\% | 1.3\% | 1.2\% | 0.1\% | 0.1\% | 0.1\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 287 | 112 | 0 | 0 | 1,283 | 596 | 3 | 9 | 7 | 2,297 |
| 1992 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 1,824 | 4,028 | 56 | 17 | 52,400 | 18,410 | 105 | 419 | 342 | 77,601 |
|  | SE of number | $0$ | $0$ | $0$ | $0$ | $0$ | $448$ | $428$ | $31$ | $16$ | $894$ | $765$ | $64$ | $119$ | $115$ |  |
|  | Proportion by age class | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $2.4 \%$ | $5.2 \%$ | $0.1 \%$ | $0.0 \%$ | $67.5 \%$ | $23.7 \%$ | $0.1 \%$ | $0.5 \%$ | $0.4 \%$ |  |
|  | SE of \% | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.6 \%$ | 0.6\% | 0.0\% | 0.0\% | 1.2\% | 1.0\% | $0.1 \%$ | 0.2\% | 0.1\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 36 | 118 | 3 | 1 | 1,277 | 577 | 3 | 14 | 10 | 2,039 |
| 1993 | Escapement by age class | 0 | 0 | 0 | 19 | 0 | 1,560 | 901 | 0 | 0 | 18,693 | 30,396 | 91 | 180 | 239 | 52,080 |
|  | SE of number | 0 | 0 | 0 | 18 | 0 | 207 | 149 | 0 | 0 | 541 | 560 | 43 | 76 | 84 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 3.0\% | 1.7\% | 0.0\% | 0.0\% | 35.9\% | 58.4\% | 0.2\% | 0.3\% | 0.5\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.4\% | 0.3\% | 0.0\% | 0.0\% | 1.0\% | 1.1\% | 0.1\% | 0.1\% | 0.2\% |  |
|  | Sample size | 0 | 0 | 0 | 1 | 0 | 54 | 37 | 0 | 0 | 739 | 1,224 | 5 | 6 | 9 | 2,075 |

-continued-

Appendix N.-Page 3 of 7 .

| Year | Weighted by stat. week | Age class |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.1 | 1.1 | 2.1 | 3.1 | 0.2 | 1.2 | 2.2 | 3.2 | 0.3 | 1.3 | 2.3 | 3.3 | 1.4 | 2.4 |  |
| 1994 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 671 | 549 | 23 | 48 | 24,876 | 10,573 | 22 | 194 | 50 | 37,007 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 112 | 98 | 23 | 34 | 392 | 378 | 21 | 56 | 24 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.8\% | 1.5\% | 0.1\% | 0.1\% | 67.2\% | 28.6\% | 0.1\% | 0.5\% | 0.1\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.3\% | 0.3\% | 0.1\% | 0.1\% | 1.1\% | 1.0\% | 0.1\% | 0.2\% | 0.1\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 35 | 32 | 1 | 2 | 1,328 | 571 | 1 | 12 | 4 | 1,986 |
| 1995 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 3,360 | 298 | 0 | 0 | 2,176 | 1,219 | 0 | 78 | 46 | 7,177 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 129 | 67 | 0 | 0 | 139 | 114 | 0 | 40 | 27 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 46.8\% | 4.2\% | 0.0\% | 0.0\% | 30.3\% | 17.0\% | 0.0\% | 1.1\% | 0.6\% |  |
|  | SE of \% | $0.0 \%$ | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.8\% | 0.9\% | 0.0\% | 0.0\% | 1.9\% | 1.6\% | 0.0\% | 0.6\% | 0.4\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 267 | 23 | 0 | 0 | 186 | 121 | 0 | 5 | 4 | 606 |
| 1996 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 3,365 | 517 | 23 | 11 | 43,232 | 3,559 | 0 | 35 | 0 | 50,741 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 338 | 145 | 22 | 10 | 461 | 308 | 0 | 18 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 6.6\% | 1.0\% | 0.0\% | 0.0\% | 85.2\% | 7.0\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.7\% | 0.3\% | 0.0\% | 0.0\% | 0.9\% | 0.6\% | 0.0\% | 0.0\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 128 | 16 | 1 | 1 | 1,737 | 176 | 0 | 4 | 0 | 2,063 |
| 1997 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 1,022 | 183 | 0 | 23 | 39,858 | 3,114 | 8 | 45 | 0 | 44,254 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 146 | 65 | 0 | 23 | 286 | 244 | 8 | 31 | 0 |  |
|  | Proportion by age class | $0.0 \%$ | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 2.3\% | 0.4\% | 0.0\% | 0.1\% | 90.1\% | 7.0\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | SE of \% | $0.0 \%$ | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.3\% | 0.1\% | 0.0\% | 0.1\% | 0.6\% | 0.6\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 47 | 8 | 0 | 1 | 1,902 | 150 | 1 | 2 | 0 | 2,111 |
| 1998 | Escapement by age class | 15 | 0 | 0 | 0 | 0 | 631 | 268 | 0 | 0 | 7,478 | 3,753 | 13 | 165 | 13 | 12,335 |
|  | SE of number | 15 | 0 | 0 | 0 | 0 | 86 | 57 | 0 | 0 | 189 | 177 | 13 | 44 | 13 |  |
|  | Proportion by age class | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 5.1\% | 2.2\% | 0.0\% | 0.0\% | 60.6\% | 30.4\% | 0.1\% | 1.3\% | 0.1\% |  |
|  | SE of \% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.7\% | 0.5\% | 0.0\% | 0.0\% | 1.5\% | 1.4\% | 0.1\% | 0.4\% | 0.1\% |  |
|  | Sample size | 1 | 0 | 0 | 0 | 0 | 47 | 20 | 0 | 0 | 570 | 288 | 1 | 13 | 1 | 941 |
| 1999 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 5,934 | 1,597 | 0 | 0 | 8,550 | 3,136 | 0 | 34 | 34 | 19,284 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 203 | 124 | 0 | 0 | 212 | 163 | 0 | 16 | 18 |  |
|  | Proportion by age class | $0.0 \%$ | 0.0\% | $0.0 \%$ | 0.0\% | 0.0\% | 30.8\% | 8.3\% | 0.0\% | 0.0\% | 44.3\% | 16.3\% | 0.0\% | 0.2\% | 0.2\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.1\% | 0.6\% | 0.0\% | 0.0\% | 1.1\% | 0.8\% | 0.0\% | 0.1\% | 0.1\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 585 | 164 | 0 | 0 | 945 | 331 | 0 | 4 | 4 | 2,033 |

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| Year | Weighted by stat. week | Age class |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.1 | 1.1 | 2.1 | 3.1 | 0.2 | 1.2 | 2.2 | 3.2 | 0.3 | 1.3 | 2.3 | 3.3 | 1.4 | 2.4 |  |
| 2000 | Escapement by age class | 0 | 0 | 0 | 0 | 24 | 6,678 | 1,041 | 0 | 0 | 25,864 | 9,903 | 0 | 29 | 15 | 43,555 |
|  | SE of number | 0 | 0 | 0 | 0 | 24 | 359 | 160 | 0 | 0 | 468 | 377 | 0 | 20 | 15 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.1\% | 15.3\% | 2.4\% | 0.0\% | 0.0\% | 59.4\% | 22.7\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.1\% | 0.8\% | 0.4\% | 0.0\% | 0.0\% | 1.1\% | 0.9\% | 0.0\% | 0.0\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 1 | 295 | 42 | 0 | 0 | 1,306 | 581 | 0 | 2 | 1 | 2,228 |
| 2001 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 3,565 | 50 | 0 | 157 | 68,859 | 3,600 | 0 | 53 | 0 | 76,283 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 436 | 29 | 0 | 62 | 606 | 437 | 0 | 52 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 4.7\% | 0.1\% | 0.0\% | 0.2\% | 90.3\% | 4.7\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | SE of \% | $0.0 \%$ | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.6\% | 0.0\% | 0.0\% | 0.1\% | 0.8\% | 0.6\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 113 | 4 | 0 | 7 | 2,106 | 114 | 0 | 1 | 0 | 2,345 |
| $2002$ | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 4,989 | 800 | 0 | 0 | 50,880 | 1,400 | 0 | 292 | 0 | 58,361 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 382 | 155 | 0 | 0 | 441 | 181 | 0 | 85 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 8.5\% | 1.4\% | 0.0\% | 0.0\% | 87.2\% | 2.4\% | 0.0\% | 0.5\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.7\% | 0.3\% | 0.0\% | 0.0\% | 0.8\% | 0.3\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 182 | 30 | 0 | 0 | 2,540 | 71 | 0 | 13 | 0 | 2,836 |
| 2003 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 42,648 | 2,594 | 0 | 0 | 24,883 | 4,776 | 0 | 132 | 33 | 75,065 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 960 | 326 | 0 | 0 | 905 | 458 | 0 | 60 | 32 |  |
|  | Proportion by age class | $0.0 \%$ | $0.0 \%$ | 0.0\% | $0.0 \%$ | 0.0\% | 56.8\% | 3.5\% | 0.0\% | 0.0\% | 33.1\% | 6.4\% | 0.0\% | 0.2\% | 0.0\% |  |
|  | SE of \% | $0.0 \%$ | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.3\% | 0.4\% | 0.0\% | 0.0\% | 1.2\% | 0.6\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 1,078 | 110 | 0 | 0 | 1,174 | 238 | 0 | 10 | 1 | 2,611 |
| 2004 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 11,846 | 5,738 | 0 | 0 | 54,309 | 5,732 | 0 | 36 | 0 | 77,660 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 611 | 460 | 0 | 0 | 770 | 414 | 0 | 25 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 15.3\% | 7.4\% | 0.0\% | 0.0\% | 69.9\% | 7.4\% | 0.0\% | 0.0\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.8\% | 0.6\% | 0.0\% | 0.0\% | 1.0\% | 0.5\% | 0.0\% | 0.0\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 399 | 161 | 0 | 0 | 1,929 | 220 | 0 | 2 | 0 | 2,711 |
| 2005 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 11,048 | 2,242 | 0 | 0 | 32,908 | 4,909 | 0 | 71 | 0 | 51,178 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 433 | 228 | 0 | 0 | 508 | 326 | 0 | 38 | 0 |  |
|  | Proportion by age class | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | 21.6\% | 4.4\% | 0.0\% | 0.0\% | 64.3\% | 9.6\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.8\% | 0.4\% | 0.0\% | 0.0\% | 1.0\% | 0.6\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 542 | 106 | 0 | 0 | 1,843 | 235 | 0 | 4 | 0 | 2,730 |

[^11]Appendix N.-Page 5 of 7.

| Year | Weighted by stat. week | Age class |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.1 | 1.1 | 2.1 | 3.1 | 0.2 | 1.2 | 2.2 | 3.2 | 0.3 | 1.3 | 2.3 | 3.3 | 1.4 | 2.4 |  |
| 2006 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 8,492 | 817 | 0 | 22 | 76,211 | 10,578 | 0 | 48 | 34 | 96,203 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 582 | 187 | 0 | 21 | 839 | 653 | 0 | 48 | 34 |  |
|  | Proportion by age class | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ | $8.8 \%$ | $0.8 \%$ | $0.0 \%$ | $0.0 \%$ | $79.2 \%$ | $11.0 \%$ | $0.0 \%$ | $0.1 \%$ | $0.0 \%$ |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.6\% | 0.2\% | 0.0\% | 0.0\% | 0.9\% | 0.7\% | 0.0\% | 0.0\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 211 | 22 | 0 | 1 | 2,076 | 269 | 0 | 1 | 1 | 2,581 |
| 2007 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 7,128 | 618 | 0 | 0 | 55,604 | 8,908 | 0 | 421 | 0 | 72,678 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 483 | 150 | 0 | 0 | 658 | 493 | 0 | 116 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 9.8\% | 0.8\% | 0.0\% | 0.0\% | 76.5\% | 12.3\% | 0.0\% | 0.6\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.7\% | 0.2\% | 0.0\% | 0.0\% | 0.9\% | 0.7\% | 0.0\% | 0.2\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 214 | 19 | 0 | 0 | 2,387 | 383 | 0 | 17 | 0 | 3,020 |
| 2008 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 3,405 | 330 | 0 | 55 | 26,672 | 1,403 | 0 | 1,213 | 39 | 33,117 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 427 | 154 | 0 | 31 | 552 | 282 | 0 | 255 | 23 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 10.3\% | 1.0\% | 0.0\% | 0.2\% | 80.5\% | 4.2\% | 0.0\% | 3.7\% | 0.1\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.3\% | 0.5\% | 0.0\% | 0.1\% | 1.7\% | 0.9\% | 0.0\% | 0.8\% | 0.1\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 103 | 6 | 0 | 3 | 851 | 44 | 0 | 47 | 3 | 1,057 |
| 2009 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 9,539 | 647 | 0 | 0 | 22,801 | 615 | 0 | 103 | 0 | 33,705 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 386 | 119 | 0 | 0 | 399 | 115 | 0 | 45 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 28.3\% | 1.9\% | 0.0\% | 0.0\% | 67.6\% | 1.8\% | 0.0\% | 0.3\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.1\% | 0.4\% | 0.0\% | 0.0\% | 1.2\% | 0.3\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 479 | 35 | 0 | 0 | 1,288 | 34 | 0 | 5 | 0 | 1,841 |
| 2010 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 4,269 | 2,922 | 34 | 0 | 58,284 | 6,099 | 0 | 48 | 0 | 71,657 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 554 | 466 | 25 | 0 | 883 | 619 | 0 | 30 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 6.0\% | 4.1\% | 0.0\% | 0.0\% | 81.3\% | 8.5\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.8\% | 0.6\% | 0.0\% | 0.0\% | 1.2\% | 0.9\% | 0.0\% | 0.0\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 122 | 72 | 3 | 0 | 2,070 | 223 | 0 | 3 | 0 | 2,493 |
| 2011 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 20,450 | 1,421 | 0 | 4 | 32,475 | 11,301 | 136 | 120 | 8 | 65,915 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 786 | 253 | 0 | 4 | 829 | 635 | 64 | 66 | 7 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 31.0\% | 2.2\% | 0.0\% | 0.0\% | 49.3\% | 17.1\% | 0.2\% | 0.2\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.2\% | 0.4\% | 0.0\% | 0.0\% | 1.3\% | 1.0\% | 0.1\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 637 | 50 | 0 | 1 | 1,441 | 431 | 7 | 4 | 1 | 2,572 |

[^12]Appendix N.-Page 6 of 7.

| Year | Weighted by stat. week | Age class |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.1 | 1.1 | 2.1 | 3.1 | 0.2 | 1.2 | 2.2 | 3.2 | 0.3 | 1.3 | 2.3 | 3.3 | 1.4 | 2.4 |  |
| 2012 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 2,730 | 449 | 0 | 0 | 102,954 | 11,803 | 0 | 230 | 0 | 118,166 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 473 | 157 | 0 | 0 | 1,116 | 1,024 | 0 | 86 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 2.3\% | 0.4\% | 0.0\% | 0.0\% | 87.1\% | 10.0\% | 0.0\% | 0.2\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.4\% | 0.1\% | 0.0\% | 0.0\% | 0.9\% | 0.9\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 76 | 18 | 0 | 0 | 2,078 | 240 | 0 | 11 | 0 | 2,423 |
| 2013 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 13,574 | 2,826 | 0 | 0 | 22,516 | 5,930 | 93 | 1,390 | 46 | 46,329 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 800 | 445 | 0 | 0 | 876 | 566 | 102 | 261 | 59 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 29.3\% | 6.1\% | 0.0\% | 0.0\% | 48.6\% | 12.8\% | 0.2\% | 3.0\% | 0.1\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.7\% | 1.0\% | 0.0\% | 0.0\% | 1.9\% | 1.2\% | 0.2\% | 0.6\% | 0.1\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 452 | 71 | 0 | 0 | 826 | 208 | 1 | 58 | 1 | 1,617 |
| 2014 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 28,648 | 5,920 | 0 | 0 | 64,274 | 6,766 | 0 | 106 | 0 | 105,713 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 1,314 | 677 | 0 | 0 | 1,403 | 678 | 0 | 54 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 27.1\% | 5.6\% | 0.0\% | 0.0\% | 60.8\% | 6.4\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.2\% | 0.6\% | 0.0\% | 0.0\% | 1.3\% | 0.6\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 421 | 101 | 0 | 0 | 1,503 | 150 | 0 | 5 | 0 | 2,181 |
| 2015 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 11,156 | 1,502 | 0 | 0 | 54,280 | 4,434 | 0 | 215 | 0 | 71,515 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 749 | 301 | 0 | 9 | 885 | 503 | 0 | 105 | 6 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 15.6\% | 2.1\% | 0.0\% | 0.0\% | 75.9\% | 6.2\% | 0.0\% | 0.3\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.1\% | 0.4\% | 0.0\% | 0.0\% | 1.2\% | 0.7\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 211 | 28 | 0 | 1 | 1,253 | 100 | 0 | 3 | 1 | 1,597 |
| 2016 | Escapement by age class | 0 | 5 | 0 | 0 | 0 | 2,186 | 362 | 0 | 0 | 73,061 | 11,024 | 9 | 73 | 0 | 86,721 |
|  | SE of number | 0 | 5 | 0 | 0 | 0 | 521 | 133 | 0 | 0 | 1,214 | 1,126 | 8 | 52 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 2.5\% | 0.4\% | 0.0\% | 0.0\% | 84.2\% | 12.7\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.6\% | 0.2\% | 0.0\% | 0.0\% | 1.4\% | 1.3\% | 0.0\% | 0.1\% | 0.0\% |  |
|  | Sample size | 0 | 1 | 0 | 0 | 0 | 33 | 9 | 0 | 0 | 1,376 | 207 | 1 | 2 | 0 | 1,629 |
| 2017 | Escapement by age class | 117 | 0 | 0 | 0 | 0 | 8,702 | 799 | 0 | 55 | 29,286 | 3,265 | 0 | 737 | 137 | 43,098 |
|  | SE of number | 116 | 0 | 0 | 0 | 0 | 867 | 328 | 0 | 38 | 1,050 | 644 | 0 | 202 | 92 |  |
|  | Proportion by age class | 0.3\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 20.2\% | 1.9\% | 0.0\% | 0.1\% | 68.0\% | 7.6\% | 0.0\% | 1.7\% | 0.3\% |  |
|  | SE of \% | 0.3\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 2.0\% | 0.8\% | 0.0\% | 0.1\% | 2.4\% | 1.5\% | 0.0\% | 0.5\% | 0.2\% |  |
|  | Sample size | 1 | 0 | 0 | 0 | 0 | 124 | 10 | 0 | 2 | 504 | 43 | 0 | 18 | 3 | 705 |

[^13]Appendix N.-Page 7 of 7.

| Year | Weighted by stat. week | Age class |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.1 | 1.1 | 2.1 | 3.1 | 0.2 | 1.2 | 2.2 | 3.2 | 0.3 | 1.3 | 2.3 | 3.3 | 1.4 | 2.4 |  |
| 2018 | Escapement by age class | 0 | 128 | 0 | 0 | 0 | 40,331 | 0 | 0 | 24 | 40,570 | 3,581 | 0 | 819 | 9 | 85,463 |
|  | SE of number | 0 | 90 | 0 | 0 | 0 | 2,885 | 0 | 0 | 24 | 2,857 | 1,198 | 0 | 673 | 9 |  |
|  | Proportion by age class | 0.0\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 47.2\% | 0.0\% | 0.0\% | 0.0\% | 47.5\% | 4.2\% | 0.0\% | 1.0\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.1\% | 0.0\% | 0.0\% | 0.0\% | 3.4\% | 0.0\% | 0.0\% | 0.0\% | 3.3\% | 1.4\% | 0.0\% | 0.8\% | 0.0\% |  |
|  | Sample size | 0 | 2 | 0 | 0 | 0 | 205 | 0 | 0 | 1 | 442 | 28 | 0 | 7 | 1 | 686 |
| 2019 | Escapement by age class | 0 | 0 | 0 | 0 | 0 | 23,987 | 557 | 0 | 0 | 113,393 | 2,034 | 0 | 407 | 0 | 140,378 |
|  | SE of number | 0 | 0 | 0 | 0 | 0 | 3,141 | 295 | 0 | 0 | 3,252 | 966 | 0 | 392 | 0 |  |
|  | Proportion by age class | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 17.1\% | 0.4\% | 0.0\% | 0.0\% | 80.8\% | 1.4\% | 0.0\% | 0.3\% | 0.0\% |  |
|  | SE of \% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 2.2\% | 0.2\% | 0.0\% | 0.0\% | 2.3\% | 0.7\% | 0.0\% | 0.3\% | 0.0\% |  |
|  | Sample size | 0 | 0 | 0 | 0 | 0 | 92 | 4 | 0 | 0 | 700 | 13 | 0 | 2 | 0 | 811 |

Appendix O.-Average length (mid eye to tail fork in mm), standard error (SE), and number of samples ( $n$ ) of male sockeye salmon in the Chilkoot Lake escapement by major age class, 1982-2019. (Dashes indicate age class not present.)

| Year | Age 1.2 |  |  | Age 1.3 |  |  | Age 2.2 |  |  | Age 2.3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Avg. | SE | $n$ | Avg. | SE | $n$ | Avg. | SE | $n$ | Avg. | SE | $n$ |
| 1982 | 469 | 4.0 | 143 | 591 | 1.1 | 675 | 538 | 17.5 | 2 | 594 | 2.9 | 11 |
| 1983 | 456 | 2.9 | 132 | 581 | 1.0 | 523 | 479 | 22.0 | 8 | 580 | 1.7 | 189 |
| 1984 | 455 | 4.1 | 73 | 581 | 0.9 | 850 | 457 | 8.7 | 5 | 580 | 2.2 | 77 |
| 1985 | 469 | 2.6 | 182 | 578 | 1.1 | 598 | 472 | 5.4 | 36 | 577 | 2.0 | 143 |
| 1986 | 470 | 2.6 | 254 | 589 | 1.0 | 810 | 476 | 5.7 | 35 | 590 | 1.8 | 213 |
| 1987 | 469 | 3.1 | 143 | 590 | 1.0 | 813 | 465 | 5.9 | 33 | 591 | 1.5 | 240 |
| 1988 | 496 | 4.9 | 89 | 587 | 0.8 | 1,126 | 500 | 5.2 | 52 | 585 | 1.9 | 176 |
| 1989 | 463 | 3.7 | 89 | 590 | 0.8 | 810 | 474 | 5.0 | 84 | 587 | 1.2 | 451 |
| 1990 | 462 | 6.7 | 40 | 589 | 0.9 | 739 | 487 | 12.4 | 20 | 586 | 1.0 | 776 |
| 1991 | 479 | 3.6 | 161 | 578 | 0.9 | 675 | 476 | 6.3 | 57 | 577 | 1.5 | 316 |
| 1992 | 469 | 9.0 | 28 | 580 | 1.0 | 632 | 460 | 4.3 | 77 | 582 | 1.6 | 268 |
| 1993 | 484 | 7.6 | 49 | 583 | 1.2 | 412 | 507 | 10.6 | 25 | 581 | 1.0 | 641 |
| 1994 | 460 | 9.4 | 27 | 576 | 1.1 | 569 | 478 | 12.5 | 17 | 579 | 1.7 | 250 |
| 1995 | 493 | 2.8 | 179 | 579 | 2.6 | 104 | 501 | 9.6 | 15 | 581 | 2.8 | 69 |
| 1996 | 506 | 4.1 | 87 | 600 | 0.9 | 833 | 514 | 16.4 | 12 | 597 | 3.2 | 77 |
| 1997 | 505 | 5.6 | 36 | 586 | 0.9 | 1,038 | 508 | 9.7 | 8 | 574 | 3.3 | 78 |
| 1998 | 495 | 5.4 | 40 | 579 | 1.5 | 291 | 513 | 9.0 | 16 | 575 | 1.9 | 170 |
| 1999 | 488 | 2.1 | 403 | 588 | 1.1 | 493 | 515 | 4.1 | 101 | 584 | 2.1 | 174 |
| 2000 | 506 | 2.7 | 250 | 589 | 1.1 | 571 | 501 | 9.2 | 36 | 591 | 1.6 | 271 |
| 2001 | 487 | 4.7 | 71 | 588 | 0.8 | 990 | - | - | - | 586 | 4.1 | 44 |
| 2002 | 475 | 3.5 | 142 | 592 | 0.8 | 1,200 | 474 | 7.4 | 19 | 596 | 5.0 | 32 |
| 2003 | 490 | 1.4 | 672 | 586 | 1.1 | 550 | 489 | 4.6 | 65 | 585 | 2.4 | 116 |
| 2004 | 498 | 2.3 | 253 | 580 | 0.9 | 801 | 499 | 4.0 | 96 | 576 | 2.3 | 96 |
| 2005 | 484 | 1.7 | 407 | 574 | 0.8 | 862 | 487 | 4.0 | 80 | 569 | 2.5 | 92 |
| 2006 | 480 | 3.1 | 160 | 569 | 0.8 | 991 | 493 | 13.1 | 14 | 567 | 1.9 | 124 |
| 2007 | 477 | 3.0 | 156 | 577 | 0.7 | 1,133 | 492 | 15.2 | 13 | 576 | 1.8 | 185 |
| 2008 | 489 | 5.4 | 67 | 583 | 1.4 | 350 | 553 | 18.9 | 4 | 583 | 4.8 | 15 |
| 2009 | 485 | 1.9 | 353 | 581 | 1.0 | 660 | 496 | 6.5 | 28 | 583 | 7.1 | 15 |
| 2010 | 480 | 4.0 | 103 | 572 | 0.7 | 887 | 476 | 4.6 | 56 | 567 | 2.5 | 101 |
| 2011 | 492 | 1.8 | 481 | 579 | 0.8 | 811 | 503 | 7.5 | 35 | 577 | 1.5 | 203 |
| 2012 | 493 | 5.7 | 54 | 583 | 0.7 | 1,044 | 508 | 9.7 | 13 | 577 | 1.9 | 124 |
| 2013 | 487 | 2.1 | 329 | 576 | 1.0 | 414 | 494 | 5.2 | 50 | 576 | 2.3 | 99 |
| 2014 | 481 | 1.8 | 347 | 576 | 1.0 | 732 | 486 | 3.9 | 84 | 576 | 3.5 | 64 |
| 2015 | 460 | 3.1 | 175 | 552 | 1.0 | 724 | 460 | 7.9 | 22 | 552 | 2.6 | 60 |
| 2016 | 476 | 6.5 | 31 | 555 | 1.0 | 644 | 474 | 8.6 | 7 | 549 | 3.0 | 97 |
| 2017 | 484 | 3.6 | 105 | 559 | 1.6 | 266 | 484 | 16.8 | 6 | 550 | 6.3 | 23 |
| 2018 | 477 | 2.7 | 176 | 562 | 1.5 | 201 | - | - | - | 576 | 5.3 | 14 |
| 2019 | 479 | 4.9 | 54 | 562 | 1.5 | 296 | 488 | 16.0 | 3 | 537 | 12.9 | 5 |
| Average | 481 |  |  | 579 |  |  | 491 |  |  | 578 |  |  |

Appendix P.-Average length (mid eye to tail fork in mm), standard error (SE), and number of samples ( $n$ ) of female sockeye salmon in the Chilkoot Lake escapement by major age class, 1982-2019. (Dashes indicate age class not present.)

| Year | Age 1.2 |  |  | Age 1.3 |  |  | Age 2.2 |  |  | Age 2.3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Avg. | SE | $n$ | Avg. | SE | $n$ | Avg. | SE | $n$ | Avg. | SE | $n$ |
| 1982 | 465 | 2.8 | 177 | 563 | 1.0 | 646 | 476 | 12.6 | 7 | 562 | 6.8 | 5 |
| 1983 | 455 | 3.8 | 82 | 565 | 0.8 | 558 | 473 | 7.4 | 17 | 560 | 1.4 | 272 |
| 1984 | 497 | 6.9 | 13 | 562 | 0.8 | 798 | 503 | 2.5 | 2 | 559 | 2.8 | 68 |
| 1985 | 507 | 5.7 | 14 | 558 | 0.9 | 480 | 503 | 6.7 | 7 | 552 | 2.0 | 115 |
| 1986 | 491 | 5.7 | 30 | 574 | 0.8 | 627 | 510 | 9.5 | 12 | 570 | 1.7 | 148 |
| 1987 | 473 | 5.4 | 40 | 576 | 0.9 | 714 | 488 | 8.4 | 16 | 573 | 2.0 | 197 |
| 1988 | 497 | 8.7 | 28 | 568 | 0.7 | 946 | 497 | 8.9 | 19 | 564 | 1.8 | 174 |
| 1989 | 486 | 4.3 | 27 | 569 | 0.9 | 608 | 494 | 4.3 | 46 | 565 | 1.2 | 414 |
| 1990 | 483 | 8.3 | 15 | 566 | 1.0 | 538 | 506 | 5.8 | 21 | 567 | 1.0 | 606 |
| 1991 | 485 | 3.2 | 126 | 552 | 1.0 | 606 | 480 | 3.8 | 55 | 553 | 1.6 | 278 |
| 1992 | 481 | 11.8 | 8 | 562 | 0.9 | 644 | 492 | 5.0 | 41 | 563 | 1.4 | 309 |
| 1993 | 525 | 16.0 | 5 | 567 | 1.3 | 323 | 506 | 8.1 | 12 | 565 | 0.9 | 568 |
| 1994 | 511 | 14.4 | 8 | 563 | 0.7 | 759 | 503 | 10.8 | 14 | 561 | 1.2 | 321 |
| 1995 | 505 | 2.5 | 87 | 561 | 2.3 | 82 | 516 | 7.1 | 8 | 563 | 3.4 | 52 |
| 1996 | 519 | 3.5 | 38 | 579 | 0.8 | 884 | 515 | 8.7 | 4 | 577 | 2.6 | 97 |
| 1997 | 526 | 4.6 | 10 | 568 | 0.8 | 861 | - | - | - | 564 | 2.6 | 69 |
| 1998 | 479 | 15.1 | 7 | 565 | 1.3 | 277 | 523 | 8.3 | 3 | 563 | 2.3 | 117 |
| 1999 | 500 | 2.3 | 181 | 569 | 1.2 | 452 | 509 | 4.1 | 62 | 564 | 1.9 | 156 |
| 2000 | 522 | 4.0 | 42 | 578 | 0.8 | 723 | 533 | 8.7 | 6 | 578 | 1.3 | 308 |
| 2001 | 508 | 5.2 | 41 | 576 | 0.6 | 1,097 | 528 | 24.4 | 4 | 566 | 2.7 | 70 |
| 2002 | 496 | 4.4 | 40 | 577 | 0.6 | 1,337 | 498 | 13.8 | 11 | 566 | 4.6 | 39 |
| 2003 | 503 | 1.3 | 383 | 570 | 0.9 | 615 | 508 | 3.5 | 44 | 572 | 1.9 | 118 |
| 2004 | 512 | 1.9 | 146 | 568 | 0.6 | 1,128 | 502 | 3.2 | 65 | 566 | 1.6 | 124 |
| 2005 | 500 | 1.9 | 134 | 561 | 0.7 | 980 | 499 | 4.8 | 26 | 555 | 1.8 | 143 |
| 2006 | 511 | 4.3 | 50 | 554 | 0.6 | 1,084 | 511 | 13.5 | 8 | 555 | 1.6 | 143 |
| 2007 | 504 | 3.6 | 57 | 566 | 0.6 | 1,199 | 521 | 11.6 | 6 | 564 | 1.5 | 196 |
| 2008 | 510 | 4.8 | 36 | 570 | 1.0 | 501 | 510 | 30.0 | 2 | 569 | 3.6 | 29 |
| 2009 | 506 | 2.2 | 126 | 570 | 0.9 | 628 | 511 | 10.3 | 7 | 568 | 5.8 | 19 |
| 2010 | 511 | 5.9 | 19 | 562 | 0.5 | 1,173 | 515 | 4.8 | 16 | 559 | 1.9 | 121 |
| 2011 | 508 | 2.2 | 156 | 567 | 0.8 | 628 | 510 | 7.1 | 15 | 565 | 1.4 | 227 |
| 2012 | 496 | 4.9 | 22 | 563 | 0.7 | 1,007 | 495 | 5.2 | 5 | 556 | 2.0 | 110 |
| 2013 | 505 | 2.2 | 122 | 558 | 1.0 | 412 | 509 | 5.5 | 21 | 558 | 1.8 | 109 |
| 2014 | 509 | 2.2 | 73 | 558 | 0.9 | 770 | 509 | 6.1 | 17 | 560 | 2.7 | 86 |
| 2015 | 476 | 7.3 | 36 | 531 | 0.9 | 527 | 485 | 15.2 | 6 | 536 | 3.0 | 40 |
| 2016 | 478 | 17.5 | 2 | 543 | 0.6 | 636 | 535 | 15.0 | 2 | 543 | 2.0 | 82 |
| 2017 | 496 | 6.7 | 19 | 547 | 1.5 | 238 | 502 | 1.7 | 3 | 538 | 3.5 | 20 |
| 2018 | 490 | 4.7 | 29 | 548 | 1.2 | 241 | - | - | - | 551 | 4.9 | 14 |
| 2019 | 478 | 4.5 | 38 | 544 | 1.2 | 403 | 453 | 0.0 | 1 | 532 | 10.9 | 8 |
| Average | 497 |  |  | 563 |  |  | 503 |  |  | 561 |  |  |

Appendix Q.-Chilkoot Lake zooplankton abundance summary from 1987 to 2019. All stations were averaged and species combined.

|  | Laboratory | Stations | Monthly mean density (no./m²) |  |  |  |  |  |  |  | May-Sep. mean density (no. $/ \mathrm{m}^{2}$ ) | $\begin{gathered} \text { May-Sep. } \\ \text { biomass } \mathrm{mg} / \mathrm{m}^{2} \text { ) } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | location | sampled | Apr. | May | June | July | Aug. | Sep. | Oct. | Nov. |  |  |
| 1987 | Soldotna | 2 | ND | 74,291 | 166,794 | 247,623 | 131,559 | 246,859 | 166,645 | 124,109 | 173,425 | 236 |
| 1988 | Soldotna | 2 | ND | 129,840 | 304,596 | 105,239 | 76,223 | 135,953 | 36,827 | 3,481 | 150,370 | 190 |
| 1989 | Soldotna | 2 | ND | 50,073 | 13,001 | 155,720 | 15,506 | 11,505 | 35,430 | 11,080 | 49,161 | 146 |
| 1990 | Soldotna | 2 | ND | 113,496 | 62,426 | 101,715 | 37,857 | 21,035 | 8,877 | 9,871 | 64,214 | 187 |
| 1991 | Soldotna | 2 | ND | 20,110 | 9,493 | 3,906 | 6,113 | 2,853 | 16,030 | ND | 8,495 | 15 |
| 1995 | Soldotna | 4 | ND | ND | 46,778 | 36,755 | 25,081 | ND | ND | 3,178 | ND | ND |
| 1996 | Soldotna | 4 | ND | 76,537 | 76,728 | 54,180 | 37,528 | 10,103 | 3,354 | ND | 58,119 | 174 |
| 1997 | Soldotna | 4 | ND | 32,320 | 43,522 | 8,287 | 6,818 | 3,136 | 4,136 | ND | 19,038 | 54 |
| 1998 | Soldotna | 4 | 118,331 | 99,399 | 72,667 | 23,930 | 2,547 | 6,801 | 3,129 | ND | 42,557 | 112 |
| 1999 | Soldotna | 4 | ND | 22,202 | 28,163 | 13,661 | 12,961 | 12,854 | 9,637 | ND | 17,968 | 46 |
| 2000 | Soldotna | 4 | ND | 102,706 | 67,418 | 105,175 | 62,123 | 22,778 | 12,738 | ND | 72,040 | 223 |
| 2001 | Soldotna | 4 | ND | 190,588 | 127,123 | 102,203 | 60,516 | 20,052 | 7,149 | ND | 100,096 | 285 |
| 2002 | Soldotna | 4 | ND | 148,739 | 76,142 | 84,416 | 44,723 | 34,841 | 11,360 | ND | 77,767 | 224 |
| 2003 | Soldotna | 4 | ND | 72,126 | 58,403 | 41,696 | 34,344 | 27,645 | ND | ND | 46,245 | 155 |
| 2004 | Kodiak | 4 | 322,445 | 204,279 | 114,239 | 103,138 | 77,528 | 60,430 | 41,911 | ND | 107,217 | 253 |
| 2005 | Kodiak | 4 | 569 | 2,433 | 3,212 | 6,392 | 4,035 | 3,362 | 1,675 | ND | 3,625 | 9 |
| 2006 | Kodiak | 4 | 119,545 | 100,484 | 54,169 | 103,498 | 49,032 | 53,999 | ND | ND | 67,155 | 227 |
| 2007 | Kodiak | 4 | ND | 106,593 | 29,610 | 6,018 | 8,639 | 20,080 | 31,563 | ND | 18,110 | 29 |
| 2008 | Kodiak | 2 | ND | 90,784 | 181,865 | 215,996 | 167,304 | 94,753 | ND | ND | 136,239 | 314 |
| 2009 | Kodiak | 2 | ND | 29,822 | 19,910 | 18,552 | 19,528 | 15,666 | ND | ND | 14,943 | 43 |
| 2010 | Kodiak | 2 | ND | 121,519 | 56,207 | 43,301 | 50,582 | 68,731 | 119,503 | ND | 65,176 | 128 |
| 2011 | Kodiak | 2 | ND | 79,789 | 68,963 | 64,187 | 111,411 | 144,698 | ND | ND | 82,545 | 212 |
| 2012 | Kodiak | 2 | ND | 125,212 | 112,583 | 18,785 | 40,160 | 60,792 | 137,035 | ND | 63,135 | 147 |
| 2013 | Kodiak | 2 | ND | 81,954 | 30,298 | 44,044 | 52,429 | 89,129 | 64,922 | ND | 47,144 | 83 |
| 2014 | Kodiak | 2 | ND | 168,620 | 147,203 | 148,561 | 137,800 | 137,291 | 218,926 | ND | 130,659 | 451 |
| 2015 | Kodiak | 2 | 484,972 | 97,045 | 211,836 ${ }^{\text {a }}$ | 156,308 | 75,904 | 30,735 | 90,338 | ND | 97,372 | 321 |
| 2016 | Kodiak | 2 | 570,131 | 303,108 | 143,064 | 76,159 | 96,069 | 151,129 | 205,638 | ND | 148,506 | 570 |
| 2017 | Kodiak | 2 | ND | 251,825 | 124,979 | 173,374 | 87,876 | 273,306 | ND | ND | 182,272 | 433 |
| 2018 | Kodiak | 2 | ND | 190,949 | 255,031 | 170,487 | ND | 156,648 | ND | ND | 193,279 | 398 |
| 2019 | Kodiak | 2 | ND | 388,351 | 186,449 | 286,806 | 263,287 | 326,541 | 181,525 | ND | 290,287 | 555 |

Notes: The vast majority of species present were Cyclops sp. and ovigerous Cyclops sp. Copepod nauplii were not included, because they were not enumerated in laboratory samples until 2002 and 2004.
a Stations were not averaged in June 2015. Only Station 2A was used in June 2015, because the Station 1A sample estimate was about 4 times larger than any other sample since 1987.


[^0]:    1 Reynolds Manney, A. M. Lynn Canal sockeye stock identification. Saltonstall-Kennedy final performance report, July 1, 2015 through June 30, 2017, NOAA Cooperative Agreement No. NA15NMF4270274, September 22, 2017.

[^1]:    2 R Development Core Team. 2019. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL http://www.R-project.org/.

[^2]:    a Weekly escapement goal targets are from Eggers et al. (2009).
    b Weir removed prior to statistical week 37.

[^3]:    ${ }^{\text {a }}$ Weekly escapement goal targets are from Eggers et al. (2009).
    b Weir removed prior to statistical week 37.

[^4]:    a Other includes Chilkat River mainstem spawning stocks.
    b Harvest proportions and numbers for statistical weeks 35-40 were estimated using the proportions from the last statistical week with genetic samples, in this case statistical week 34 .

[^5]:    a Other includes Chilkat River mainstem spawning stocks.
    b Harvest proportions and numbers for statistical weeks $35-40$ were estimated using the proportions from the last statistical week with genetic samples, in this case statistical week 34.

[^6]:    a Average values use 1976-2015 data and are based on standardized dates (1 June through 27 September).

[^7]:    -continued-

[^8]:    ${ }^{\text {a }}$ Weir pickets were removed from 1200 hours on 28 July through 1500 hours on 29 July due to flood event; interpolated (bold) values were calculated for 28-29 July.

[^9]:    a Weir pickets were removed from 0600 hours on 9 June through 1530 hours on 11 June due to flood event; interpolated (bold) values were calculated for 9-11 June.

[^10]:    a Weir pickets were removed from 1100 hours on 9 August through 1400 hours on 10 August due to flood event; interpolated (bold) values calculated for 9-10 August.

[^11]:    -continued-

[^12]:    -continued-

[^13]:    -continued-

